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Provisional Peer-Reviewed Toxicity Values for

2,3-Toluenediamine (CASRN 2687-25-4)

U.S. EPA Office of Research and Development Center for Public Health and Environmental Assessment

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Center for Public Health and Environmental Assessment Office of Research and Development U.S. Environmental Protection Agency Cincinnati, OH 45268

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Questions regarding the content of this PPRTV assessment should be directed to the U.S. EPA Office of Research and Development (ORD) Center for Public Health and Environmental Assessment (CPHEA) website at [https://ecomments.epa.gov/pprtv.](https://ecomments.epa.gov/pprtv)

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COMMONLY USED ABBREVIATIONS AND ACRONYMS

Abbreviations and acronyms not listed on this page are defined upon first use in the PPRTV document.

DRAFT PROVISIONAL PEER-REVIEWED TOXICITY VALUES FOR 2,3-TOLUENEDIAMINE (CASRN 2687-25-4)

BACKGROUND

A Provisional Peer-Reviewed Toxicity Value (PPRTV) is defined as a toxicity value derived for use in the Superfund Program. PPRTVs are derived after a review of the relevant scientific literature using established U.S. Environmental Protection Agency (U.S. EPA) guidance on human health toxicity value derivations.

The purpose of this document is to provide support for the hazard and dose-response assessment pertaining to chronic and subchronic exposures to substances of concern, to present the major conclusions reached in the hazard identification and derivation of the PPRTVs, and to characterize the overall confidence in these conclusions and toxicity values. It is not intended to be a comprehensive treatise on the chemical or toxicological nature of this substance.

Currently available PPRTV assessments can be accessed on the U.S. EPA's PPRTV website at [https://www.epa.gov/pprtv.](https://www.epa.gov/pprtv) PPRTV assessments are eligible to be updated on a 5-year cycle and revised as appropriate to incorporate new data or methodologies that might impact the toxicity values or affect the characterization of the chemical's potential for causing adverse human-health effects. Questions regarding nomination of chemicals for update can be sent to the appropriate U.S. EPA Superfund and Technology Liaison [\(https://www.epa.gov/research/fact](https://www.epa.gov/research/fact-sheets-regional-science)[sheets-regional-science\)](https://www.epa.gov/research/fact-sheets-regional-science).

QUALITY ASSURANCE

This work was conducted under the U.S. EPA Quality Assurance (QA) program to ensure data are of known and acceptable quality to support their intended use. Surveillance of the work by the assessment managers and programmatic scientific leads ensured adherence to QA processes and criteria, as well as quick and effective resolution of any problems. The QA manager, assessment managers, and programmatic scientific leads have determined under the QA program that this work meets all U.S. EPA quality requirements. This PPRTV was written with guidance from the CPHEA Program Quality Assurance Project Plan (PQAPP), the QAPP titled *Program Quality Assurance Project Plan (PQAPP) for the Provisional Peer-Reviewed Toxicity Values (PPRTVs) and Related Assessments/Documents (L-CPAD-0032718-QP),* and the PPRTV development contractor QAPP titled *Quality Assurance Project Plan—Preparation of Provisional Toxicity Value (PTV) Documents (L-CPAD-0031971-QP)*. As part of the QA system, a quality product review is done prior to management clearance. A Technical Systems Audit may be performed at the discretion of the QA staff.

All PPRTV assessments receive internal peer review by at least two CPHEA scientists and an independent external peer review by at least three scientific experts. The reviews focus on whether all studies have been correctly selected, interpreted, and adequately described for the purposes of deriving a provisional reference value. The reviews also cover quantitative and qualitative aspects of the provisional value development and address whether uncertainties associated with the assessment have been adequately characterized.

DISCLAIMERS

The PPRTV document provides toxicity values and information about the adverse effects of the chemical and the evidence on which the value is based, including the strengths and limitations of the data. All users are advised to review the information provided in this document to ensure that the PPRTV used is appropriate for the types of exposures and circumstances at the site in question and the risk management decision that would be supported by the risk assessment.

Other U.S. EPA programs or external parties who may choose to use PPRTVs are advised that Superfund resources will not generally be used to respond to challenges, if any, of PPRTVs used in a context outside of the Superfund program.

This document has been reviewed in accordance with U.S. EPA policy and approved for publication. Mention of trade names or commercial products does not constitute endorsement or recommendation for use.

QUESTIONS REGARDING PPRTVS

Questions regarding the content of this PPRTV assessment should be directed to the U.S. EPA ORD CPHEA website at [https://ecomments.epa.gov/pprtv.](https://ecomments.epa.gov/pprtv)

1. INTRODUCTION

2,3-Toluenediamine (2,3-TDA), also known as 2,3-diaminotoluene (CASRN 2687-25-4), belongs to the class of compounds known as anilines and is an *ortho* (*o*)-substituted compound. The principal commercial use for *o*-TDA isomers, including 2,3-TDA, is in the production of tolyltriazoles used in corrosion and nitrification inhibitors. 2,3-TDA is also used as an intermediate in the manufacture of urethane products, dyes, corrosion inhibitors, polyols, and benzimidazole thiol antioxidants, and as a starting material for a pharmaceutical intermediate [\(HSDB, 2013a;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102062) [Cartolano, 2005\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102063). It is listed on U.S. EPA's Toxic Substances Control Act's public inventory [\(U.S. EPA, 2015\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3036228), but it is not registered with Europe's Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH) program [\(ECHA, 2016\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3108686).

TDA isomers, including 2,3-TDA, are produced by the catalytic hydrogenation of dinitrotoluenes under a variety of temperatures, pressures, and solvents. 2,3-TDA is then separated from *meta* (*m*)-substituted TDA isomers by vacuum distillation [\(Cartolano, 2005\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102063).

2,3-TDA is one of six TDA isomers that are components of crude or commercial-grade mixtures used as intermediates in the production of dyes and pigments for commercial products [\(WHO, 1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356). The crude mixture contains all six isomeric forms, while the two commercial mixtures are composed primarily of two isomers each. One commercial mixture, *m*-TDA, contains the 2,4- and 2,6- isomers (80:20 or 65:35), and the other, *o*-TDA, contains the 2,3- and 3,4- isomers (40:60).

The empirical formula for 2,3-TDA is C₇H₁₀N₂, and its structure is shown in Figure 1. Table 1 summarizes the physicochemical properties of 2,3-TDA. The compound is a light gray to purple solid at room temperature [\(Cartolano, 2005\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102063). The low vapor pressure and low estimated Henry's law constant for 2,3-TDA indicate that it is unlikely to volatilize from either dry or moist surfaces. 2,3-TDA has an estimated atmospheric half-life of 0.6 hours for the reaction with hydroxyl radicals, but this is not expected to be an important fate process because the compound is not likely to partition to the atmosphere. The estimated water solubility and low soil adsorption coefficient for 2,3-TDA indicate that it has the potential to leach to groundwater or undergo runoff after a rain event. However, given its estimated acid dissociation constant (pKa), 2,3-TDA may exist partially as a cation in the environment, and cations generally adsorb more strongly to soils containing organic carbon and clay than their neutral counterparts. Also, aromatic amines contain highly reactive amino groups that may cause strong bonding to soil organic matter.

Figure 1. 2,3-Toluenediamine (CASRN 2687-25-4) Structure

Table 1. Physicochemical Properties of 2,3-Toluenediamine

a Data were extracted from the U.S. EPA CompTox Chemicals Dashboard (2,3-toluenediamine, CASRN 2687-25-4. [https://comptox.epa.gov/dashboard/DTXSID4027494.](https://comptox.epa.gov/dashboard/DTXSID9024930) Accessed on April 20, 2021). All values are experimental averages unless otherwise specified.

b[U.S. EPA \(2012\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102231) c [HSDB \(2013a\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3102062)

NV = not available.

No toxicity values for 2,3-TDA are available from U.S. EPA or other agencies/organizations searched, as shown in Table 2.

^aSources: ACGIH = American Conference of Governmental Industrial Hygienists; ATSDR = Agency for Toxic Substances and Disease Registry; CalEPA = California Environmental Protection Agency; DWSHA = Drinking Water Standards and Health Advisories; HEAST = Health Effects Assessment Summary Tables; IARC = International Agency for Research on Cancer; IPCS = International Programme on Chemical Safety; IRIS = Integrated Risk Information System; NIOSH = National Institute for Occupational Safety and Health; NTP = National Toxicology Program; OSHA = Occupational Safety and Health Administration. ^bReference date is the publication date for the database and not the date the source was accessed.

 $NA = not applicable; NV = not available.$

Non-date-limited literature searches were conducted in November 2017 and updated in May 2020 and April 2021 for studies relevant to the derivation of provisional toxicity values for 2,3-toluenediamine (CASRN 2687-25-4). Searches were conducted using U.S. EPA's Health and Environmental Research Online (HERO) database of scientific literature. HERO searches the following databases: PubMed, $TOXLINE¹$ $TOXLINE¹$ $TOXLINE¹$ (including $TSCATS1$), and Web of Science. The following resources were searched outside of HERO for health-related values: American Conference of Governmental Industrial Hygienists (ACGIH), Agency for Toxic Substances and Disease Registry (ATSDR), California Environmental Protection Agency (CalEPA), Defense Technical Information Center (DTIC), European Centre for Ecotoxicology and Toxicology of Chemicals (ECETOC), European Chemicals Agency (ECHA), U.S. EPA Chemical Data Access Tool (CDAT), U.S. EPA ChemView, U.S. EPA Integrated Risk Information System (IRIS), U.S. EPA Health Effects Assessment Summary Tables (HEAST), U.S. EPA Office of Water (OW), International Agency for Research on Cancer (IARC), U.S. EPA TSCATS2/TSCATS8e, U.S. EPA High Production Volume (HPV), Chemicals via IPCS INCHEM, European Centre for Ecotoxicology and Toxicology of Chemicals (ECETOC), Japan Existing Chemical Data Base (JECDB), European Chemicals Agency (ECHA), Organisation for Economic Cooperation and Development (OECD) Screening Information Data Sets (SIDS), OECD International Uniform Chemical Information Database (IUCLID), OECD HPV, National Institute for Occupational Safety and Health (NIOSH), National Toxicology Program (NTP), Occupational Safety and Health Administration (OSHA), and World Health Organization (WHO).

¹Note that this version of TOXLINE [\(https://www.nlm.nih.gov/databases/download/toxlinesubset.html\)](https://www.nlm.nih.gov/databases/download/toxlinesubset.html) is no longer updated; therefore, it was not included in the literature search updates from May 2020 and April 2021.

2. REVIEW OF POTENTIALLY RELEVANT DATA (NONCANCER AND CANCER)

As shown in Tables 3A and 3B, there are no potentially relevant subchronic or chronic studies or developmental or reproductive toxicity studies of 2,3-TDA in humans or animals for noncancer and cancer endpoints following oral or inhalation exposures. [WHO \(1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356) described a small number of occupational health surveys of male workers exposed to diaminotoluene and dinitrotoluene mixtures; however, these studies are not useful for determining the effects of 2,3-TDA because they did not discuss or otherwise verify the presence of this isomer in the mixtures. The phrase "statistical significance" and term "significant," used throughout the document, indicate a *p*-value of < 0.05 unless otherwise specified.

 $ND = no data.$

 $ND = no data.$

2.1. HUMAN STUDIES

2.1.1. Oral Exposures

No human studies following oral exposure to 2,3-TDA have been identified.

2.1.2. Inhalation Exposures

No human studies following inhalation exposure to 2,3-TDA have been identified.

2.2. ANIMAL STUDIES

2.2.1. Oral Exposures

No animal studies following oral exposure to 2,3-TDA have been identified.

2.2.2. Inhalation Exposures

No animal studies following inhalation exposure to 2,3-TDA have been identified.

2.3. OTHER DATA (SHORT-TERM TESTS, OTHER EXAMINATIONS)

Toxicity data available for the 2,3-TDA isomer are limited to two genotoxicity studies. Toxicity studies evaluating TDA mixtures containing the 2,3-TDA isomer include a reproductive/toxicity study, two developmental toxicity studies, several acute lethality studies, eye and skin irritation assays, and a genotoxicity study.

2.3.1. Genotoxicity

Limited data are available on the genotoxicity of 2,3-TDA. [Cheung et al. \(1996\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1413137) reported that 2,3-TDA (0, 5, 10, 15, and 20 µg/plate) was weakly mutagenic in *Salmonella typhimurium* strain TA98 with metabolic activation. The use of liver microsomes from rats treated with either Aroclor 1254 (a potent inducer of CYP1A enzymes) or the parent compound (i.e., 2,3-TDA) did not alter the mutagenic response. Administration of 0.2 or 0.5 mg/kg of 2,3-TDA by intraperitoneal (i.p.) injection did not cause an increase in chromosomal aberrations (CAs) or mitotic frequency in bone marrow or ascites tumor cells in male CFW mice [\(Adam, 1985\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1808403). Similarly, the CA frequency was not increased in ascites tumor cells implanted in mice that were exposed subcutaneously to *o*-TDA (a mixture of 2,3- and 3,4-toluenediamine [3,4-TDA]) 4 days after implantation [\[Bogajewski and Bogajewska \(1982\),](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2277077) available only as an abstract].

2.3.2. Acute Toxicity Studies of TDA Mixtures

Available acute toxicity studies of TDA mixtures including 2,3-TDA are summarized in Table 4. Oral median lethal dose (LD₅₀) values in rats for mixtures including 2,3-TDA range from 660 to 1,760 mg/kg [\(Dow Chemical, 1993,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229226) [1983;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229189) [WIL Research, 1978;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2276452) [Air Products and](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178) [Chemicals, 1976;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178) [Carpenter et al., 1974\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=655409). An inhalation median lethal concentration (LC50) value >670 ppm was reported in rats exposed to a mixture of 2,3- and 3,4-TDA; the duration of exposure was not reported [\(Air Products and Chemicals, 1976\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178). Exposure to concentrated vapors of a mixture of 2,3- and 3,4-TDA was lethal after >8 hours of exposure [\(Carpenter et al., 1974\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=655409). Dermal LD₅₀ values in rabbits exposed to mixtures of 2,3- and 3,4-TDA ranged from 1,120 to >5,750 mg/kg [\(Air Products and Chemicals, 1976;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178) [Carpenter et al., 1974\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=655409). TDA mixtures including 2,3-TDA are slightly to moderately irritating to rabbit skin, and the undiluted liquid is irritating to the rabbit eye [\(Dow Chemical, 1993,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229226) [1983;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229189) [Air Products and Chemicals, 1976;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178) [Carpenter et al., 1974\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=655409). Skin sensitization was "insignificant" in guinea pigs exposed to a mixture of 2,3- and 3,4-TDA [\(Air Products and Chemicals, 1976\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229178).

CI = confidence interval; F = female(s); GI = gastrointestinal; LC₅₀ = median lethal concentration; LD₅₀ = median lethal dose; M = male(s); TDA = toluenediamine.

2.3.3. Reproductive/Developmental Studies of TDA Mixtures *[Becci et al. \(1983\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=65182)*

Groups of 22 pregnant Sprague Dawley rats were administered *o*-TDA (a 40:60 mixture of 2,3- and 3,4-TDA) at doses of 0, 10, 30, 100, or 300 mg/kg-day via gavage in corn oil from Gestation Days (GDs) 6−15. Observations were conducted daily for general appearance, behavior, and mortality. Body weights of the dams were recorded on GDs 0, 6, 9, 12, 15, and 20. On GD 20, all dams were sacrificed, and uterine contents were removed and examined. One-half of the fetuses were examined for soft-tissue anomalies and the remaining fetuses were examined for skeletal anomalies. No treatment-related effects on appearance or behavior were observed in treated dams, and all dams survived the duration of the study. There was a statistically significant decrease in weight gain (−20%) during gestation for treated dams receiving 300 mg/kg-day compared with controls. No significant differences in the number of live fetuses, implantation sites, or resorption sites were indicated. Fetal effects indicative of developmental delay included significant reductions in fetal body weight (−18%) in the 300-mg/kg-day group and significant increases in skeletal variations per litter (missing sternebrae at 300 mg/kg-day and incomplete ossification of vertebrae at 100 and 300 mg/kg-day), compared with controls. No exposure-related skeletal or soft-tissue malformations were observed. No maternal or developmental effects were seen at ≤ 30 mg/kg-day.

Additionally, groups of 15 pregnant female Dutch belted rabbits were exposed to *o*-TDA at doses of 0, 3, 10, 30, or 100 mg/kg-day via gavage in corn oil from GDs 6−18. Observations were conducted daily for general appearance, behavior, and mortality. Body weights of the does were recorded on GDs 0, 6, 9, 12, 15, 18, and 29. On GD 29, all does were sacrificed, and uterine contents were removed and examined. All the fetuses were examined for both soft-tissue and skeletal anomalies. Appearance and behavior of does were unaffected by treatment. All does survived the duration of the study. Body-weight gain during gestation was significantly decreased (−60%) in treated does receiving 100 mg/kg-day compared with controls. Other observations at this dose included a significant 2.5-fold increase in the incidence of resorptions, a 16% decrease in the mean number of live fetuses per doe (reported as statistically significant in the text, but not in the table showing the data), and a significant 22% decrease in fetal body weight. No exposure-related skeletal or soft-tissue malformations or variations were observed. No maternal or developmental effects were seen at ≤30 mg/kg-day.

[BASF \(2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2298768)

In an OECD 421 reproductive/developmental (R/D) study available only as an industry-submitted summary, groups of male and female Wistar rats (10/sex/group) were administered *o*-TDA (45:50 mixture of 2,3- and 3,4-TDA) at doses of 0, 10, 50, or 250 mg/kg-day via gavage (vehicle not reported) from premating through mating (males, at least 28 days) or premating through Postnatal Day (PND) 4 (up to 60 days for females). The pups were sacrificed and examined on PND 4 (endpoints examined at sacrifice were not reported). The available summary reported only "the most relevant results"; no statistics were provided, and the summary did not include the magnitude/incidence for many of the findings.

Clinical signs of toxicity (reduced activity, eyelid drop, salivation, and/or piloerection) were observed in males and dams at >50 mg/kg-day. Decreased food consumption was observed during premating in males at 250 mg/kg-day and females at \geq 50 mg/kg-day; decreased food consumption was also observed in dams during gestation at 250 mg/kg-day. Body-weight gain was decreased throughout the study in high-dose males, with a decreased terminal body weight compared with controls (magnitude not reported). Decreased body weight and body-weight gains

were observed during premating and gestation in dams at 250 mg/kg-day, with a decreased terminal body weight compared with controls (magnitude not reported). In males, a decrease in the number of spermatids/g testis was reported at 250 mg/kg-day; however, the study summary did not indicate whether decreased fertility was observed. Reproductive effects observed in high-dose dams included a 39% decrease in the number of implantation sites compared with controls, a 27.4% post implantation loss (control value not reported), and a 42% decrease in the number of delivered pups/litter. In offspring, a decreased viability index of 91% was observed at 250 mg/kg-day (viability index in controls was not reported). No adverse effects were noted in males or dams administered 10 mg/kg-day.

2.3.4. Mode-of-Action/Mechanistic Studies

Mechanistic data for 2,3-TDA are limited. Following a single i.p. injection of 0, 10, 20, or 40 mg/kg 2,3-TDA, CYP1A, CYP1A2, and 7-ethoxyresorufin *O*-deethylation levels in rat liver were elevated in a dose-dependent fashion; significant elevations in 7-methoxyresorufin *O*-demethylation were only observed at the highest dose [\(Cheung et al., 1996\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1413137). In vitro, 2,3-TDA was able to bind to the rat cytosolic aromatic hydrocarbon receptor; the median effective concentration for 2,3,7,8-tetrachlorodibenzo-*p*-dioxin displacement was 1.5 × 10[−]⁵ M [\(Cheung et](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1413137) [al., 1996\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1413137).

3. DERIVATION OF PROVISIONAL VALUES

3.1. DERIVATION OF PROVISIONAL REFERENCE DOSES

No subchronic or chronic studies were located regarding toxicity of 2,3-TDA to humans or animals via oral administration. Gavage exposure studies evaluating TDA mixtures containing the 2,3-TDA isomer (approximately 40−45% 2,3-TDA) showed some evidence of R/D effects in rats exposed to 250 mg/kg-day following premating through PND 4 and rats and rabbits exposed to ≥100 mg/kg-day following gestational exposure, primarily at doses associated with potential maternal toxicity [\(BASF, 2010;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2298768) [Becci et al., 1983\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=65182). The scope and design of these studies are inadequate to support the derivation of a subchronic or chronic provisional reference dose (p-RfD) for 3,4-TDA using chemical-specific data. Instead, screening p-RfDs are derived in Appendix A using an alternative read-across approach.

3.2. DERIVATION OF PROVISIONAL REFERENCE CONCENTRATIONS

The absence of relevant inhalation data precludes derivation of provisional reference concentrations (p-RfCs) for 2,3-TDA directly. An alternative read-across approach was attempted, but screening p-RfCs could not be derived due to a lack of inhalation toxicity values for potential analogues (see Appendix A). Based on the overall analogue approach presented in Appendix A, 2,5-TDA was selected as the most appropriate analogue for 2,3-TDA for deriving a screening subchronic and chronic p-RfD (see Table 5).

3.3. SUMMARY OF NONCANCER PROVISIONAL REFERENCE VALUES

The noncancer provisional reference values for 2,3-TDA are summarized in Table 5.

 $AST =$ aspartate aminotransferase; $F =$ female(s); $HED =$ human equivalent dose; $NDr =$ not determined; NOAEL = no-observed-adverse-effect level; $POD = point$ of departure; $p-RfC =$ provisional reference concentration; $p-RfD =$ provisional reference dose; $UF_C =$ composite uncertainty factor.

3.4. CANCER WEIGHT-OF-EVIDENCE DESCRIPTOR

Under the U.S. EPA Cancer Guidelines [\(U.S. EPA, 2005a\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6324329), there is *"Inadequate Information to Assess the Carcinogenic Potential"* of 2,3-TDA (see Table 6). No relevant studies

are available in humans or animals. Within the current U.S. EPA Cancer Guidelines (U.S. EPA, [2005a\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6324329), there is no standard methodology to support the identification of a weight-of-evidence (WOE) descriptor and derivation of provisional cancer risk estimates for data-poor chemicals using an analogue approach. In the absence of an established framework, a screening evaluation of potential carcinogenicity is provided using the methodology described in Appendix B. This evaluation determined that there was a *concern for potential carcinogenicity* of 2,3-TDA (see Appendix C).

 $NA = not applicable; NS = not selected; TDA = tolerance$ toluenediamine; $WOE = weight of evidence$.

3.5. DERIVATION OF PROVISIONAL CANCER RISK ESTIMATES

The absence of suitable data precludes development of cancer risk estimates for 2,3-TDA (see Table 7).

 $NDr = not determined; p-IUR =$ provisional inhalation unit risk; $p-OSF =$ provisional oral slope factor.

APPENDIX A. SCREENING NONCANCER PROVISIONAL VALUES

Due to the lack of evidence described in the main Provisional Peer-Reviewed Toxicity Value (PPRTV) document, it is inappropriate to derive provisional toxicity values for 2,3-toluenediamine (2,3-TDA). However, some information is available for this chemical, which although insufficient to support derivation of a provisional toxicity value under current guidelines, may be of limited use to risk assessors. In such cases, the Center for Public Health and Environmental Assessment (CPHEA) summarizes available information in an appendix and develops a "screening value." Appendices receive the same level of internal and external scientific peer review as the provisional reference values to ensure their appropriateness within the limitations detailed in the document. Users of screening toxicity values in an appendix to a PPRTV assessment should understand that there could be more uncertainty associated with deriving an appendix screening toxicity value than for a value presented in the body of the assessment. Questions or concerns about the appropriate use of screening values should be directed to the CPHEA.

APPLICATION OF AN ALTERNATIVE ANALOGUE APPROACH

The analogue read-across approach allows for the use of data from related compounds to calculate screening values when data for the compound of interest are limited or unavailable. Details regarding searches and methods for analogue analysis are presented in [Wang et al.](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1239453) (2012). Three types of potential analogues (structural, metabolic, and toxicity-like) are identified to facilitate the final analogue chemical selection. The analogue approach may or may not be route specific or applicable to multiple routes of exposure. All information was considered together as part of the final weight-of-evidence (WOE) approach to select the most suitable analogue both toxicologically and chemically.

Structural Analogues

An initial analogue search focused on the identification of structurally similar chemicals with toxicity values from the Integrated Risk Information System (IRIS), PPRTV, Agency for Toxic Substances and Disease Registry (ATSDR), or California Environmental Protection Agency (CalEPA) databases to take advantage of the well-characterized chemical-class information. This was accomplished by searching structural similarity software tools, namely the National Library of Medicine's (NLM) ChemIDplus database [\(NLM, 2019\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6302807) and Organisation for Economic Co-operation and Development (OECD) quantitative structure-activity relationship (QSAR) Toolbox [\(OECD, 2019\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6311654). These software tools employ slightly different quantitative methods to make similarity comparisons between chemical structures based on fingerprints, ChemIDplus uses a modified Tanimoto index and the OECD Toolbox uses the Dice index. Two TDA isomers that have oral noncancer toxicity values were identified as potential structural analogues of 2,3-TDA: 2,6-TDA [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682) and 2,5-TDA [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) (see Table A-1). In addition, 3,4-TDA (a compound being evaluated in a separate PPRTV assessment) and 2,4-TDA were included in the read-across analysis to provide information on the influence of the position of the amino groups (*ortho* [*o-*], *meta* [*m-*], or *para* [*p-*]) on toxicity (see Table A-1). Previous structure-activity relationship (SAR) analyses have suggested increased chemical reactivity and toxicity for *o*- and *p*- versus *m*-substituted aromatic amines [\(Bajot et al., 2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128). The target and 2,3-TDA are *o*- isomers, 2,5-TDA is a *p*- isomer and 2,4- and 2,6-TDAs are *m*- isomers.

Table A-1 summarizes the physicochemical properties and similarity scores for all analogues. 2,3-TDA and the identified analogues are aromatic amines that share a common basic structure, which consists of a benzene ring, two amino groups, and a methyl group, differing only in the position of the amino functional groups. These compounds are major components of commercial-grade TDA mixtures [\(WHO, 1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356) and have physicochemical properties that are of the same relative order of magnitude; therefore, differences in the absorption and distribution between the analogues and the target are not expected to be significant or result in a preference in the selection of one analogue over another. These compounds are all weak bases and are expected to be substantially ionized at physiological pH values. Furthermore, their water solubility and octanol-water partition coefficient ($log K_{ow}$) values are consistent with a high degree of hydrophilicity (see Table A-1). Additionally, all of these diamines have low volatility and are not expected to be eliminated in exhaled breath.

Structural alert (SA) predictions for relevant toxicity endpoints were generated for the TDA isomers using the OECD QSAR Toolbox [\[OECD \(2019\);](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6311654) see Table A-2]. These included the repeated-dose profiler based on the Hazard Evaluation Support System (HESS) database and the developmental and reproductive toxicity (DART) scheme adapted from the [Wu et al. \(2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6311597) framework for identifying chemicals with structural features associated with potential reproductive/developmental (R/D) toxicants. The model predictions suggest concern for hepatotoxicity, hemolytic anemia with methemoglobinemia, and for R/D toxicity for 2,3-TDA and all analogues. The predictions are based on SAs for aniline and toluene/small alkyl toluene derivatives, respectively. The HESS model also showed concern for renal toxicity for 2,3-, 3,4-, and 2,6-TDA based on SA for toluene.

In summary, the candidate analogues are considered suitable analogues for 2,3-TDA based on their similarities in structural and physicochemical properties and SA predictions.

^aData were extracted from the U.S. EPA CompTox Chemicals Dashboard [\(https://comptox.epa.gov/dashboard.](https://comptox.epa.gov/dashboard) Accessed on April 20, 2021). All values are experimental averages unless otherwise specified.

b ChemIDplus advanced similarity scores [\(NLM, 2019\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6302807). c OECD QSAR Toolbox, similarity scores [\(OECD, 2019\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6311654).

 $\frac{\text{dBBB (2013b)}}{\text{Chem} \Delta \text{xon} (201)}$

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 $\sqrt{\frac{ChemAxon(2017)}}$.

OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship.

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^aOECD QSAR Toolbox [\(OECD, 2019\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6311654).

DART = developmental and reproductive toxicity; HESS = Hazard Evaluation Support System; OECD = Organisation for Economic Co-operation and Development; $QSAR =$ quantitative structure-activity relationship; $SA =$ structural alert.

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Metabolic Analogues

Table A-3 summarizes available toxicokinetic data for 2,3-TDA and the structurally similar compounds identified as candidate analogues.

No toxicokinetic data has been identified for 2,3-TDA. Available information on the 2,4-, 2,5-, and 2,6-TDA analogues suggest that these compounds are rapidly and extensively absorbed following oral exposure and are rapidly eliminated in the urine, which is a predominant route of excretion (see Table A-3). Major metabolic steps for the TDA analogues are acetylation of amino groups and ring hydroxylation, with some evidence for oxidation of the methyl groups in rats and mice exposed via intraperitoneal (i.p.) administration (see Table A-3).

In the absence of in vivo toxicokinetic data on 2,3-TDA, a selection of available software tools, specifically the in vivo and in vitro rat metabolic simulators available within the Tissue Metabolism Simulator (TIMES) program [\(Dimitrov et al., 2005;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147) [Mekenyan et al., 2004\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129) and Meteor Nexus [\(Marchant et al., 2008\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=613223) were used to predict metabolites for the target compound and analogues. Predicted metabolites for the TDA isomers are summarized in Tables D-1 and additional information on the in silico analysis can be found in Appendix D. The analysis reveals some overlap in terms of metabolites for the individual TDA compounds across the different tools and when comparing predictions with experimental data from in vivo rodent studies (captured in Table A-3), which increases confidence in the in silico results. Furthermore, the corresponding metabolic pathway transformations were extracted from Meteor Nexus to allow for similarity comparisons across the TDAs. This level of information was not available from other tools (i.e., TIMES). Table A-4 shows a consistent pattern of pathway transformations among the TDA compounds, and Figure A-1 confirms a high degree of similarity between 2,3-TDA and the candidate analogues with regards to the Meteor Nexus pathway predictions. There is also concordance between the in silico results (see Table A-4) and the major pathways expected for this group of compounds (*N*-acetylation, ring hydroxylation, and oxidation of methyl groups). Importantly, no outstanding differences in the predicted metabolic profiles between the target and analogues are noted. The metabolic tree for 2,4-TDA is displayed in Figure D-1 of Appendix D to illustrate the relationship of the predicted metabolites for this specific analogue that correspond to the pathway transformations shared among the TDAs.

In summary, in vivo data demonstrate toxicokinetic commonalities among the analogues, particularly with respect to metabolism pathways, and according to in silico predictions, a similar metabolism pattern is expected for 2,3-TDA. Therefore, the candidate analogues are considered suitable analogues for 2,3-TDA based on their toxicokinetic properties.

ADME = absorption, distribution, metabolism, excretion; C_{max} = maximum concentration; GI = gastrointestinal; i.p. = intraperitoneal; i.v. = intravenous; NA = not applicable; $ND = no$ data.

Table A-4. Comparison of Metabolic Pathway Transformations for 2,3-Toluenediamine (CASRN 2687-25-4) and Candidate Analogues from Meteor Nexusa, b

^aMeteor Nexus (*Dimitrov et al., 2005; Mekenyan et al., 2004*).

^b1/0 denotes whether pathway transformation was identified/not identified.

TDA = toluenediamine.

Figure A-1. Metabolic Pathway Similarities for 2,3-Toluenediamine (CASRN 2687-25-4) and Candidate Analogues. Heatmap displays Jaccard pairwise similarities rounded to two decimal places for the TDA compounds, comparing metabolic pathway transformations from Meteor Nexus (**[Dimitrov et al., 2005](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147)**; **[Mekenyan et al., 2004](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129)**).

Toxicity-Like Analogues

Table A-5 summarizes available subchronic and chronic oral toxicity data for 2,3-TDA and the compounds identified as candidate analogues. None of the analogues had subchronic or chronic inhalation toxicity values from U.S. EPA, ATSDR, or CalEPA.

No repeated-dose oral toxicity data are available for the 2,3-TDA isomer. Oral toxicity values are available for 2,5- and 2,6-TDA. Hepatic effects were the basis for the subchronic and chronic provisional reference doses (p-RfDs) for 2,5-TDA sulfate using a point of departure (POD) of 2.5 mg/kg-day (1.4 mg/kg-day for the free base estimated for this assessment) [\(U.S.](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). Thyroid and body-weight effects were the basis for the subchronic p-RfD for 2,6-TDA using a POD of 62 mg/kg-day [\(U.S. EPA,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682) 2005b). The chronic p-RfD for 2,6-TDA was derived based on a POD of 25 mg/kg-day for no adverse effects [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). Although thyroid toxicity was only noted following 2,6-TDA exposure, liver effects (ranging from changes in serum biomarkers and liver weight to gross and histopathological lesions) were reported after exposure to 2,4- (≥ 6 mg/kg-day), 2,5- (≥ 3 mg/kg-day), and 2,6-TDA (692 mg/kg-day) (U.S. [EPA, 2013,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [2005b;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682) [Criteria Group for Occupational Standards, 2001;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229354) [WHO, 1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356).

R/D toxicity was commonly seen with exposure to TDA compounds, but these endpoints were generally less sensitive than the systemic effects described above. Impaired male fertility and sperm damage were reported in male rats orally exposed to 2,4-TDA at 15 mg/kg-day [\(Criteria Group for Occupational Standards, 2001\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229354). Developmental effects were observed in laboratory animals orally exposed to 2,6- (\geq 100 mg/kg-day) or 2,5-TDA (\geq 44 mg/kg-day) during gestation, primarily at doses associated with potential maternal toxicity [\(U.S. EPA, 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [WHO,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356) [1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356). Data on the *o*-TDA mixture (40:60 or 45:50 mixture of 2,3- and 3,4-TDA) showed possible evidence of R/D effects in rats and rabbits exposed to \geq 100 mg/kg-day via gavage, including alterations in sperm measures (decreased spermatid number) and/or fetal viability and growth (decreased implantation sites, litter size, pup viability, and fetal weight, as well as increased post implantation loss, resorptions, and skeletal variations) often accompanied by reductions in maternal body-weight gain [\(BASF, 2010;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2298768) [Becci et al., 1983\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=65182). These findings suggest that the reproductive system and developing embryo/fetus may be toxicity targets of 2,3- and 3,4-TDA.

Acute lethality studies via different exposure routes were available for *o*- (2,3- and 3,4-) and *m*- (2,4- and 2,6-) TDA mixtures and individual TDA isomers (see Table A-6). The oral median lethal dose (LD₅₀) value for 2,5-TDA (102 mg/kg) in rats was lower than the LD₅₀ values for o - (660 and 810 mg/kg) and m -TDA (270 and 300 mg/kg) mixtures. In mice, the i.p. LD₅₀ values for 2,3-TDA (286 mg/kg) and a *m*-TDA mixture (240 mg/kg) were similar. Likewise, the rabbit dermal LD₅₀ value for a *o*-TDA mixture (1,120 mg/kg) was similar to the rat dermal LD₅₀ value for a *m*-TDA mixture (1,200 mg/kg). Central nervous system depression and methemoglobinemia have been associated with high-dose, acute TDA toxicity in animals [\(WHO,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356) [1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229356).

SAR evaluations have suggested increased chemical reactivity for *o*- and *p*-substituted aromatic amines such as 2,3-, 3,4-, and 2,5-TDA based on their oxidation potential into reactive quinones that can interact with glutathione to produce reactive oxygen species (ROS) [\(Bajot et](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128) [al., 2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128). In contrast, *m*-substituted aromatic amines such as 2,4- and 2,6-TDA are less likely to form quinones and are therefore expected to have decreased chemical reactivity [\(Bajot et al.,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128) [2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128). The *o*- and *p*- substituents were also associated with enhanced, acute aquatic toxicity compared to *m*- substituents [\(Bajot et al., 2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128). No experimental data was found to evaluate

potential differences in chemical reactivity for the TDA isomers. The available evidence in animals shows consistency with respect to toxicity targets (primarily liver and R/D effects) among this group of compounds, and although potency differences are apparent in some cases, there is no clear pattern with respect to the position of the amino groups.

In summary, limited toxicity data for 2,3-TDA from mixture studies reveals similarities in acute toxicity potency and R/D outcomes between the target and analogues. As such, the candidate analogues are considered suitable analogues for 2,3-TDA on the basis of toxicity similarity comparisons.

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^aThe screening p-RfD values for 2,5-TDA as free base were calculated as follows: p-RfD for 2,5-TDA sulfate × (MW of 2,5-TDA as free base [122.17] ÷ MW of 2,5-TDA sulfate $[220.25] = 0.55$) [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821).

AST = aspartate aminotransferase; BW = body weight; CPK = creatine phosphokinase; GD = gestation day; LDH = lactate dehydrogenase; MW = molecular weight; NA = not applicable; NOAEL = no-observed-adverse-effect level; NTP = National Toxicology Program; POD = point of departure; R/D = reproductive/developmental; $p-RfD =$ provisional reference dose; TDA = toluenediamine; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.

i.p. = intraperitoneal; LD_{50} = median lethal dose; NA = not available; TDA = toluenediamine.

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Weight-of-Evidence Approach

A WOE approach is used to evaluate information from candidate analogues as described by [Wang et al. \(2012\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1239453) Commonalities in structural/physicochemical properties, toxicokinetics, metabolism, toxicity, or mode of action (MOA) between candidate analogues and chemical(s) of concern are identified. Emphasis is given to toxicological and/or toxicokinetic similarity over structural similarity. Analogues are excluded if they do not have commonality or demonstrate significantly different physicochemical properties and toxicokinetic profiles that set them apart from the pool of analogues and/or chemical(s) of concern. From the remaining analogues, the most appropriate analogue (most biologically or toxicologically relevant analogue chemical) with the highest structural similarity and/or most conservative toxicity value is selected.

Oral

2,5- and 2,6-TDA were identified as structural analogues of 2,3-TDA with available noncancer oral toxicity values. Two additional structural analogues were included in the read-across analysis, 3,4- and 2,4-TDA, to provide information on the potential influence of the position of the amino groups on toxicity. The analogues share a basic structure with the target compound (a benzene ring, two amino groups, and a methyl group, differing only in the position of the amino functional groups) and have similar physicochemical properties (i.e., water solubility, $log K_{ow}$, volatility, etc.; see Table A-1) important for bioavailability. 2,3-TDA and its analogues also showed similar SA predictions for repeated-dose toxicity and R/D endpoints (see Table A-2). Evidence from oral and i.p. exposure studies in rodents suggests that the TDA analogues are predominantly metabolized via acetylation of amino groups, ring hydroxylation, and potential oxidation of methyl groups (see Table A-3). A comparative analysis of metabolite predictions across different software tools revealed a similar metabolic profile for the target compound and analogues and confirmed observations from in vivo studies (see "Metabolic Analogues" section above and Appendix D for more details). Repeated-dose toxicity studies showed commonalities in target tissues for the TDA analogues, most notably, liver and R/D toxicities (see Table A-5). No adequate toxicity data is available for the 2,3-TDA; however, studies evaluating TDA mixtures containing 2,3-TDA suggest similarities between the target and analogues with respect to acute toxicity potency and R/D outcomes (see Tables A-5 and A-6).

Similarities in structure, physicochemical properties, SA and metabolite predictions and limited toxicity data support the suitability of both 2,5- and 2,6-TDA (the two analogues with available toxicity values) as analogues of 2,3-TDA. 2,5-TDA is selected as the most appropriate analogue for deriving screening p-RfDs based on mechanistic considerations and health protectiveness. Although it is unclear how the position of the amino groups affects the repeated-dose toxicity of TDA compounds, the *o*- and *p*- isomers (2,3- and 2,5-TDA, respectively) are expected to have greater chemical reactivity (related to quinone formation) than the *m*- isomer (2,6-TDA). Furthermore, the POD values for 2,5-TDA (1.4 mg/kg-day for both the subchronic and chronic p-RfDs) are more than an order of magnitude lower than the POD values for 2,6-TDA (62 and 25 mg/kg-day for the subchronic and chronic p-RfDs, respectively).

Inhalation

None of the candidate analogues have repeated-dose inhalation toxicity values, precluding derivation of screening provisional reference concentrations (p-RfCs).

NONCANCER ORAL TOXICITY VALUES Derivation of a Screening Subchronic Provisional Reference Dose

Based on the overall analogue approach presented in this PPRTV assessment, 2,5-TDA is selected as the analogue for 2,3-TDA for deriving a screening subchronic p-RfD. The study used for the U.S. EPA screening subchronic p-RfD for 2,5-TDA was a 13-week rat study [Hill (1997) as cited in [SCCP \(2007\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2345411) and reported in [U.S. EPA \(2013\)\]](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). [U.S. EPA \(2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) described the study as follows:

Hill (1997, as cited in SCCP, 2007) administered toluene-2,5-diamine sulfate (99.7% pure) via gavage in deionized water to Sprague-Dawley rats (15/sex/dose) at 0, 2.5, 5, 10, or 20 mg/kg-day for 13 weeks. The original report for this study is not available; SCCP briefly described the study. Animals were observed daily for mortality and clinical signs. Body weights and food intake were recorded weekly. Ophthalmoscopic examinations were performed on all animals before the initiation of treatment and during Week 13. Blood and urine samples were collected during Week 4 and during Week 12 or 13. Following treatment, all animals were sacrificed and necropsied. Organ weights were recorded, and tissues were subjected to microscopic examination. No dose-related changes in mortality, clinical signs, body weights, body-weight gains, or food consumption were reported (data not shown). The researchers did not consider hematological variations (not further described) to be treatment-related. Aspartate aminotransferase (AST) levels were statistically significantly (p < 0.05) increased in females at doses of ≥5 mg/kg-day (data not shown). Increased urine levels, associated with a statistically $(p < 0.05)$ *significant decrease in specific gravity, were observed at ≥10 mg/kg-day (females) or 20 mg/kg-day (males) (data not shown). Although retinopathy was observed in some animals, a pathology peer review concluded that the incidence of these effects in the treatment groups was similar to the spontaneous incidence for Sprague-Dawley rats. At 20 mg/kg-day, an increased incidence of abnormally shaped pituitary glands was reported. The SCCP (2007) identified a NOAEL of 2.5 mg/kg-day for toluene-2,5-diamine sulfate in this study based on significantly elevated AST levels at 5 mg/kg-day. However, experimental data were not presented in the summary, and the adversity of the reported effects has not been demonstrated (there was no mention of the magnitude or dose-response of the observed change in AST, or corresponding changes in other serum enzymes or liver pathology). The available description of this study lacked information to support independent evaluation of the study.*

The apparent NOAEL of 2.5 mg/kg-day and LOAEL of 5 mg/kg-day for increased serum AST levels in rats treated with toluene-2,5-diamine sulfate by gavage in water for 13 weeks (Hill, 1997, as cited in SCCP, 2007) can be used as the basis for derivation of screening provisional toxicity values for toluene-2,5-diamine sulfate and toluene-2,5-diamine. Based on available information, this appeared to be the most sensitive endpoint identified in the available studies. The choice of endpoint was supported by the results of the 14-day range-finding study, which reported changes in AST and other clinical chemistry measures at 30 mg/kg-day (Hill, 1994, as cited in SCCP, 2007).

Reproductive and developmental toxicity studies reported effects only at higher doses (80−160 mg/kg-day) (Kavlock et al., 1987; Seidenburg et al., 1986; Osterberg, 1982a,b, as cited in SCCP, 2007 and reviewed in Pang, 1992).

The critical effect identified in the [U.S. EPA \(2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) assessment for 2,5-TDA was increased serum aspartate aminotransferase (AST) in female rats exposed for 13 weeks [\(Hill,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2345404) [1997\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2345404). A no-observed-adverse-effect level (NOAEL) of 2.5 mg/kg-day for increased AST was selected as the POD in the screening subchronic p-RfD for 2,5-TDA sulfate [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). The corresponding POD for the free base is calculated by multiplying the POD for the sulfate by the ratio of the molecular weights (MW of 2,5-TDA as free base $[122.17] \div MW$ of 2,5-TDA sulfate $[220.25] = 0.55$). The resulting NOAEL of 1.4 mg/kg-day for 2,5-TDA is adopted as the POD for deriving the screening subchronic p-RfD for 2,3-TDA. The NOAEL of 1.4 mg/kg-day is not adjusted for molecular-weight differences between 2,3- and 2,5-TDA (both as free base), because the molecular weights are identical.

The NOAEL of 1.4 mg/kg-day is converted to a human equivalent dose (HED) according to current [\(U.S. EPA, 2011c\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=752972) guidance. In *Recommended Use of Body Weight3/4 as the Default Method in Derivation of the Oral Reference Dose* [\(U.S. EPA, 2011c\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=752972), the Agency endorses body-weight scaling to the $3/4$ power (i.e., BW^{3/4}) as a default to extrapolate toxicologically equivalent doses of orally administered agents from all laboratory animals to humans for the purpose of deriving an RfD from effects that are not portal-of-entry effects.

Following [U.S. EPA \(2011c\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=752972) guidance, the POD for increased serum AST in female rats is converted to an HED through the application of a dosimetric adjustment factor (DAF) derived as follows:

where

 $\text{DAF} = (\text{BW}_{a}^{1/4} \div \text{BW}_{h}^{1/4})$

DAF = dosimetric adjustment factor $BW_a =$ animal body weight $BW_h =$ human body weight

Using a reference BW_a of 0.204 kg for female Sprague Dawley rats in a subchronic study and a reference BWh of 70 kg for humans [\(U.S. EPA, 1988\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=64560), the resulting DAF is 0.23. Applying this DAF to the NOAEL of 1.4 mg/kg-day yields a POD (HED) as follows:

In deriving a screening p-RfD for 2,3-TDA, a composite uncertainty factor (UFC) of 300 is applied, based on a 3-fold uncertainty factor value for interspecies extrapolation (interspecies uncertainty factor [UFA], reflecting use of a dosimetric adjustment) and 10-fold uncertainty factor values for both intraspecies variability (UFH) and database deficiencies (database uncertainty factor [UFD], reflecting lack of adequate repeated-dose toxicity information for 2,3-TDA). The screening subchronic p-RfD for 2,3-TDA is derived as follows:

Table A-7 summarizes the uncertainty factors for the screening subchronic p-RfD for 2,3-TDA.

DAF = dosimetric adjustment factor; HED = human equivalent dose; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfD = provisional reference dose; TDA = toluenediamine; UF = uncertainty factor; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; $UF_D =$ database uncertainty factor; $UF_H =$ intraspecies uncertainty factor; $UF_L = LOAEL-to-NOAEL uncertainty factor$; $UF_S = subchronic-to-chronic uncertainty factor$.

Derivation of a Screening Chronic Provisional Reference Dose

2,5-TDA is also selected as the most appropriate analogue for 2,3-TDA for deriving the screening chronic p-RfD. [U.S. EPA \(2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) used the critical effect of increased AST levels in female rats and associated POD of 2.5 mg/kg-day (HED of 0.32 mg/kg-day estimated for this assessment) identified in the 13-week rat study [\(Hill, 1997\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2345404) to derive a screening chronic p-RfD for 2,5-TDA. The principal study and calculation of the POD (HED) is described above. Although a cancer bioassay in rats and mice exposed to 2,5-TDA for 78 weeks was available [\(NTP, 1978\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4433157), the [U.S. EPA \(2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) assessment concluded that the study was inadequate in scope and design for evaluating noncancer oral toxicity based on the following: (1) evaluations were limited to measures of body weight, food consumption, clinical signs, and non-neoplastic histopathology; (2) histopathological examinations were conducted after a lengthy recovery period (28−31 weeks); and (3) treatment was initiated at different times for the low- and high-dose groups $(\sim 11$ months apart).

In deriving the screening chronic p-RfD for 2,3-TDA, the POD (HED) of 0.32 mg/kg-day from the 13-week rat study with 2,5-TDA is selected, applying an additional uncertainty factor of 10 to account for increased uncertainty associated with extrapolating from a subchronic to a chronic exposure (UFs). A UF_C of 3,000 was derived, reflecting a 3-fold UF_A, and 10-fold uncertainty factor values for UFH, UFS, and UFD. Finally, the screening chronic p-RfD for 2,3-TDA is derived as follows:

Table A-8 summarizes the uncertainty factors for the screening chronic p-RfD for 2,3-TDA.

DAF = dosimetric adjustment factor; HED = human equivalent dose; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfD = provisional reference dose; $TDA =$ toluenediamine; $UF =$ uncertainty factor; $UF_A =$ interspecies uncertainty factor; $UF_C =$ composite uncertainty factor; $UF_D =$ database uncertainty factor; $UF_H =$ intraspecies uncertainty factor; $UF_L = LOAEL-to-NOAEL$ uncertainty factor; $UF_S = subchronic-to-chronic$ uncertainty factor.

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APPENDIX B. BACKGROUND AND METHODOLOGY FOR THE SCREENING EVALUATION OF POTENTIAL CARCINOGENICITY

For reasons noted in the main Provisional Peer-Reviewed Toxicity Value (PPRTV) document, there is inadequate information to assess the carcinogenic potential of 2,3-toluenediamine (2,3-TDA). However, information is available for this chemical which, although insufficient to support a weight-of-evidence (WOE) descriptor and derivation of provisional cancer risk estimates under current guidelines, may be of use to risk assessors. In such cases, the Center for Public Health and Environmental Assessment (CPHEA) summarizes available information in an appendix and develops a "screening evaluation of potential carcinogenicity." Appendices receive the same level of internal and external scientific peer review as the provisional cancer assessments in PPRTVs to ensure their appropriateness within the limitations detailed in the document. Users of the information regarding potential carcinogenicity in this appendix should understand that there could be more uncertainty associated with this evaluation than for the cancer WOE descriptors presented in the body of the assessment. Questions or concerns about the appropriate use of the screening evaluation of potential carcinogenicity should be directed to the CPHEA.

The screening evaluation of potential carcinogenicity includes the general steps shown in Figure B-1. The methods for Steps 1−8 apply to any target chemical and are described in this appendix. Chemical-specific data for all steps in this process are summarized in Appendix C.

Figure B-1. Steps Used in the Screening Evaluation of Potential Carcinogenicity

STEP 1. USE OF AUTOMATED TOOLS TO IDENTIFY STRUCTURAL ANALOGUES WITH CARCINOGENICITY AND/OR GENOTOXICITY DATA ChemACE Clustering

The U.S. EPA's Chemical Assessment Clustering Engine [ChemACE; [U.S. EPA](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442545) (2011a)] is an automated tool that groups (or clusters) a user-defined list of chemicals based on chemical structure fragments. The methodology used to develop ChemACE was derived from U.S. EPA's Analog Identification Methodology (AIM) tool, which identifies structural analogues for a chemical based on common structural fragments. ChemACE uses the AIM structural fragment recognition approach for analogue identification and applies advanced queries and user-defined rules to create the chemical clusters. The ChemACE cluster outputs are available in several formats and layouts (i.e., Microsoft Excel, Adobe PDF) to allow rapid evaluation of structures, properties, mechanisms, and other parameters which are customizable based on an individual user's needs. ChemACE clustering has been successfully used with chemical inventories for identifying trends within a series of structurally similar chemicals, demonstrating structural diversity in a chemical inventory, and detecting structural analogues to fill data gaps and/or perform read-across analysis.

For this project, ChemACE is used to identify potential structural analogues of the target compound that have available carcinogenicity assessments and/or carcinogenicity data. An overview of the ChemACE process in shown in Figure B-2.

Figure B-2. Overview of ChemACE Clustering Process

The chemical inventory was populated with chemicals from the following databases and lists:

- Carcinogenic Potency Database [CPDB; [CPDB \(2011\)\]](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442529)
- Agents classified by the International Agency for Research on Cancer (IARC) monographs [\(IARC, 2018\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4235828)
- National Toxicology Program (NTP) Report on Carcinogens [ROC; [NTP \(2016a\)\]](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3827262)
- NTP technical reports [\(NTP, 2017\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442566)
- Integrated Risk Information System (IRIS) carcinogens [\(U.S. EPA, 2017\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442576)
- California EPA (CalEPA) Prop 65 list [\(CalEPA, 2017\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442577)
- European Chemicals Agency (ECHA) carcinogenicity data available in the Organisation for Economic Co-operation and Development (OECD) Quantitative Structure-Activity Relationship (QSAR) Toolbox [\(OECD, 2017\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3970963)
- PPRTVs for Superfund [\(U.S. EPA, 2020b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4443402)

In total, 2,123 distinct substances were identified from the sources above. For the purpose of ChemACE clustering, each individual substance needed to meet the following criteria:

- 1) Substance is not a polymer, metal, inorganic, or complex salt because ChemACE is not designed to accommodate these substances;
- 2) Substance has CASRN or unambiguous chemical identification; and
- 3) A unique Simplified Molecular Input Line Entry System (SMILES) notation (encoded molecular structure format used in ChemACE) for the substance can be identified from one of these sources:
	- a. Syracuse Research Corporation (SRC) and Distributed Structure-Searchable Toxicity (DSSTox) lists of known SMILES associated with unique CASRNs (the combined lists contained >200,000 SMILES); or
	- b. ChemIDplus, U.S. EPA CompTox Chemicals Dashboard, or internet searches.

Of the initial list of 2,123 substances, 201 were removed because they did not meet one of the first two criteria, and 155 were removed because they did not meet the third. The final inventory of substances contained 1,767 unique compounds.

Two separate ChemACE approaches were compared for clustering of the chemical inventory. The restrictive clustering approach, in which all compounds in a cluster contain all of the same fragments and no different fragments, resulted in 208 clusters. The less restrictive approach included the following rules for remapping the chemical inventory:

- treat adjacent halogens as equivalent, allowing fluorine (F) to be substituted for chlorine (Cl), Cl for bromine (Br), Br for iodine (I);
- allow methyl, methylene, and methane to be equivalent;
- allow primary, secondary, and tertiary amines to be equivalent; and
- exclude aromatic thiols (removes thiols from consideration).

Clustering using the less restrictive approach (Pass 2) resulted in 284 clusters. ChemACE results for clustering of the target chemical within the clusters of the chemical inventory are described in Appendix C.

Analogue Searches in the OECD QSAR Toolbox (Dice Method)

The OECD QSAR Toolbox (Version 4.1) is used to search for additional structural analogues of the target compound. There are several structural similarity score equations available in the Toolbox (Dice, Tanimoto, Kulczynski-2, Ochiai/Cosine, and Yule). Dice is considered the default equation. The specific options that are selected for the performance of this search include a comparison of molecular features (atom-centered fragments) and atom characteristics (atom type, count hydrogens attached, and hybridization). Chemicals identified in these similarity searches are selected if their similarity scores exceeded 50%.

The OECD QSAR Toolbox Profiler is used to identify those structural analogues from the Dice search that have carcinogenicity and/or genotoxicity data. Nine databases in the OECD QSAR Toolbox (Version 4.1) provide data for carcinogenicity or genotoxicity (see Table B-1).

Analogue search results for the target chemical are described in Appendix C.

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^aDescriptions were obtained from the OECD QSAR Toolbox documentation [Version 4.1; [OECD \(2017\)\]](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3970963).

CA = chromosomal aberration; CCRIS = Chemical Carcinogenesis Research Information System; CPBD = Carcinogenic Potency Database; CRADA = cooperative research and development agreement; DNA = deoxyribonucleic acid; ECHA = European Chemicals Agency; ECVAM = European Centre for the Validation of Alternative Methods; FDA = Food and Drug Administration; IARC = International Agency for Research on Cancer; ISSCAN = Istituto Superiore di Sanità Chemical Carcinogen; ISSCTA = Istituto Superiore di Sanità Cell Transformation Assay; ISSMIC = Istituto Superiore di Sanità Micronucleus; ISSSTY = Istituto Superiore di Sanità *Salmonella typhimurium*; MLA = mouse lymphoma gene mutation assay; MN = micronuclei; MNT = micronucleus test; NCI = National Cancer Institute; NTP = National Toxicology Program; OECD = Organization for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; REACH = Registration, Evaluation, Authorization and Restriction of Chemicals; TD_{50} = median toxic dose.

STEPS 2−5. ANALOGUE REFINEMENT AND SUMMARY OF EXPERIMENTAL DATA FOR GENOTOXICITY, TOXICOKINETICS, CARCINOGENICITY, AND MODE OF ACTION

The outcome of the Step 1 analogue identification process using ChemACE and the OECD QSAR Toolbox is an initial list of structural analogues with genotoxicity and/or carcinogenicity data. Expert judgment is applied in Step 2 to refine the list of analogues based on physicochemical properties; absorption, distribution, metabolism, and excretion (ADME); and mechanisms of toxicity. The analogue refinement process is chemical-specific and is described in Appendix C. Steps 3, 4, and 5 (summary of experimental data for genotoxicity, toxicokinetics, carcinogenicity, and mode of action [MOA]) are also chemical specific (see Appendix C for further details).

STEP 6. STRUCTURAL ALERTS AND STRUCTURE-ACTIVITY RELATIONSHIP PREDICTIONS FOR 2,3-TDA AND ANALOGUES

Structural alerts (SAs) and predictions for genotoxicity and carcinogenicity are identified using six freely available structure-based tools (described in Table B-2).

^aThere is some overlap between the tools. For example, OncoLogic classification is provided by the QSAR Toolbox, but the prediction is available only through OncoLogic, and alerts or decision trees were used in or adapted from several models (e.g., Benigni and Bossa alerts and Toxtree decision tree) [\(OECD, 2017\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=3970963).

ANTARES = Alternative Non-Testing Methods Assessed for REACH Substances; CA = chromosomal aberration; CAESAR = Computer Assisted Evaluation of industrial chemical Substances According to Regulations; CONSENSUS = consensus assessment based on multiple models (CAESAR, SARpy, ISS, and *k*-NN); CRS4 = Center for Advanced Studies, Research and Development in Sardinia; CPDB = Carcinogenic Potency Database; DNA = deoxyribonucleic acid; FN = false negative; IRFMN = Istituto di Ricerche Farmacologiche Mario Negri; ISS = Istituto Superiore di Sanità; ISSCAN-CGX = Istituto Superiore di Sanità Chemical Carcinogen; *k*-NN = *k*-nearest neighbor; LMC = Laboratory for Mathematical Chemistry; MN = micronucleus; MNT = micronucleus test; OCHEM = Online Chemical Monitoring Environment; OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; REACH = Registration, Evaluation, Authorisation and Restriction of Chemicals; SA = structural alert; SAR = structure-activity relationship; SVM = support vector machine; TIMES = The Integrated MARKEL-EFOM System; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

The tool results for the target and analogue compounds are provided in Appendix C.

STEP 7. EVIDENCE INTEGRATION FOR SCREENING EVALUATION OF 2,3-TDA CARCINOGENICITY

Data identified across multiple lines of evidence from Steps 1−6 (outlined above) are integrated to determine the qualitative level of *concern for potential carcinogenicity* of the target compound (Step 8). In the absence of information supporting carcinogenic portal-of-entry effects, the qualitative level of concern for the target chemical should be considered applicable to all routes of exposure.

Evidence integration for the target compound is provided in Appendix C.

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APPENDIX C. RESULTS OF THE SCREENING EVALUATION OF POTENTIAL CARCINOGENICITY

STEP 1. USE OF AUTOMATED TOOLS TO IDENTIFY STRUCTURAL ANALOGUES WITH CARCINOGENICITY AND/OR GENOTOXICITY DATA

U.S. EPA's Chemical Assessment Clustering Engine (ChemACE) clustering was performed as described in Appendix B. The cluster containing 2,3-toluenediamine (2,3-TDA; less restrictive approach; Cluster 71) also contains 3,4-toluenediamine (3,4-TDA; an additional target compound being evaluated in a separate Provisional Peer-Reviewed Toxicity Value [PPRTV] document) and 13 structural analogues. The 15 cluster members all contain a benzene ring substituted with one or more amino groups (−NR2) and one or more methyl groups (−CH3). The methyl groups are present on the ring or the nitrogen substituent $(-N(CH_3)_2)$ (see Figure C-1).

Figure C-1. Illustration of Common Fragments in Cluster 71

The Organisation for Economic Co-operation and Development (OECD) Quantitative Structure-Activity Relationship (QSAR) Toolbox Profiler was used to identify structural analogues from the Dice analogue search with carcinogenicity and/or genotoxicity data (see Step 1 methods in Appendix B). This process identified an additional 31 compounds to be considered as potential analogues for 2,3-TDA. Refinement of selection of final analogues is described below.

STEP 2. ANALOGUE REFINEMENT USING EXPERT JUDGMENT

Expert judgment was applied to refine the initial list of 44 potential analogues based on physicochemical properties; absorption, distribution, metabolism, and excretion (ADME); and mechanisms of toxicity.

Compounds were considered potential analogues if they had (1) one aromatic ring (benzene) substituted with (2) two unsubstituted amines on the ring, in a *meta* (*m*)- or *para* (*p*)-substitution pattern, (3) a methyl group on the ring, and (4) no other functional group. Such compounds are similar to the target chemical in all attributes except for the proximity of the two amine substituents to one another on the aromatic ring. The closest analogue for 2,3-TDA structurally would be 3,4-TDA because of the similar *ortho* (*o*)-substitution pattern; however, 3,4-TDA does not have adequate experimental data for evaluating potential carcinogenicity and was not considered further as an analogue. Simple salts (e.g., hydrochlorides or sulfates) of the *m*- and *p*-substituted diamines are also considered as potential analogues.

Of the 44 chemicals identified as potential analogues by ChemACE clustering and the OECD Toolbox analogue selection tool (Dice), 36 were not selected for further review. Common rationales for not selecting these chemicals included the presence of polycyclic aromatics or ring systems other than toluene; lack of two amine substituents; occurrence of functional groups not present in the target chemicals (e.g., phenols, halogens, carboxylic acids); *N*-alkyl-substituted amines and acetamide derivatives of aromatic amines. In addition, nitro amines and dinitro compounds were not selected. Each of these attributes introduce significant differences in bioavailability, reactivity, and applicable metabolic pathways relative to 2,3-TDA. Additionally, ar-methyl-1,3-benzenediamine (CASRN 25376-45-8) was not selected for further review because it can exist as a mixture of two TDA isomers, in which the location of the methyl on the aromatic ring is not defined.

The remaining eight possible analogues for 2,3-TDA are listed in Table C-1. The existence of a cancer risk estimate and/or a weight-of-evidence (WOE) determination for cancer is indicated for each analogue. Compounds are grouped with their respective simple salts, which were identified by Dice only. Salts did not cluster with free acids in ChemACE because it is fragment-based; therefore, salts and free acids have different fragments and will not cluster without special adjustment (i.e., modify the Simplified Molecular Input Line Entry System [SMILES] being clustered so that representative free acid structures are entered for salts). The analogue results from Dice are based on SMILES arbitrary target specification (SMARTS) substructure searching, allowing for identification of both free acid and respective salt analogues.

Table C-1. Summary of Cancer Assessment Information for Analogues of

a Found by ChemACE.

^bFound by Dice.

 $IUR =$ inhalation unit risk; $OSF =$ oral slope factor; $p-OSF =$ provisional oral slope factor; $TDA =$ toluenediamine; $WOE = weight of evidence.$

2,3-Dimethylbenzene-1,4-diamine, which lacks a cancer risk estimate or WOE determination (highlighted in gray in Table C-1), was not further considered as a potential analogue for the screening evaluation of potential carcinogenicity of 2,3-TDA. Compounds selected for further consideration were 2,4-, 2,5-, and 2,6-TDA and their simple salts.

STEP 3. COMPARISON OF THE EXPERIMENTAL GENOTOXICITY DATA FOR 2,3-TDA AND ANALOGUES

The limited genotoxicity data available for 2,3-TDA are described in detail in the "Other Data" section in the main body of this report. The data indicate that 2,3-TDA is a weak mutagen but does not cause chromosomal aberrations (CAs) following in vivo exposure. A summary of the genotoxicity data for the structural analogues, 2,4-, 2,5-, and 2,6-TDA, is provided below for comparative purposes.

2,4-, 2,5-, and 2,6-TDA are mutagenic to *Salmonella typhimurium* in the presence of metabolic activation [\(U.S. EPA, 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [ECHA, 2008;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). Sex-linked recessive mutations were observed in *Drosophila melanogaster* exposed to 2,4-TDA [\(ECHA, 2008\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467). However, 2,4-, 2,5-, and 2,6-TDA were generally nonmutagenic to mammalian cells in vitro or in vivo [\(U.S. EPA, 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [ECHA, 2008;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682).

2,4-, 2,5-, and 2,6-TDA show evidence of in vitro clastogenicity in mammalian cells, both with and without metabolic activity. CAs were induced by 2,4- and 2,5-TDA, sister chromatid exchanges (SCEs) were induced by 2,4-TDA, and micronuclei (MN) were induced by 2,6-TDA [\(U.S. EPA, 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [ECHA, 2008;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). Induction of MN in bone marrow or hepatocytes was generally not observed following in vivo exposure to 2,4-, 2,5-, or 2,6-TDA. However, weak induction of MN in bone marrow following exposure to 2,4- or 2,6-TDA was reported in some studies [\(Takasawa et al., 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229089) [U.S. EPA, 2013;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [ECHA, 2008;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [U.S. EPA,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682) [2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682).

The majority of in vitro studies indicate that 2,4-, 2,5-, and 2,6-TDA are capable of damaging mammalian deoxyribonucleic acid (DNA). Results were most consistent with 2,4-TDA, which induced DNA damage and/or unscheduled DNA synthesis (UDS) in human skin fibroblasts, human hepatocytes, and primary rat hepatocytes, and formed DNA adducts in rat hepatocytes and purified calf thymus DNA [\(ECHA, 2008\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467). 2,5-TDA also induced DNA damage in rat and hamster hepatocytes [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). UDS was observed in primary cultured human hepatocytes exposed to 2,6-TDA, but not primary rat hepatocytes [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). Low levels of covalent binding to DNA were observed for 2,6-TDA [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). DNA strand breaks and UDS were consistently reported in rodents following in vivo exposure to 2,4-TDA [\(ECHA,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [2008\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467), but results in rodents exposed to 2,5- or 2,6-TDA were mixed [\(U.S. EPA, 2013,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). DNA adducts were observed in multiple organs following in vivo exposure to 2,4-, but not 2,6-TDA [\(ECHA, 2008;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442467) [U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). 2,5- and 2,6-TDA induced cell transformation in hamster embryo cells [\(U.S. EPA, 2013,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821) [2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682).

In summary, 2,3-TDA and its analogues show some evidence of mutagenicity in bacterial systems with metabolic activation. Additionally, 2,4-, 2,5-, and 2,6-TDA display potential for causing clastogenic effects and DNA damage in mammalian cells under certain conditions. No genotoxic effects were reported for 2,3-TDA in studies using mice but the evidence is too limited to adequately assess its genotoxic potential.

STEP 4. TOXICOKINETICS OF 2,3-TDA AND ANALOGUES

The toxicokinetics of 3,4-, 2,4-, 2,5-, and 2,6-TDA are briefly described in Table C-2 (see additional information in Table A-3). Experimental data indicate that 2,4-, 2,5-, and 2,6-TDA are rapidly absorbed following oral exposure and excreted in the urine (see Table A-3). The primary metabolic pathways for 2,4-, 2,5-, and 2,6-TDA include acetylation of the amino groups and ring hydroxylation with some evidence of oxidation of the methyl group (see Table A-3). No toxicokinetic data are available for 2,3-TDA, but similar metabolic pathways are expected for the target compound based on a comparative in silico metabolism analysis (see section on "Metabolic Analogues" in Appendix A and Appendix D for additional details).

 $NA = not applicable; ND = no data; TDA = tolerance.$

STEP 5. CARCINOGENICITY OF 2,3-TDA ANALOGUES AND MODE-OF-ACTION DISCUSSION

U.S. EPA cancer WOE descriptors for 2,3-TDA and its analogue compounds are shown in Table C-3. As noted in the main PPRTV document, there is inadequate information to assess the carcinogenic potential of 2,3-TDA. The analogue 2,5-TDA is characterized by U.S. EPA as having evidence of carcinogenic potential. Under the 2005 *Guidelines for Carcinogen Risk Assessment* [\(U.S. EPA, 2005a\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6324329), there is *"Suggestive Evidence of Carcinogenic Potential"* for 2,5-TDA [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). The U.S. EPA has not assessed the potential carcinogenicity of 2,4-TDA [\(U.S. EPA, 1991\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4731470); however, this compound is listed as a carcinogen by [CalEPA](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2215636) (2011b), considered *possibly carcinogenic to humans* by [IARC \(1987\),](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4731469) and reasonably *anticipated to be a human carcinogen* by [NTP \(2016b\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442462) The U.S. EPA determined that there is *"Inadequate Information to Assess Carcinogenic Potential"* for 2,6-TDA [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). Oral slope factor (OSF) values varied by an order of magnitude, with the highest potency value calculated for 2,4-TDA (4 \times 10⁰ [mg/kg-day]⁻¹) [\(CalEPA, 2011b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2215636) and the lowest (screening) potency value for 2,5-TDA (1.8×10^{-1} [mg/kg-day]⁻¹ as a sulfate) [\(U.S. EPA, 2013\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257821). Exposure-related increases were observed in liver tumors in male and female rats and female mice, subcutaneous fibromas in male rats, mammary tumors in female rats, and lymphoma in female mice following dietary 2,4-TDA exposure [\(NTP, 2016b;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442462) [CalEPA, 2011b;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2215636) [IARC, 1987\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4731469). Testicular tumors were observed in male rats and lung tumors were observed in female mice following dietary exposure to 2,5-TDA $(U.S. EPA, 2013)$. Potential carcinogenic effects of

2,6-TDA were evaluated in rats and mice in 2-year feeding studies [\(U.S. EPA, 2005b;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682) [NTP,](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229145) [1980\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=2229145). Dose-related trends for increased incidence of hepatocellular carcinomas and islet-cell adenomas of the pancreas were observed in male rats, a slight increase in vascular neoplasm of the spleen and liver and a significant trend in increased lymphomas were observed in male mice, and a significant trend for increased hepatocellular carcinomas was reported in female mice [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). The study authors did not consider the neoplastic lesions observed with exposure to 2,6-TDA to be treatment-related due to the absence of statistically significant effects in any treatment group compared to controls, but it was unclear whether exposure levels were adequate to assess carcinogenic potential [\(U.S. EPA, 2005b\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=1257682). The carcinogenic mode of action (MOA) has not been established for 2,4- or 2,5-TDA, although both compounds (along with 2,6-TDA and the target compound, 2,3-TDA) exhibit some evidence of genotoxicity (see "Step 3. Comparison of the Experimental Genotoxicity Data for 2,3-TDA and Analogues" above for more information).

^aThere is no U.S. EPA WOE descriptor for 2,4-TDA; however, this compound is listed as a carcinogen by CalEPA (2011a), considered *possibly carcinogenic to humans* b[y IARC \(1987\),](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4731469) and *reasonably anticipated to be a human carcinogen* by [NTP \(2016b\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4442462)

^bOSF derived b[y CalEPA \(2011a\).](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=4003790)

 $BMDL_{10} = 10\%$ benchmark dose lower confidence limit; $F = female(s)$; HED = human equivalent dose; $M = male(s)$; $NA = not applicable$; $ND = no data$; $OSF = oral slope factor$; $POD = point of departure$; $p-OSF =$ provisional oral slope factor; $TDA =$ toluenediamine; $WOE =$ weight of evidence.

STEP 6. STRUCTURAL ALERTS AND STRUCTURE-ACTIVITY RELATIONSHIP PREDICTIONS FOR 2,3-TDA AND ANALOGUES

Structural alerts (SAs) and predictions for genotoxicity and carcinogenicity were identified using computational tools as described in Appendix B. The model results for 2,3-TDA and its analogue compounds are shown in Table C-4. Concerns for carcinogenicity and/or mutagenicity of 2,3-TDA and its analogues were indicated by several models within each predictive tool (see Table C-4). Table C-5 provides a list of the specific SAs that underlie the findings of a concern for carcinogenicity or mutagenicity in Table C-4.

OECD QSAR Toolbox models showed a concern for mutagenicity, CAs, MN, and protein binding for 2,3-TDA and all analogues based on SAs (see Table C-4 and Table C-5). The ToxRead and VEGA models also indicated a concern for mutagenicity for 2,3-TDA and all analogues. The Toxtree tool indicated a concern for 2,4- and 2,5-TDA mutagenicity in *S. typhimurium* TA100 but indicated that 2,3- and 2,6-TDA were unlikely to be mutagenic in *S. typhimurium* TA100. The Toxtree results for 2,3- and 2,6-TDA are inconsistent with positive experimental data (see "Step 3. Comparison of the Experimental Genotoxicity Data for 2,3-TDA and Analogues" above for more information) and the results of the other QSAR models.

^aAll tools and models described in Appendix B were used. Models with results or alerts are presented in the heat map (models without results were omitted).

ANTARES = Alternative Non-Testing Methods Assessed for REACH Substances; CA = chromosomal aberration; CAESAR = Computer-Assisted Evaluation of industrial chemical Substances According to Regulations; CONSENSUS = consensus assessment based on multiple models (CAESAR, SARpy, ISS, and *k*-NN); DNA = deoxyribonucleic acid; IRFMN = Istituto di Ricerche Farmacologiche Mario Negri; ISS = Istituto Superiore di Sanità; ISSCAN-CGX = Istituto Superiore di Sanità Chemical Carcinogen; *k-*NN = *k*-nearest neighbor; OECD = Organisation for Economic Co-operation and Development; SAR = structure-activity relationship; QSAR = quantitative structure-activity relationship; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

a The SA in OncoLogic for 2,6-TDA dihydrochloride was reported as "marginal."

OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; $SA =$ structural alert; $TDA =$ toluenediamine.

OECD QSAR Toolbox models showed a concern for carcinogenicity for 2,3-TDA and all analogues based on SAs (see Table C-4 and Table C-5). The level of carcinogenicity concern in OncoLogic for 2,3-TDA was "high-moderate" based on structure-activity relationship (SAR) analysis only (aromatic amine with amino groups *ortho* to one another). OncoLogic indicated the level of concern for carcinogenicity as "moderate" for 2,4-TDA based on animal carcinogenicity data and SAR analysis (aromatic amine with amino groups *meta* to one another). The level of carcinogenicity concern in OncoLogic for 2,6-TDA, 2,5-TDA, 2,5-TDA hydrochloride, 2,5-TDA sulfate and 2,4-TDA dihydrochloride was "moderate" based on SAR analysis only (aromatic amine with amino groups *meta* or *para* to one another). OncoLogic reported a "marginal" level of concern for 2,6-TDA dihydrochloride (shown as no results for model in Table C-4) based on a lack of evidence of carcinogenicity from animal studies and SAR analysis (aromatic amine with amino groups *meta* to one another). VEGA showed concern for carcinogenicity of 2,3-TDA using the CAESAR and ISS models (no data for the IRFMN/ANTARES or IRFMN/ISSCAN-CGX models). All four VEGA models showed concern for carcinogenicity for 2,4-TDA and 2,4-TDA dihydrochloride. Carcinogenicity models in VEGA produced inconsistent results for 2,5- and 2,6-TDA (and their salts). While the CAESAR model showed concern for both compounds (and their salts), the ISS model showed concern only for 2,5-TDA (and its salts), and the IRFMN/ISSCAN-CGX model did not show concern for either compound (or their salts). There were no data for the IRFMN/ANTARES model for 2,5- or 2,6-TDA (or their salts). The Toxtree tool indicated that 2,3-, 2,4-, and 2,6-TDA were potential carcinogens based on QSAR, but 2,5-TDA was not. The Toxtree tool showed there was no concern for nongenotoxic carcinogenicity for 2,3-TDA or any of its analogues.

The ToxAlerts tool showed a concern for genotoxic carcinogenicity and/or mutagenicity for 2,3-TDA and all analogues based on various SAs (see Table C-4 and Table C-5). The Toxtree models also suggest a concern for genotoxic carcinogenicity for 2,3-TDA and all analogues based on SAs.

In general, SAR predictions indicate a concern for genotoxicity and carcinogenicity for 2,3-TDA and TDA analogues across several software systems evaluated. Moreover, a clear pattern or relationship between the position of the amino groups (2,3-TDA is a *o*- isomer, 2,5- is a *p*- isomer, and 2,4- and 2,6- are *m*- isomers) and potential differences in SAR predictions are not apparent for the TDA compounds. Previous SAR evaluations have suggested enhanced chemical reactivity for the *o*- and *p*-substituted aromatic amines due to quinone formation [\(Bajot](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128) [et al., 2010\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=7343128). However, based on the available experimental and in silico data discussed above, the influence of the position of the amino groups on the potential genotoxicity and carcinogenicity of the TDA compounds is unclear.

STEP 7. EVIDENCE INTEGRATION FOR SCREENING EVALUATION OF 2,3-TDA CARCINOGENICITY

Table C-6 presents the data for multiple lines of evidence pertinent to the screening evaluation of the carcinogenic potential of 2,3-TDA.

ADME = absorption, distribution, metabolism, and excretion; CA = chromosomal aberration; DNA = deoxyribonucleic acid; *m-* = *meta*; MOA = mode of action; ND = no data; o - = *ortho*; p - = para; QSAR = quantitative structure-activity relationship; SA = structural alert; SAR = structure-activity relationship; TDA = toluenediamine; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

STEP 8. QUALITATIVE LEVEL OF CONCERN FOR 2,3-TDA POTENTIAL CARCINOGENICITY

A *concern for potential carcinogenicity* for 2,3-TDA is identified based on the multiple lines of evidence, including similarities in structural features, in silico metabolism profiles, SAs, and SAR predictions, and experimental data for carcinogenicity and/or genotoxicity for the target and analogues (see Table C-7 for additional details). Because of the lack of information supporting carcinogenic portal-of-entry effects, the qualitative level of concern for this chemical is considered to be applicable to all routes of exposure.

 MOA = mode of action; NA = not applicable; SA = structural alert; SAR = structure-activity relationship; TDA = toluenediamine.

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APPENDIX D. METHODOLOGY AND RESULTS FOR IN SILICO METABOLITE ANALYSIS OF TARGET AND ANALOGUES

An in silico analysis of metabolism was conducted for 2,3-toluenediamine (2,3-TDA) and its analogues using different software tools. The main objective of this analysis is to provide a qualitative comparison of metabolite predictions for TDA compounds in the absence of experimental data for the target. The focus is on the major metabolism pathways characterized in the literature, highlighting any notable differences between the target and analogues.

Chemical structures were extracted from the U.S. EPA CompTox Chemicals Dashboard for 2,3-TDA and the identified structural analogues [3,4-, 2,4-, 2,5-, and 2,6-TDA; [U.S. EPA](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=5794424) (2019)]. The metabolite predictions for the chemicals of interest were generated using commercially available software systems, including the Tissue Metabolism Simulator (TIMES) [\(Dimitrov et al., 2005;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147) [Mekenyan et al., 2004\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129) and Meteor Nexus [\(Marchant et al., 2008\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=613223). A structure data file (SDF) was imported into the TIMES program (Version 2.29.1; [http://oasis-lmc.org/products/software/times.aspx\)](http://oasis-lmc.org/products/software/times.aspx), using the in vitro rat S9 metabolic simulator (Version 11.16) and the rat in vivo metabolic simulator (Version 07.12) to make predictions of likely metabolites. The predictions were exported as a .txt file for subsequent processing.

For the Meteor Nexus predictions, the SDF was split into separate molecular data (MOL) files for batch processing in Meteor Nexus. A python script (Python; Version 3.6.5; [python.org\)](http://www.python.org/) was used to split the SDF, and a second script was used to concatenate the individual substance prediction files that were created as separate excel workbooks. Default settings were used in Meteor Nexus (Version 3.1.0) developed by Lhasa Limited

[\(https://www.lhasalimited.org/library/publishing.htm\)](https://www.lhasalimited.org/library/publishing.htm). The settings were for a maximum depth of tree to be 3, for the maximum number of metabolites to be capped at 1,000, and for the scoring method to be Site of Metabolism Scoring (with Molecular Mass Variance). The results are described as a score that uses experimental data for compounds that match the same biotransformation, have similar molecular weights, and are structurally similar around the site of metabolism to the query compound (for more details, see

[https://www.lhasalimited.org/products/meteor-reasoning-methodologies.htm\)](https://www.lhasalimited.org/products/meteor-reasoning-methodologies.htm). The prediction files were then processed further within a Jupyter notebook [\(jupyter.org\)](http://www.jupyter.org/) imported with python libraries RDKit Version 2018.03.2.0 [\(RDKit.org\)](http://rdkit.org/), Pandas (Version 0.23.1; [pandas.pydata.org\)](http://pandas.pydata.org/), NumPy (Version 1.14.3; [numpy.org\)](https://numpy.org/), and Matplotlib (Version 2.2.2; [matplotlib.org\)](http://matplotlib.org/).

The software systems provided Simplified Molecular Input Line Entry System (SMILES) representations for the predicted metabolites. These were converted into RDKit mol objects and exported as a Pandas Tools worksheet that provided depictions of chemical structure. International Union of Pure and Applied Chemistry (IUPAC) International Chemical Identifier (InChI™) keys were created using RDKit because SMILES representations are not unique. The structures of the predicted metabolites from each of the tools evaluated are presented in Table D-1, which compares metabolites identified across the different software tools and experimental data from in vivo animal studies captured in Table A-3 of the "Metabolic Analogues" section in Appendix A. Additionally, pathway transformations corresponding to the metabolite predictions were extracted from Meteor Nexus to facilitate similarity comparisons between the target and candidate analogues. Other software tools (i.e., TIMES) did not provide the same level of information. The pathway transformations for target and candidate analogues were extracted from the Meteor Nexus summary report, then grouped by substance and pivoted

to provide a representation of the TDA compounds as rows and the unique pathways as columns (see Table A-4 in Appendix A). A pairwise distance matrix was then computed using the Jaccard distance as a metric, which was then transformed to a similarity matrix (see Figure A-1 in Appendix A). A metabolic tree for 2,4-TDA was constructed to highlight the relationships of predicted metabolites for this specific analogue that correspond to the pathway transformations shared among the TDA compounds (see Figure D-1).

^a1/0 denotes whether a metabolite was identified/not identified by a software tool or experimental animal data captured in Table A-3.

^bMeteor Nexus [\(Dimitrov et al., 2005;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147) [Mekenyan et al., 2004\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129).

^cIn Vivo/In Vitro Rat Tissue Metabolism Simulator [\(Dimitrov et al., 2005;](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147) [Mekenyan et al., 2004\)](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129). d Metabolites reported from in vivo animal studies for the TDA isomers. Experimental data for 2,3- and 3,4-TDA were not available. Refer to Table A-3 for additional details.

InChI = IUPAC International Chemical Identifier; IUPAC = International Union of Pure and Applied Chemistry; NDr = not determined; SMILES = Simplified Molecular Input Line Entry System; TDA = toluenediamine; TIMES = Tissue Metabolism Simulator.

Figure D-1. Metabolic Tree for the 2,4-Toluenediamine (CASRN 95-80-7) Analogue. Diagram displays the relationship of the metabolites identified from Meteor Nexus (**[Dimitrov et al., 2005](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318147)**; **[Mekenyan et al., 2004](https://hero.epa.gov/hero/index.cfm?action=search.view&reference_id=6318129)**) to the parent compound and notes the corresponding pathway transformations (a−g).

APPENDIX E. REFERENCES

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