EPA/690/R-21/005F | August 2021 | FINAL



Provisional Peer-Reviewed Toxicity Values for

1-Bromo-2-Chloroethane (CASRN 107-04-0)



U.S. EPA Office of Research and Development Center for Public Health and Environmental Assessment



Provisional Peer-Reviewed Toxicity Values for

1-Bromo-2-Chloroethane (CASRN 107-04-0)

Center for Public Health and Environmental Assessment Office of Research and Development U.S. Environmental Protection Agency Cincinnati, OH 45268

AUTHORS, CONTRIBUTORS, AND REVIEWERS

CHEMICAL MANAGER

Jeffry L. Dean II, PhD Center for Public Health and Environmental Assessment, Cincinnati, OH

CONTRIBUTOR

Laura M. Carlson, PhD Center for Public Health and Environmental Assessment, Research Triangle Park, NC

Roman Mezencev, PhD Center for Public Health and Environmental Assessment, Washington, DC

SCIENTIFIC TECHNICAL LEAD

Jeffry L. Dean II, PhD Center for Public Health and Environmental Assessment, Cincinnati, OH

J. Phillip Kaiser, PhD, DABT Center for Public Health and Environmental Assessment, Cincinnati, OH

Q. Jay Zhao, PhD, MPH, DABT Center for Public Health and Environmental Assessment, Cincinnati, OH

DRAFT DOCUMENT PREPARED BY

SRC, Inc. 7502 Round Pond Road North Syracuse, NY 13212

PRIMARY INTERNAL REVIEWERS

Andrew Kraft, PhD Center for Public Health and Environmental Assessment, Washington, DC

Paul G. Reinhart, PhD, DABT Center for Public Health and Environmental Assessment, Research Triangle Park, NC

PRIMARY EXTERNAL REVIEWERS

Eastern Research Group, Inc. 110 Hartwell Avenue Lexington, MA 02421-3136

PPRTV PROGRAM MANAGEMENT

Teresa L. Shannon Center for Public Health and Environmental Assessment, Cincinnati, OH J. Phillip Kaiser, PhD, DABT Center for Public Health and Environmental Assessment, Cincinnati, OH

Questions regarding the content of this PPRTV assessment should be directed to the U.S. EPA Office of Research and Development (ORD) Center for Public Health and Environmental Assessment (CPHEA) website at https://ecomments.epa.gov/pprtv.

TABLE OF CONTENTS

| COMMONLY | USED ABBREVIATIONS AND ACRONYMS | v |
|------------------|---|----|
| | ۱D | |
| QUALITY AS | SURANCE | 1 |
| DISCLAIMER | S | 1 |
| QUESTIONS | REGARDING PPRTVs | 2 |
| 1. INTRODUC | CTION | 3 |
| 2. REVIEW O | F POTENTIALLY RELEVANT DATA (NONCANCER AND CANCER) | 7 |
| | AN STUDIES 1 | |
| 2.1.1. | Oral Exposures1 | 0 |
| 2.1.2. | Inhalation Exposures1 | 0 |
| 2.2. ANIN | IAL STUDIES 1 | 0 |
| 2.2.1. | Inhalation Exposures1 | 0 |
| | Oral Exposures1 | |
| | ER DATA (SHORT-TERM TESTS, OTHER EXAMINATIONS) 1 | |
| 2.3.1. | Acute Toxicity in Animals1 | 0 |
| | Genotoxicity1 | |
| | Absorption, Distribution, Metabolism, and Excretion Studies 1 | |
| | ON OF PROVISIONAL VALUES 1 | |
| | VATION OF PROVISIONAL REFERENCE DOSES 1 | |
| 3.2. DERI | VATION OF PROVISIONAL REFERENCE CONCENTRATIONS 1 | 8 |
| 3.3. SUM | MARY OF NONCANCER PROVISIONAL REFERENCE VALUES 1 | 8 |
| | CER WEIGHT-OF-EVIDENCE DESCRIPTOR 1 | |
| | VATION OF PROVISIONAL CANCER RISK ESTIMATES | |
| APPENDIX A | . SCREENING NONCANCER PROVISIONAL VALUES | 21 |
| APPENDIX B | BACKGROUND AND METHODOLOGY FOR THE SCREENING | |
| | ON OF POTENTIAL CARCINOGENICITY | 54 |
| | RESULTS OF THE SCREENING EVALUATION OF POTENTIAL | |
| | GENICITY | |
| APPENDIX D | . REFERENCES |)1 |

COMMONLY USED ABBREVIATIONS AND ACRONYMS

| 2 | | LD | |
|-----------|---|--------------------|--|
| α2u-g | alpha 2u-globulin | LD ₅₀ | median lethal dose |
| ACGIH | American Conference of Governmental | LOAEL | lowest-observed-adverse-effect level |
| | Industrial Hygienists | MN | micronuclei |
| AIC | Akaike's information criterion | MNPCE | micronucleated polychromatic |
| ALD | approximate lethal dosage | | erythrocyte |
| ALT | alanine aminotransferase | MOA | mode of action |
| AR | androgen receptor | MTD | maximum tolerated dose |
| AST | aspartate aminotransferase | NAG | N-acetyl-β-D-glucosaminidase |
| atm | atmosphere | NCI | National Cancer Institute |
| ATSDR | Agency for Toxic Substances and | NOAEL | no-observed-adverse-effect level |
| | Disease Registry | NTP | National Toxicology Program |
| BMD | benchmark dose | NZW | New Zealand White (rabbit breed) |
| BMDL | benchmark dose lower confidence limit | OCT | ornithine carbamoyl transferase |
| BMDS | Benchmark Dose Software | ORD | Office of Research and Development |
| BMR | benchmark response | PBPK | physiologically based pharmacokinetic |
| BUN | blood urea nitrogen | PCNA | proliferating cell nuclear antigen |
| BW | body weight | PND | postnatal day |
| CA | chromosomal aberration | POD | point of departure |
| CAS | Chemical Abstracts Service | POD _{ADJ} | duration-adjusted POD |
| CASRN | Chemical Abstracts Service registry | QSAR | quantitative structure-activity |
| | number | | relationship |
| CBI | covalent binding index | RBC | red blood cell |
| СНО | Chinese hamster ovary (cell line cells) | RDS | replicative DNA synthesis |
| CL | confidence limit | RfC | inhalation reference concentration |
| CNS | central nervous system | RfD | oral reference dose |
| CPHEA | Center for Public Health and | RGDR | regional gas dose ratio |
| | Environmental Assessment | RNA | ribonucleic acid |
| CPN | chronic progressive nephropathy | SAR | structure-activity relationship |
| CYP450 | cytochrome P450 | SCE | sister chromatid exchange |
| DAF | dosimetric adjustment factor | SD | standard deviation |
| DEN | diethylnitrosamine | SDH | sorbitol dehydrogenase |
| DMSO | dimethylsulfoxide | SE | standard error |
| DNA | deoxyribonucleic acid | SGOT | serum glutamic oxaloacetic |
| EPA | Environmental Protection Agency | | transaminase, also known as AST |
| ER | estrogen receptor | SGPT | serum glutamic pyruvic transaminase, |
| FDA | Food and Drug Administration | | also known as ALT |
| FEV_1 | forced expiratory volume of 1 second | SSD | systemic scleroderma |
| GD | gestation day | TCA | trichloroacetic acid |
| GDH | glutamate dehydrogenase | TCE | trichloroethylene |
| GGT | γ-glutamyl transferase | TWA | time-weighted average |
| GSH | glutathione | UF | uncertainty factor |
| GST | glutathione-S-transferase | UFA | interspecies uncertainty factor |
| Hb/g-A | animal blood-gas partition coefficient | UFc | composite uncertainty factor |
| Hb/g-H | human blood-gas partition coefficient | UFD | database uncertainty factor |
| HEC | human equivalent concentration | UF _H | intraspecies uncertainty factor |
| HED | human equivalent dose | UFL | LOAEL-to-NOAEL uncertainty factor |
| i.p. | intraperitoneal | UFs | subchronic-to-chronic uncertainty factor |
| IRIS | Integrated Risk Information System | U.S. | United States of America |
| IVF | in vitro fertilization | WBC | white blood cell |
| LC_{50} | median lethal concentration | W DC | white blood cell |
| LC30 | meanan remai concentration | | |

Abbreviations and acronyms not listed on this page are defined upon first use in the PPRTV document.

DRAFT PROVISIONAL PEER-REVIEWED TOXICITY VALUES FOR 1-BROMO-2-CHLOROETHANE (CASRN 107-04-0)

BACKGROUND

A Provisional Peer-Reviewed Toxicity Value (PPRTV) is defined as a toxicity value derived for use in the Superfund Program. PPRTVs are derived after a review of the relevant scientific literature using established U.S. Environmental Protection Agency (U.S. EPA) guidance on human health toxicity value derivations.

The purpose of this document is to provide support for the hazard and dose-response assessment pertaining to chronic and subchronic exposures to substances of concern, to present the major conclusions reached in the hazard identification and derivation of the PPRTVs, and to characterize the overall confidence in these conclusions and toxicity values. It is not intended to be a comprehensive treatise on the chemical or toxicological nature of this substance.

Currently available PPRTV assessments can be accessed on the U.S. EPA's PPRTV website at <u>https://www.epa.gov/pprtv</u>. PPRTV assessments are eligible to be updated on a 5-year cycle and revised as appropriate to incorporate new data or methodologies that might impact the toxicity values or affect the characterization of the chemical's potential for causing adverse human-health effects. Questions regarding nomination of chemicals for update can be sent to the appropriate U.S. EPA Superfund and Technology Liaison (<u>https://www.epa.gov/research/fact-sheets-regional-science</u>).

QUALITY ASSURANCE

This work was conducted under the U.S. EPA Quality Assurance (QA) program to ensure data are of known and acceptable quality to support their intended use. Surveillance of the work by the assessment managers and programmatic scientific leads ensured adherence to QA processes and criteria, as well as quick and effective resolution of any problems. The QA manager, assessment managers, and programmatic scientific leads have determined under the QA program that this work meets all U.S. EPA quality requirements. This PPRTV was written with guidance from the CPHEA Program Quality Assurance Project Plan (PQAPP), the QAPP titled *Program Quality Assurance Project Plan (PQAPP) for the Provisional Peer-Reviewed Toxicity Values (PPRTVs) and Related Assessments/Documents (L-CPAD-0032718-QP)*, and the PPRTV development contractor QAPP titled *Quality Assurance Project Plan—Preparation of Provisional Toxicity Value (PTV) Documents (L-CPAD-0031971-QP)*. As part of the QA system, a quality product review is done prior to management clearance. A Technical Systems Audit may be performed at the discretion of the QA staff.

All PPRTV assessments receive internal peer review by at least two CPHEA scientists and an independent external peer review by at least three scientific experts. The reviews focus on whether all studies have been correctly selected, interpreted, and adequately described for the purposes of deriving a provisional reference value. The reviews also cover quantitative and qualitative aspects of the provisional value development and address whether uncertainties associated with the assessment have been adequately characterized.

DISCLAIMERS

The PPRTV document provides toxicity values and information about the adverse effects of the chemical and the evidence on which the value is based, including the strengths and

limitations of the data. All users are advised to review the information provided in this document to ensure that the PPRTV used is appropriate for the types of exposures and circumstances at the site in question and the risk management decision that would be supported by the risk assessment.

Other U.S. EPA programs or external parties who may choose to use PPRTVs are advised that Superfund resources will not generally be used to respond to challenges, if any, of PPRTVs used in a context outside of the Superfund program.

This document has been reviewed in accordance with U.S. EPA policy and approved for publication. Mention of trade names or commercial products does not constitute endorsement or recommendation for use.

QUESTIONS REGARDING PPRTVs

Questions regarding the content of this PPRTV assessment should be directed to the U.S. EPA ORD CPHEA website at <u>https://ecomments.epa.gov/pprtv</u>.

1. INTRODUCTION

1-Bromo-2-chloroethane (CASRN 107-04-0) belongs to the class of compounds known as halogenated alkanes. 1-Bromo-2-chloroethane currently has no commercial uses, but small amounts may be imported and used for research and development (U.S. EPA, 2015b). It is listed under the U.S. Environmental Protection Agency (U.S. EPA) Toxic Substances Control Act (TSCA) Significant New Use Rule (SNUR), which requires notification of the U.S. EPA prior to manufacturing, importing, or processing of a chemical substance for a significant new use (e.g., commercial purpose) (U.S. EPA, 2015c). Former uses of 1-bromo-2-chloroethane include as a solvent for cellulose esters and ethers, in organic synthesis, and as a fumigant for fruits and vegetables (Lewis and Hawley, 2007). 1-Bromo-2-chloroethane is listed on TSCA's public inventory (U.S. EPA, 2015a), but it is not registered with Europe's Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH) program (ECHA, 2015).

1-Bromo-2-chloroethane can be produced by the reaction of bromine and chlorine on ethylene gas (Lewis and Hawley, 2007). However, it is no longer produced in the United States because of potential carcinogenic activity based on its in vitro mutagenicity and the carcinogenicity of the structurally similar chemicals 1,2-dibromoethane and 1,2-dichloroethane (U.S. EPA, 2015b).

The empirical formula for 1-bromo-2-chloroethane is C₂H₄BrCl, and its structure is shown in Figure 1. Table 1 summarizes its physicochemical properties. 1-Bromo-2-chloroethane is a clear, colorless liquid with a sweet chloroform-like odor at room temperature (NOAA, 2015). Its high vapor pressure indicates that it will exist solely as a vapor in the atmosphere. Given its vapor pressure and moderate estimated Henry's law constant, it is likely to volatilize from either dry or moist soil surfaces and from water surfaces. The high water solubility and low soil adsorption coefficient indicate that it will have the potential to leach to groundwater or undergo runoff after a rain event. However, volatilization to the atmosphere is expected to be the main transport pathway. Estimated hydrolysis half-lives of 3 and 32 years at pH 8 and 7, respectively, indicate that hydrolysis is not likely to be an important fate process.

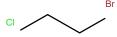


Figure 1. 1-Bromo-2-Chloroethane (CASRN 107-04-0) Structure

| Property (unit) | Value |
|--|---|
| Physical state | Liquid |
| Boiling point (°C) | 107 ^a |
| Melting point (°C) | -16.3ª |
| Density (g/cm ³) | 1.63 (predicted average) ^a |
| Vapor pressure (mm Hg at 20°C) | 33.1ª |
| pH (unitless) | NA |
| pKa (unitless) | NA |
| Solubility in water (mol/L) | $4.79 	imes 10^{-2}$ a |
| Octanol-water partition coefficient (log Kow) | 1.86 (predicted average) ^a |
| Henry's law constant (atm-m ³ /mol at 25°C) | $9.09 	imes 10^{-4}$ a |
| Soil adsorption coefficient Koc (L/kg) | 55.1 (predicted average) ^a |
| Atmospheric OH rate constant (cm ³ /molecule-sec at 25°C) | 3.22×10^{-13} (predicted average) ^a |
| Atmospheric half-life (days) | 42 (estimated) ^c |
| Relative vapor density (air = 1) | 4.94 ^b |
| Molecular weight (g/mol) | 143ª |
| Flash point (closed cup, °C) | NA |

Table 1. Physicochemical Properties of 1-Bromo-2-Chloroethane (CASRN 107-04-0)

^aU.S. EPA CompTox Chemicals Dashboard (as of June 2021;

https://comptox.epa.gov/dashboard/dsstoxdb/results?search=DTXSID4024775#properties).

^b<u>NOAA (2015)</u>. ^c<u>U.S. EPA (2012)</u>.

NA = not applicable.

No toxicity values are available for 1-bromo-2-chloroethane from U.S. EPA or other agencies/organizations (see Table 2).

| Source ^a | Value | Notes | Reference ^b |
|---------------------|-------|-------|----------------------------|
| Noncancer | | | |
| IRIS | NV | NA | <u>U.S. EPA (2020c)</u> |
| HEAST | NV | NA | <u>U.S. EPA (2011b)</u> |
| DWSHA | NV | NA | <u>U.S. EPA (2018)</u> |
| ATSDR | NV | NA | <u>ATSDR (2018)</u> |
| IPCS | NV | NA | <u>IPCS (2020)</u> |
| CalEPA | NV | NA | <u>CalEPA (2019)</u> |
| OSHA | NV | NA | OSHA (2020a); OSHA (2020b) |
| NIOSH | NV | NA | <u>NIOSH (2018)</u> |
| ACGIH | NV | NA | <u>ACGIH (2020)</u> |
| HEEP | NV | NA | <u>U.S. EPA (1985)</u> |
| Cancer | | | |
| IRIS | NV | NA | <u>U.S. EPA (2020c)</u> |
| HEAST | NV | NA | <u>U.S. EPA (2011b)</u> |
| DWSHA | NV | NA | <u>U.S. EPA (2018)</u> |
| NTP | NV | NA | <u>NTP (2016a)</u> |
| IARC | NV | NA | <u>IARC (2018)</u> |
| CalEPA | NV | NA | <u>CalEPA (2019)</u> |
| ACGIH | NV | NA | ACGIH (2020) |

Table 2. Summary of Available Toxicity Values for 1-Bromo-2-Chloroethane

^aSources: ACGIH = American Conference of Governmental Industrial Hygienists; ATSDR = Agency for Toxic Substances and Disease Registry; CalEPA = California Environmental Protection Agency; DWSHA = Drinking Water Standards and Health Advisories; HEAST = Health Effects Assessment Summary Tables; HEEP = Health and Environmental Effects Profile; IARC = International Agency for Research on Cancer; IPCS = International Programme on Chemical Safety; IRIS = Integrated Risk Information System; NIOSH = National Institute for Occupational Safety and Health; NTP = National Toxicology Program; OSHA = Occupational Safety and Health Administration.

^bReference date is the publication date for the database and not the date the source was accessed.

NA = not applicable; NV = not available.

Non-date-limited literature searches were conducted in November 2017 and updated in May 2021, for studies relevant to the derivation of provisional toxicity values for 1-bromo-2-chloroethane (CASRN 107-04-0). Searches were conducted using U.S. EPA's Health and Environmental Research Online (HERO) database of scientific literature. HERO searches the following databases: PubMed, TOXLINE¹ (including TSCATS1), and Web of Science. The following resources were searched outside of HERO for health-related values: American Conference of Governmental Industrial Hygienists (ACGIH), Agency for Toxic Substances and Disease Registry (ATSDR), California Environmental Protection Agency (CalEPA), Defense Technical Information Center (DTIC), European Centre for Ecotoxicology and Toxicology of Chemicals (ECETOC), European Chemicals Agency (ECHA), U.S. EPA Chemical Data Access Tool (CDAT), U.S. EPA ChemView, U.S. EPA Integrated Risk Information System (IRIS), U.S. EPA Health Effects Assessment Summary Tables (HEAST), U.S. EPA Office of Water (OW), International Agency for Research on Cancer (IARC), U.S. EPA TSCATS2/TSCATS8e, U.S. EPA High Production Volume (HPV), Chemicals via IPCS INCHEM, Japan Existing Chemical Data Base (JECDB), Organisation for Economic Cooperation and Development (OECD) Screening Information Data Sets (SIDS), OECD International Uniform Chemical Information Database (IUCLID), OECD HPV, National Institute for Occupational Safety and Health (NIOSH), National Toxicology Program (NTP), Occupational Safety and Health Administration (OSHA), and World Health Organization (WHO).

¹Note that this version of TOXLINE is no longer updated

^{(&}lt;u>https://www.nlm.nih.gov/databases/download/toxlinesubset.html</u>); therefore, it was not included in the literature search update from May 2021.

2. REVIEW OF POTENTIALLY RELEVANT DATA (NONCANCER AND CANCER)

As shown in Tables 3A and 3B, there are no potentially relevant short-term, subchronic, chronic, developmental, or reproductive toxicity studies of 1-bromo-2-chloroethane in humans or animals exposed by oral or inhalation routes. The phrase "statistical significance" or the term "significant," used throughout the document, indicates a *p*-value of < 0.05 unless otherwise noted.

| Category | Number of Male/Female, Strain, Species, Study Type, Reported Doses, Study Duration | Dosimetry | Critical Effects | Reference | Notes |
|----------|---|------------------------------------|------------------|-----------|-------|
| Human | | | | | |
| | | 1. Oral (mg/kg-d) | | | |
| ND | | | | | |
| | | 2. Inhalation (mg/m ³) | | | |
| ND | | | | | |
| Animal | | | | | |
| | | 1. Oral (mg/kg-d) | | | |
| ND | | | | | |
| | | 2. Inhalation (mg/m ³) | | | |
| ND | | | | | |
| | | | | | |

| Category | Number of Male/Female, Strain, Species, Study Type, Reported Doses, Study Duration | Dosimetry | Critical Effects | Reference | Notes |
|----------|---|----------------------------|-------------------------|-----------|-------|
| Human | | | | | |
| | 1. Ora | l (mg/kg-d) | | | |
| ND | | | | | |
| | 2. Inhala | ation (mg/m ³) | | | |
| ND | | | | | |
| Animal | | | | | |
| | 1. Ora | l (mg/kg-d) | | | |
| ND | | | | | |
| | 2. Inhala | ation (mg/m ³) | | | |
| ND | | | | | |

ND = no data.

2.1. HUMAN STUDIES

2.1.1. Oral Exposures

No human studies following oral exposure to 1-bromo-2-chloroethane have been identified.

2.1.2. Inhalation Exposures

No human studies following inhalation exposure to 1-bromo-2-chloroethane have been identified.

2.2. ANIMAL STUDIES

2.2.1. Oral Exposures

No animal studies following oral exposure to 1-bromo-2-chloroethane have been identified.

2.2.2. Inhalation Exposures

No animal studies following inhalation exposure to 1-bromo-2-chloroethane have been identified.

2.3. OTHER DATA (SHORT-TERM TESTS, OTHER EXAMINATIONS)

Toxicity data for 1-bromo-2-chloroethane are limited to acute lethality values reported in secondary sources, an acute oral hepatotoxicity study, an acute intraperitoneal (i.p.) injection study, and genotoxicity studies.

2.3.1. Acute Toxicity in Animals

An oral median lethal dose (LD₅₀) value of 64 mg/kg was identified for 1-bromo-2-chloroethane in rats [Frear (1969) as cited in <u>NLM (2016)</u>]; no further data regarding acute oral toxicity were reported. Valade (1957), as cited in <u>U.S. EPA (1985)</u>, reported a 30-minute median lethal concentration (LC₅₀) value of 15,000–25,000 mg/m³ for 1-bromo-2-chloroethane in unspecified laboratory animals. This study also indicated that acute inhalation of "several alkyl halides" caused ataxia in dogs, rats, and guinea pigs; however, it is unclear whether these halides included 1-bromo-2-chloroethane.

<u>Moody et al. (1980)</u> treated male Sprague Dawley rats with 1-bromo-2-chloroethane via gavage and found decreases in cytochrome P450 (CYP450) content of hepatic microsomes to 51% of controls, as well as alterations in relative content of fatty acids within hepatic microsomes 18 hours after exposure to a single dose (0.15 mL/kg). A high correlation between CYP450 loss and decreased arachidonic acid, increased linoleic acid, and increased oleic acid was observed.

Significant mortality was observed in male B6C3F1 mice (5/11 treated animals) at 24 hours following a single i.p. injection of 1.5 mmol/kg (215 mg/kg) of 1-bromo-2-chloroethane. Statistically significant increases in serum sorbitol dehydrogenase (SDH), alanine aminotransferase (ALT), and blood urea nitrogen (BUN) and relative liver and kidney weights were also observed at this dose (Storer and Conolly, 1983). No mortalities or liver or kidney effects were noted at $\leq 1 \text{ mmol/kg}$ (143 mg/kg); no other endpoints were evaluated.

EPA/690/R-21/005F

2.3.2. Genotoxicity

Overview: Studies evaluating the potential genotoxicity of 1-bromo-2-chloroethane are summarized below (see Table 4 for details). Available data indicate that 1-bromo-2-chloroethane and/or its metabolites display genotoxic, mutagenic, clastogenic, and deoxyribonucleic acid (DNA) damaging activity.

| | Table 4. Summary of 1-Bromo-2-Chloroethane (CASRN 107-04-0) Genotoxicity | | | | | | | |
|-----------------|---|--|--|--|--|---|--|--|
| Endpoint | Test System | Doses/ Concentrations Tested | Results without Activation ^a | Results with Activation ^a | Comments | References (comments) | | |
| Genotoxicity st | udies in prokaryotic organi | sms | | | | · | | |
| Mutation | Salmonella typhimurium strains TA1535, TA1537, TA1538, TA98, and TA100 | 0, 1.6, 3.0, 4.7, 6.8, 8.1 μmol/plate | + (TA1535, TA100) - (TA1537, TA1538, TA98) | + (TA1535, TA100) - (TA1537, TA1538, TA98) | A concentration-dependent increase in revertants was observed in TA1535 and TA100 with or without metabolic activation (rat liver S9 fractions). The addition of S9 activation did not significantly increase the mutagenic response. | Barber et al. (1981) | | |
| Mutation | <i>S. typhimurium</i> strains TA1530, TA1535, and TA1538 | 0–15 µmol/plate | + (TA1530, TA1535) - (TA1538) | NT | A concentration-dependent increase in revertants was observed without metabolic activation. | Brem et al. (1974) | | |
| Mutation | <i>S. typhimurium</i> strains TA100 and TA98 | 0, 33, 100, 200, 333, 667, 1,000, 1,500, 2,000 µg/plate | + (TA100, TA98) | + (TA100) - (TA98) | In TA100, the number of revertants was increased >twofold at \geq 667 µg/plate without metabolic activation and \geq 333 µg/plate with metabolic activation. In TA98, the number of revertants was increased >twofold at \geq 333 µg/plate without metabolic activation. | <u>NTP (1990)</u> | | |
| Mutation | <i>S. typhimurium</i> strains TA100 and TA102 | NR | + | (see comments) | Preincubation assay. It is unclear from the abstract whether or not metabolic activation was used. | Hughes et al. (1987) (abstract only) | | |

| | Table 4. Summary of 1-Bromo-2-Chloroethane (CASRN 107-04-0) Genotoxicity | | | | | | | |
|---|---|--|---|---|--|--------------------------------------|--|--|
| Endpoint | Test System | Doses/ Concentrations Tested | Results without Activation ^a | Results with Activation ^a | Comments | References (comments) | | |
| Mutation | <i>S. typhimurium</i> strains TA1535/pK233-2 (empty plasmid) TA1535/DM11 (bacterial dichloromethane dehalogenase) TA1535/GST 5-5 (rat GST) TA1535/GST T1 (human GST) | 0-10.0 μM (DM11); 0-150 μM (GST 5-5, GST T1, and control plasmid pK233-2) | (pK233-2) | + (DM11, GST 5-5, GST T1) | Metabolic activation was achieved using S. typhimurium TA1535 cells expressing mammalian theta-class GSH transferases (rat GST 5-5 or human GST T1) or a bacterial dichloromethane dehalogenase (DM11). A concentration-dependent increase in revertants was observed for each of the enzymes induced; an increase in revertants was not observed in TA1535 with the expression plasmid only. The numbers of revertants/µM per plate for GST 5-5, GST T1, and DM11 were 45, 3.3, and 2,200, respectively. It was <0.1 for the plasmid only. | Wheeler et al. (2001) | | |
| Mutation | <i>S. typhimurium</i> strain TA100 | 0–4.0 mM | + | + | Metabolic activation was tested either with rat GSH added or rat GSH + rat liver cytosol (S100). The number of revertants following GSH + S100 activation was increased almost fourfold over GSH only or without a metabolizing system. | <u>van Bladeren et al.</u> (1981) | | |
| DNA damage (SOS/ <i>umu</i> chromotest) | <i>S. typhimurium</i> strains TA1535/pSK1002 (empty plasmid) NM5004 (rat GST 5-5 and <i>umuC-lac Z</i> operon) | 0–0.5 mM | _ (pSK1002) | + (NM5004) | Strain NM5004 was generated by introducing a plasmid <i>S. typhimurium</i> TA1535 cells containing both rat GSH transferase (GST 5-5) cDNA and the <i>umuC-lac Z</i> operon. <i>Umu</i> gene expression was increased in a dose-dependent manner for NM5004 at noncytotoxic concentrations. Cytotoxicity was observed at \geq 0.25 mM in the NM5004 strain. | <u>Shimada et al.</u> (1996) | | |

| | Table 4. Summary of 1-Bromo-2-Chloroethane (CASRN 107-04-0) Genotoxicity | | | | | | | |
|--|---|---|---|---|---|---|--|--|
| Endpoint | Test System | Doses/ Concentrations Tested | Results without Activation ^a | Results with Activation ^a | Comments | References (comments) | | |
| DNA damage (SOS/ <i>umu</i> chromotest) | <i>S. typhimurium</i> strains TA1535/pSK1002 (empty plasmid) NM5004 (rat GST 5-5 and <i>umuC-lac Z</i> operon) | 0–0.1 mM | _ (pSK1002) | + (NM5004) | Strain NM5004 was generated as describe above by <u>Shimada et al. (1996)</u> . <i>Umu</i> gene expression was increased in a dose-dependent manner for NM5004 at noncytotoxic concentrations. Cytotoxicity was observed at 0.1 mM in the NM5004 strain. | <u>Oda et al. (1996)</u> | | |
| Genotoxicity stu | dies in nonmammalian eu | karyotic organisms | | | | | | |
| Mitotic malsegregation | <i>Aspergillus nidulans</i> diploid strain P1 | 0, 6.0, 12.0, 18.0, 24.0 mM | + | + | The number of malsegregations was significantly elevated at ≥12.0 mM; survival was <50% at ≥18.0 mM. | <u>Crebelli et al. (1995)</u> | | |
| Mitotic recombination (wing-spot test) | Drosophila melanogaster ($flr^3 \times mwh$); 48-h inhalation exposure | 0, 0.31 μg/L | + | + | The tested concentration was based on the calculated LC_{50} . The frequency of wing spots at the LC_{50} was significantly elevated by 17-fold compared with controls. | <u>Chroust et al. (2006)</u> | | |
| Genotoxicity stu | dies in mammalian cells in | n vitro | | | | | | |
| HGPRT mutation (6-thioguanine resistance) | CHO cells | +S9: 0, 0.05, 0.1, 0.2, 0.3, 0.4, 0.5, 0.6, 0.7, 0.8, 1.0 mM; -S9: 0, 0.5, 1, 2, 3, 4, 6, 8 mM | + | + | A concentration-dependent increase in mutants was observed with and without metabolic activation; the mutagenic activity was fourfold greater with metabolic activation (rat liver S9 mix) than without S9. Enhancement by S9 mix did not occur when the NADPH pool was reduced (via omission of NADP from the S9 fraction). Cell survival (relative to untreated control) was reduced by >50% at 6 mM without S9 and 1 mM with S9. | <u>Tan and Hsie (1981)</u> | | |
| CAs | CHL cells | 0.1-0.5 mg/mL (6 h; ±S9) 0.5-3.0 mg/mL (24 h; -S9) 0.5-4.0 mg/mL (48 h); -S9 | + (24 h; 48 h) - (6 h) | + (6 h) | NA | Japan Chemical Industry Ecology-Toxicology Information Center (1996) as cited in <u>NLM (2005)</u> | | |

| | Table 4. Summary of 1-Bromo-2-Chloroethane (CASRN 107-04-0) Genotoxicity | | | | | | | |
|-----------------------|---|------------------------------------|---|---|--|------------------------------|--|--|
| Endpoint | Test System | Doses/ Concentrations Tested | Results without Activation ^a | Results with Activation ^a | Comments | References (comments) | | |
| Genotoxicity stud | Genotoxicity studies in mammals in vivo | | | | | | | |
| (DNA unwinding assay) | Male B6C3F1 mice (six/group); single i.p. injection; sacrifice 4 h after exposure; hepatic cell nuclei were harvested | 0, 0.5, 0.75, 1.0 mmol/kg | + | + | 1-Bromo-2-chloroethane induced 8.9, 15.8, and 24.7% reduction in double-stranded DNA in liver cells harvested from mice exposed to 0.5, 0.75, and 1.0 mM/kg (72, 107, and 143 mg/kg), respectively, compared with control. | Storer and Conolly (1983) | | |

 a + = positive; ± = weakly positive; - = negative.

CA = chromosomal aberration; cDNA = complementary DNA; CHL = Chinese hamster lung; CHO = Chinese hamster ovary (cell line cells); DNA = deoxyribonucleic acid; GSH = glutathione; GST = glutathione-S-transferase; HGPRT = hypoxanthine-guanine phosphoribosyltransferase; i.p. = intraperitoneal; LC₅₀ = median lethal concentration; NA = not applicable; NADP = nicotinamide adenine dinucleotide phosphate; NADPH = reduced form of NADP; NR = not reported, NT = not tested.

1-Bromo-2-chloroethane is mutagenic in bacterial and mammalian cells in vitro, both with and without metabolic activation (Wheeler et al., 2001; NTP, 1990; Hughes et al., 1987; Barber et al., 1981; Tan and Hsie, 1981; van Bladeren et al., 1981; Brem et al., 1974). Addition of glutathione-*S*-transferases (GST) or glutathione (GSH) with or without S100 into *Salmonella typhimurium* bacterial test systems enhanced mutagenicity, indicating that GSH conjugation results in the formation of mutagenic metabolites (Wheeler et al., 2001; van Bladeren et al., 1981). In the Chinese hamster ovary cell/hypoxanthine-guanine phosphoribosyl transferase (CHO/HGPRT) system, mutagenicity was enhanced in the presence of an S9 metabolic activation fraction; this enhancement did not occur when nicotinamide adenine dinucleotide phosphate (NADP) was omitted from the S9 fraction, suggesting that CYP450 enzymes could be involved in the increased metabolic activation (Tan and Hsie, 1981).

Chromosomal damage and mitotic malsegregation/recombination have been induced by 1-bromo-2-chloroethane in mammalian cells and nonmammalian eukaryotic organisms. 1-bromo-2-chloroethane induced chromosomal aberrations (CAs) in Chinese hamster lung (CHL) cells, both in the presence and absence of a metabolic activation system, in a study by the Japan Chemical Industry Ecology-Toxicology Information Center, available only as a brief summary in the Chemical Carcinogenesis Research Information System (CCRIS) (NLM, 2005). 1-Bromo-2-chloroethane also induced mitotic malsegregation in *Aspergillus nidulans* (Crebelli et al., 1995) and mitotic recombination in *Drosophila melanogaster* (Chroust et al., 2006).

DNA damage has been reported in bacterial cells exposed to 1-bromo-2-chloroethane in the presence of GSTs (Oda et al., 1996; Shimada et al., 1996). DNA damage was assessed in the SOS/*umu* test system, which measures the induction of DNA damage responsive *umuC* gene expression by cellular β -galactosidase activity produced by an *umuC-lac Z* fusion gene. Shimada et al. (1996) constructed a strain of *S. typhimurium* for use in the SOS/*umu* test system that possessed enhanced GST activity by introducing a rat GST 5-5 cDNA plasmid into *S. typhimurium* TA1535 (resulting in strain NM5004). Expression of the *umuC* gene was not increased in the tester TA1535 strain containing only the empty plasmid (pSK1002), but it was increased in a dose-related manner in the NM5004 strain expressing rat GST. These data suggest that GSH conjugates of 1-bromo-2-chloroethane were DNA-reactive intermediates.

DNA damage was also observed in liver cells from male B6C3F1 mice following a single i.p. dose of 1-bromo-2-chloroethane at greater than or equal to 0.5 mmol/kg (72 mg/kg) (Storer and Conolly, 1983). Liver cell nuclei harvested from exposed mice showed a dose-related decrease in the percentage of double-stranded DNA observed (indicating a dose-related increase in DNA damage), as assessed by an alkaline DNA unwinding/hydroxylapatite batch chromatography method.

In summary, 1-bromo-2-chloroethane displayed mutagenic activity among in vitro models employing bacterial and mammalian cells, and this mutagenic activity was increased concomitantly with an increase in GSH conjugation (Wheeler et al., 2001; NTP, 1990; Hughes et al., 1987; Tan and Hsie, 1981; van Bladeren et al., 1981; Brem et al., 1974). Increased CYP450 activity (as suggested by manipulating NADP levels within the S9 fraction) promoted elevated mutagenicity in a CHO/HGPRT model system, indicating a potential role for CYP450 in the mutagenicity resulting from 1-bromo-2-chloroethane metabolism (Tan and Hsie, 1981). Chromosomal aberrations and mitotic malsegregation were increased in both nonmammalian eukaryotic organisms and mammalian cell model systems (Chroust et al., 2006; NLM, 2005; Crebelli et al., 1995). Lastly, the induction of DNA damage in bacterial systems, as well as in in

vivo liver cells of male B6C3F1 mice, occurred in a dose-dependent manner after 1-bromo-2-chloroethane exposure (Oda et al., 1996; Shimada et al., 1996; Storer and Conolly, 1983).

2.3.3. Absorption, Distribution, Metabolism, and Excretion Studies

Available in vivo and in vitro data indicate that 1-bromo-2-chloroethane is metabolized by a GSH conjugation pathway (Jean and Reed, 1992; Marchand and Reed, 1989). GSH conjugation was demonstrated in vivo by the detection of *S*-(2-chloroethyl)glutathione (CEG) and its hydrolysis product, *S*-(2-hydroxyethyl)glutathione (HEG), in the bile of bile-duct-cannulated rats following intravenous (i.v.) injection with 75 mg/kg 1-bromo-2-chloroethane (Marchand and Reed, 1989). The total amount of CEG detected in bile was 2% of the administered dose. Marchand and Reed (1989) suggested that the low levels of CEG detected in bile may result from rapid conversion of CEG to other metabolites (e.g., HEG) or intrahepatic transport of CEG into plasma rather than bile. Data from the structurally similar dihaloalkanes 1,2-dichloroethane and 1,2-dibromoethane indicate that once formed, CEG will cyclize to form an electrophilic episulfonium ion in vivo, which subsequently binds to DNA [reviewed by Guengerich (1994) and Dekant and Vamvakas (1993)].

In isolated rat hepatocytes, 1-bromo-2-chloroethane was metabolized to HEG, S-(carboxymethyl)glutathione (CMG), and S,S'-(1,2-ethanediyl)bis(glutathione) (EDG) conjugates. HEG was produced in the largest amounts, followed by EDG and CMG (Jean and <u>Reed, 1992</u>). HEG and EDG are products of direct GSH conjugation (EDG results from conjugation at both carbon atoms), whereas CMG is formed after an initial oxidation to 2-haloacetaldehyde metabolites (Jean and Reed, 1992). Formation of GSH conjugates was concomitant with intracellular GSH depletion, measured as an 84% decrease in intracellular GSH levels in response to 1-bromo-2-chloroethane treatment. The addition of extracellular GSH into the incubation medium increased the formation of EDG conjugates by 179% (Jean and <u>Reed, 1992</u>). Combined, these data suggest that GSH conjugation is a primary and likely rate-limiting step in the metabolic activation of 1-bromo-2-chloroethane.

No data are available on the absorption, distribution, or excretion of 1-bromo-2-chloroethane.

3. DERIVATION OF PROVISIONAL VALUES

3.1. DERIVATION OF PROVISIONAL REFERENCE DOSES

No studies have been identified regarding toxicity of 1-bromo-2-chloroethane to humans by oral exposure. Animal studies of oral exposure to 1-bromo-2-chloroethane are limited to acute studies of inadequate design, duration, and scope to support derivation of a subchronic or chronic provisional reference dose (p-RfD). As a result of the limitations of the available oral toxicity data for 1-bromo-2-chloroethane, subchronic and chronic p-RfDs are not derived directly. Instead, screening subchronic and chronic p-RfDs are derived in Appendix A using an alternative analogue approach. Based on the overall analogue approach presented in Appendix A, 1,2-dibromo-3-chloropropane is selected as the most appropriate analogue for 1-bromo-2-chloroethane for deriving a screening subchronic and chronic p-RfD (see Table 5).

3.2. DERIVATION OF PROVISIONAL REFERENCE CONCENTRATIONS

No studies have been identified on the toxicity of 1-bromo-2-chloroethane to humans by inhalation exposure. Animal studies of inhalation exposure to 1-bromo-2-chloroethane are limited to a single acute lethality study of inadequate design, duration, and scope to support derivation of a subchronic or chronic provisional reference concentrations (p-RfCs). Due to lack of adequate inhalation toxicity data for 1-bromo-2-chloroethane, subchronic and chronic p-RfCs are not derived directly. Instead, screening subchronic and chronic p-RfCs are derived in Appendix A using an alternative analogue approach. Based on the overall analogue approach presented in Appendix A, 1,2-dibromo-3-chloropropane is selected as the most appropriate analogue for 1-bromo-2-chloroethane for deriving a screening subchronic and chronic p-RfC (see Table 5).

3.3. SUMMARY OF NONCANCER PROVISIONAL REFERENCE VALUES

The noncancer screening provisional reference values for 1-bromo-2-chloroethane are summarized in Table 5.

| | Table 5. Summary of Noncancer Reference Values for 1-Bromo-2-Chloroethane (CASRN 107-04-0) | | | | | | | | |
|---|---|--------------------------------|----------------------|----------------|------------------------------------|-------|--|--|--|
| Toxicity Type (units) | Species/ Sex | Critical Effect | p-Reference Value | POD Method | POD | UFc | Principal Study | | |
| Screening subchronic p-RfD (mg/kg-d) | Rabbit/M | Reduced germ cell number | 1×10^{-3} | NOAEL (HED) | 0.3 (based on analogue POD) | 300 | Foote et al. (1986a, b) as cited in <u>U.S. EPA</u> (2006) | | |
| Screening chronic p-RfD (mg/kg-d) | Rabbit/M | Reduced germ cell number | 1×10^{-4} | NOAEL (HED) | 0.3 (based on analogue POD) | 3,000 | Foote et al. (1986a, b) as cited in <u>U.S. EPA</u> (2006) | | |
| Screening subchronic p-RfC (mg/m ³) | Rabbit/M | Testicular toxicity | 6×10^{-4} | NOAEL (HEC) | 0.17 (based on analogue POD) | 300 | Rao et al. (1982) as cited in <u>U.S. EPA</u> (2006) | | |
| Screening chronic p-RfC (mg/m ³) | Rabbit/M | Testicular toxicity | 6 × 10 ⁻⁵ | NOAEL (HEC) | 0.17 (based on analogue POD) | 3,000 | Rao et al. (1982) as cited in <u>U.S. EPA</u> (2006) | | |

HEC = human equivalent concentration; HED = human equivalent dose; M = male(s);

NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfC = provisional reference concentration; p-RfD = provisional reference dose; UF_C = composite uncertainty factor.

3.4. CANCER WEIGHT-OF-EVIDENCE DESCRIPTOR

Although the scientific literature provides information on the mutagenicity and genotoxicity of 1-bromo-2-chloroethane, no studies have been conducted to directly assess its carcinogenicity. Under the U.S. EPA Cancer Guidelines (U.S. EPA, 2005), there is "Inadequate Information to Assess the Carcinogenic Potential" of 1-bromo-2-chloroethane (see Table 6).

Within the current U.S. EPA Cancer Guidelines (U.S. EPA, 2005), there is no standard methodology to support the identification of a weight-of-evidence (WOE) descriptor and derivation of provisional cancer risk estimates for data-poor chemicals using an analogue approach. In the absence of an established framework, a screening evaluation of potential carcinogenicity is provided using the methodology described in Appendix B. This evaluation determined that there was a qualitative level of *concern for potential carcinogenicity* for 1-bromo-2-chloroethane (see Appendix C).

| Table 6. Cancer W | OE Descrip | tor for 1-Bromo-2- | Chloroethane (CASRN 107-04-0) |
|---|-------------|---|---|
| Possible WOE Descriptor | Designation | Route of Entry (oral, inhalation, or both) | Comments |
| "Carcinogenic to Humans" | NS | NA | No human data are available. |
| <i>"Likely to Be Carcinogenic to Humans"</i> | NS | NA | No adequate chronic-duration animal cancer bioassays are available. |
| "Suggestive Evidence of Carcinogenic Potential" | NS | NA | No adequate chronic-duration animal cancer bioassays are available. |
| "Inadequate Information to Assess Carcinogenic Potential" | Selected | Both | No studies are available assessing the carcinogenic potential of 1-bromo-2-chloroethane in humans or animals. |
| "Not Likely to Be Carcinogenic to Humans" | NS | NA | No evidence of noncarcinogenicity is available. No adequate chronic-duration animal cancer bioassays are available. |

NA = not applicable; NS = not selected; WOE = weight of evidence.

3.5. DERIVATION OF PROVISIONAL CANCER RISK ESTIMATES

The absence of suitable data precludes development of cancer risk estimates for 1-bromo-2-chloroethane (see Table 7).

| Table 7. Summary of Cancer Risk Estimates for1-Bromo-2-Chloroethane (CASRN 107-04-0) | | | | |
|--|-------------|------------|-------------------------|-----------------|
| Toxicity Type (units) | Species/Sex | Tumor Type | Cancer Risk Estimate | Principal Study |
| Screening p-OSF (mg/kg-d) ⁻¹ | NDr | | | |
| Screening p-IUR (µg/m ³) ⁻¹ | NDr | | | |

NDr = not determined; p-IUR = provisional inhalation unit risk; p-OSF = provisional oral slope factor.

APPENDIX A. SCREENING NONCANCER PROVISIONAL VALUES

Due to the lack of evidence described in the main Provisional Peer-Reviewed Toxicity Value (PPRTV) document, it is inappropriate to derive provisional toxicity values for 1-bromo-2-chloroethane. However, some information is available for this chemical, which although insufficient to support derivation of a provisional toxicity value under current guidelines, may be of limited use to risk assessors. In such cases, the Center for Public Health and Environmental Assessment (CPHEA) summarizes available information in an appendix and develops a "screening value." Appendices receive the same level of internal and external scientific peer review as the provisional reference values to ensure their appropriateness within the limitations detailed in the document. Users of screening toxicity values in an appendix to a PPRTV assessment should understand that there could be more uncertainty associated with deriving an appendix screening toxicity value than for a value presented in the body of the assessment. Questions or concerns about the appropriate use of screening values should be directed to the CPHEA.

APPLICATION OF AN ALTERNATIVE ANALOGUE APPROACH

The analogue approach allows for the use of data from related compounds to calculate screening values when data for the compound of interest are limited or unavailable. Details regarding searches and methods for analogue analysis are presented in <u>Wang et al. (2012)</u>. Three types of potential analogues (structural, metabolic, and toxicity-like) are identified to facilitate the final analogue chemical selection. The analogue approach may or may not be route-specific or applicable to multiple routes of exposure. All information was considered together as part of the final weight-of-evidence (WOE) approach to select the most suitable analogue both toxicologically and chemically.

Structural Analogues

An initial analogue search focused on the identification of structurally similar chemicals with toxicity values from the Integrated Risk Information System (IRIS), PPRTVs, Agency for Toxic Substances and Disease Registry (ATSDR), or California Environmental Protection Agency (CalEPA) databases to leverage the body of well-characterized chemical-class information. As described in Wang et al. (2012), structural similarity for analogues is evaluated using the National Library of Medicine's (NLM's) ChemIDplus database (ChemIDplus, 2021) and the Organisation for Economic Co-operation and Development (OECD) Toolbox. Both structural similarity approaches were used to calculate structural similarity using the Tanimoto method. Three structural analogues to 1-bromo-2-chloroethane that have oral and/or inhalation toxicity values were identified: 1,2-dibromoethane (U.S. EPA, 2004), 1,2-dichloroethane (U.S. EPA, 2010), and 1,2-dibromo-3-chloropropane (U.S. EPA, 2006, 2003). Table A-1 summarizes the analogues' physicochemical properties and similarity scores. ChemIDplus similarity scores were available only for 1,2-dibromoethane (92%) and 1,2-dichloroethane (68%). OECD Quantitative Structure-Activity Relationship (QSAR) pairwise toolbox similarity scores were similar for all three candidate analogues (24-33%) and the target chemical. The target compound and all candidate analogues are expected to be bioavailable by oral and inhalation routes (based on water solubility and Kow values). A range of vapor pressure values (0.580-78.9 mm Hg at 25°C) is identified for 1-bromo-2-chloroethane and relevant analogue chemicals, suggesting that these compounds may exhibit differences in volatility. All of the candidate analogues are considered appropriate structural analogues based on the approach described in Wang et al. (2012).

| and Candidate Analogues ^a | | | | | |
|--|----------------------------|-------------------------|-----------------------|--|--|
| Property | 1-Bromo- 2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo- 3-Chloropropane (2R and 2S Isomers) | |
| Structure | CI Br | Br | CI CI | Cl Br | |
| CASRN | 107-04-0 | 106-93-4 | 107-06-2 | 96-12-8 | |
| Molecular weight | 143 | 188 | 99 | 236 | |
| OECD QSAR Toolbox similarity score (%) ^b | 100 | 33 | 33 | 24 | |
| ChemIDplus similarity score (%) ^c | 100 | 92 | 68 | NV | |
| Melting point (°C) | -16.3 | 9.8 | -35.3 | 6.04 | |
| Boiling point (°C) | 107 | 132 | 83.2 | 196 | |
| Vapor pressure (mm Hg at 25°C) | 33.1 | 11.2 | 78.9 | 0.580 | |
| Henry's law constant (atm-m ³ /mole at 25°C) | 9.09×10^{-4} | 6.50×10^{-4} | 1.18×10^{-3} | 2.93×10^{-4} (predicted) | |
| Water solubility (mol/L) | 4.79×10^{-2} | 2.11 × 10 ⁻² | 8.68×10^{-2} | 4.71×10^{-3} | |
| Octanol-water partition coefficient (log K _{ow}) | 1.86 (predicted) | 1.96 | 1.48 | 2.96 | |

Table A-1. Physicochemical Properties of 1-Bromo-2-Chloroethane (CASRN 107-04-0)and Candidate Analogues^a

^aData represent average values as reported on the U.S. EPA's CompTox Chemicals Dashboard unless otherwise specified (<u>https://comptox.epa.gov/dashboard/dsstoxdb/results?search=DTXSID4024775#properties</u>). Data accessed June 2021.

^b<u>OECD (2018)</u>.

^cChemIDplus advanced similarity scores (<u>ChemIDplus, 2021</u>).

NV = not available; OECD =Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship.

Metabolic Analogues

Table A-2 summarizes available toxicokinetic data for 1-bromo-2-chloroethane and the structurally similar compounds identified as candidate analogues.

Absorption, Distribution, and Excretion

No data are available to describe the absorption, distribution, or excretion of 1-bromo-2-chloroethane. Oral absorption is rapid and extensive for all three candidate analogue compounds (U.S. EPA, 2004; ATSDR, 2001; Gingell et al., 1987). Absorption is also rapid and extensive following inhalation exposure to 1,2-dibromoethane and 1,2-dichloroethane (U.S. EPA, 2004; ATSDR, 2001). There are no data available on the rate or extent of absorption following inhalation exposure to 1,2-dibromo-3-chloropropane. Widespread distribution to multiple organs and excretion primarily in the urine has been demonstrated for all three candidate analogue compounds (U.S. EPA, 2004; ATSDR, 2004; ATSDR, 2004; ATSDR, 2001).

| Parameter | 1-Bromo-2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo-3-Chloropropane (2R and 2S isomers) |
|--|------------------------|--|--|---|
| Structure | CI | Br | CICI | Cl Br |
| CASRN | 107-04-0 | 106-93-4 | 107-06-2 | 96-12-8 |
| Absorption | | | | |
| Rate and extent of oral absorption | ND | Rat: rapid and extensive (C_{max} within 30 min; >80% of dose absorbed) | Rat: rapid and extensive (C_{max} within 15 min; >80% of dose absorbed) | Rat: rapid and extensive (C_{max} within 5–40 min; >80% of dose absorbed) |
| Rate and extent of inhalation absorption | ND | Rat: rapid and extensive (C_{max} within 20 min; 40 to <60% of dose absorbed) | Rat: rapid and extensive (C_{max} within $1-3$ h; >80% of dose absorbed) | ND |
| Human blood-gas partition coefficient ^a | 29.2 | ND | 19.5 | ND |
| Rat blood-gas partition coefficient ^a | 52.7 | 119 | 30.4 | ND |
| Distribution | | | · | · |
| Extent of distribution | ND | Rat and mouse: wide distribution to multiple organs; covalent binding to tissues (kidney, stomach, liver, and testis after oral exposure; nasal mucosa, liver, lung, kidney, and small intestine after i.p. injection) | Rat: wide distribution to multiple organs; transplacental distribution was demonstrated | Rat: wide distribution to multiple organs; $V_d = 4.98 \text{ L/kg}$ (oral) |

| Parameter | 1-Bromo-2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo-3-Chloropropane (2R and 2S isomers) |
|---------------------------------|--|---|--|---|
| Metabolism—oxidati | on by CYP450 | | | |
| Primary reactive metabolites | By analogy to 1,2-dibromoethane and 1,2-dichloroethane: (1) 2-bromo-acetaldehyde; (2) 2-chloro-acetaldehyde X = Br, Cl By analogy to 1,2-dibromoethane and 1,2-dichloroethane: (3) HBr, HCl | Demonstrated: (1) 2-bromo-acetaldehyde | Demonstrated: (1) 2-chloro-acetaldehyde | Demonstrated: (1) 2-bromoacrolein Br + H (2) 2-chloro-3-(bromomethyl) oxirane Br + Cl (3) 1-bromo-3-chloroacetone Cl + Br Demonstrated: (4) HBr, HCl |

| Parameter | 1-Bromo-2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo-3-Chloropropane (2R and 2S isomers) |
|---------------------------------|--|---|---|--|
| Potential role in toxicity | By analogy to 1,2-dibromoethane and 1,2-dichloroethane: (1) Forms DNA adducts (2) Lipid peroxidation (3) Protein binding | Demonstrated: (1) Forms DNA adducts (2) Lipid peroxidation (3) Protein binding | Demonstrated: (1) Forms DNA adducts (2) Lipid peroxidation (3) Protein binding | Demonstrated: (1) 2-bromoacrolein and 2-chloro-3-(bromomethyl) oxirane are direct-acting mutagens in <i>Salmonella</i> (2) 2-bromoacrolein forms DNA adducts |
| Metabolism—GSH con | jugation ^b | | | |
| Primary reactive metabolites | Inferred from CEG as primary GSH conjugate: (1) HBr released during GSH conjugation (HCl released in minor amounts) | Inferred from BEG as GSH conjugate: (1) HBr released during GSH conjugation Demonstrated: | Inferred from CEG as GSH conjugate: (1) HCl released during GSH conjugation Demonstrated: | Inferred from CBPG as primary GSH conjugate: (1) HBr released during GSH conjugation (HCl released in minor amounts) |
| | By analogy to 1,2-dichloroethane: (2) Reactive episulfonium ion formed following cyclization of GSH adduct HO = HO = HO = HO | (2) Reactive episulfonium ion formed following cyclization of GSH adduct HO = HO = | (2) Reactive episulfonium ion formed following cyclization of GSH adduct HO = HO = | Demonstrated: (2) Reactive episulfonium ion formed following cyclization of GSH adduct $Br^{-} \int_{N+a}^{C} H_{-} H_{-}$ |

| Parameter | 1-Bromo-2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo-3-Chloropropane (2R and 2S isomers) |
|----------------------------|--|---|---|---|
| Potential role in toxicity | By analogy to all potential analogues: (1) Forms DNA adducts By analogy to 1,2-dibromoethane and 1,2-dibromo-3-chloropropane: (2) Causes DNA damage in cells including spermatocytes Demonstrated: (3) Mutagenicity among genotoxicity assays mediated by GSH activity | Demonstrated: (1) Forms DNA adducts, (2) Causes DNA damage in hepatocytes and spermatocytes (unscheduled DNA synthesis) | Demonstrated: (1) Forms DNA adducts, (2) Binds to protein and DNA in kidney and liver | Demonstrated: (1) Forms DNA adducts, (2) Causes DNA damage in multiple tissues including kidneys and testes (unscheduled DNA synthesis), (3) DNA strand breaks in spermatocytes, (4) DNA damage is correlated with tissue injury at higher doses |
| Excretion | | | | |
| Half-life in blood (h) | ND | ND | Rat, oral: <1 | Rat, oral: 2.64 |
| Route of excretion | ND | Rat, oral: 60 to <80% urine <20% feces <20% exhaled air (form not indicated) | Rat, oral: 60 to <80% urine <20% feces 20 to <40% excreted unchanged in exhaled air Rat, inhalation: >80% urine <20% feces <20% as CO ₂ in exhaled air | Rat, oral: 40 to <60% urine <20% feces 20 to <40% exhaled air (<1% unmetabolized; 20% as CO ₂) |

| Table A-2. Co | Table A-2. Comparison of Available ADME Data for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Candidate Analogues | | | | |
|---------------|---|--|---|---|--|
| Parameter | 1-Bromo-2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo-3-Chloropropane (2R and 2S isomers) | |
| References | Guengerich (1994); Dekant and Vamvakas (1993); Jean and Reed (1992); Gargas et al. (1989); Marchand and Reed (1989); Guengerich et al. (1987); van Bladeren et al. (1981); Wheeler et al. (2001); Shimada et al. (1996) | <u>U.S. EPA (2004); Gargas et al.</u> (1989); Guengerich et al. (1987); NTP (2021) | <u>ATSDR (2001); Gargas et al.</u> (1989); <u>Guengerich et al. (1987)</u> | Guengerich (1994); van Beerendonk et al. (1994); Dekant and Vamvakas (1993); Humphreys et al. (1991); Pearson et al. (1990); Dohn et al. (1988); Omichinski et al. (1988a); Omichinski et al. (1988b); Gingell et al. (1987); Guengerich et al. (1987) | |

^aValues represent means in the case of multiple studies.

Г

^bMetabolism information was categorized as: (1) experimentally **demonstrated**, (2) **inferred from** other experimental data for the same compound, or (3) **by analogy** to experimental data for a different compound.

ADME = absorption, distribution, metabolism, and excretion; BEG = S-(2-bromoethyl)glutathione; Br = bromine; CBPG = S-(3-chloro-2-bromopropyl) glutathione; CEG = S-(2-chloroethyl)glutathione; C_{max} = maximum concentration; Cl = chlorine; CO₂ = carbon dioxide; CYP450 = cytochrome P450; DNA = deoxyribonucleic acid; GSH = glutathione; HBr = hydrogen bromide; HCl = hydrogen chloride; i.p. = intraperitoneal; ND = no data; V_d = volume of distribution.

Metabolism

Each of the candidate analogue compounds is metabolized by a cytochrome P450 (CYP450) oxidation pathway and a direct glutathione (GSH) conjugation pathway. Contributions from both pathways are presumed by analogy for the target compound as well; however, experimental data for 1-bromo-2-chloroethane (the target compound) are available only for the GSH conjugation pathway. Formation of a reactive episulfonium ion occurs, or is expected to occur, for all compounds. Although metabolic pathways are similar among the target chemical and potential analogues, available data indicate that the rate of GSH conjugation and the specific products of oxidative metabolism differ between compounds.

Studies suggest that GSH conjugation leads to the formation of reactive intermediates for the target compound and potential analogues. As discussed in the "Absorption, Distribution, Metabolism, and Excretion" (ADME) section in the main document, 1-bromo-2-chloroethane was metabolized to S-(2-chloroethyl)glutathione (CEG) in bile-duct-cannulated rats (Marchand and Reed, 1989). GSH conjugation to CEG is also suggested by formation of S-(2-hydroxyethyl)glutathione (HEG), a hydrolysis product of CEG, in isolated rat hepatocytes (Jean and Reed, 1992). Metabolic studies with 1,2-dichloroethane (which also forms CEG by GSH conjugation) report that once formed, CEG cyclizes to form an electrophilic episulfonium ion in vivo, which subsequently binds to deoxyribonucleic acid (DNA) [reviewed by Guengerich (1994); Dekant and Vamvakas (1993)]. Similar electrophilic metabolites (i.e., episulfonium ions) were also produced following GSH conjugation to 1,2-dibromoethane to form S-(2-bromoethyl)glutathione (BEG) and to 1,2-dibromo-3-chloropropane to form S-(3-chloro-2-bromopropyl) glutathione (CBPG), also resulting in DNA and protein binding and subsequent damage to DNA (NTP, 2021; U.S. EPA, 2004; ATSDR, 2001; Guengerich, 1994; van Beerendonk et al., 1994; Dekant and Vamvakas, 1993; Pearson et al., 1990; Dohn et al., 1988; Omichinski et al., 1988b; Omichinski et al., 1988a; Guengerich et al., 1987). The rate of GSH conjugation is anticipated to be faster for the brominated compounds compared with 1,2-dichloroethane because of the increased relative reactivity of the bromine versus the chlorine leaving group. This is demonstrated by the initial GSH conjugation products identified for both 1-bromo-2-chloroethane and the proposed analogue 1,2-dibromo-3-chloropropane. When both chlorine and bromine substituents are present, GSH conjugation occurs preferentially at the brominated site (yielding CEG and CBPG, respectively) (Dekant and Vamvakas, 1993; Jean and Reed, 1992; Humphreys et al., 1991; Marchand and Reed, 1989). The concomitant release of hydrogen bromide (HBr) (with minor release of hydrogen chloride [HCl]) is inferred (i.e., not experimentally demonstrated) by the preference for GSH conjugation at the brominated site. Halogen reactivity (i.e., bromine vs. chlorine) is expected to have less influence on the rate of the cyclization reaction that occurs subsequent to GSH conjugation, because the reacting centers are held in close proximity to each other, resulting in rapid formation of the episulfonium ion from the GSH adduct in all cases.

Oxidative metabolites differ among the candidate analogue compounds. CYP450-mediated oxidation of 1,2-dibromoethane and 1,2-dichloroethane results in the production of haloacetaldehyde metabolites that can cause lipid peroxidation and bind to cellular proteins and DNA (<u>NTP, 2021</u>; <u>U.S. EPA, 2004</u>; <u>ATSDR, 2001</u>). HBr and HCl are released during the formation of the haloacetaldehyde metabolite (<u>NTP, 1991</u>; <u>van Bladeren et al., 1981</u>). Although not experimentally demonstrated, this pathway is expected for the target compound as well, based on analogy to the candidate analogue compounds. Oxidative metabolism of 1,2-dibromo-3-chloroethane produces three primary metabolites (2-bromoacrolein, 2-chloro-3-[bromomethyl] oxirane, 1-bromo-3-chloroacetone) (<u>Pearson et al., 1990</u>; <u>Dohn et al.,</u> <u>1988</u>; <u>Omichinski et al., 1988b</u>), none of which can be produced by oxidation of the target compound due to structural constraints. Both HBr and HCl are released during the formation of 2-bromoacrolein, while only HBr is released during the formation of 2-chloro-3-(bromomethyl) oxirane and 1-bromo-3-chloroacetone (<u>Omichinski et al., 1988b</u>). 2-Bromoacrolein and 2-chloro-3-(bromomethyl) oxirane are electrophilic and have been shown to be clastogenic and direct-acting mutagens in *Salmonella* (van Beerendonk et al., 1994; Pearson et al., 1990). 2-Bromoacrolein has also been demonstrated to bind to DNA, and the resulting adducts have been characterized (van Beerendonk et al., 1994</u>).

Because of the preference for relative reactivity of the bromine versus the chlorine leaving group during GSH conjugation, 1,2-dibromoethane and 1,2-dibromo-3-chloropropane are favored as metabolic analogues over 1,2-dichloroethane. Assuming that both CYP450 oxidation and GSH conjugation are equally relevant to the toxicity of the target compound, 1,2-dibromoethane is the most appropriate choice as a metabolic analogue. If the GSH conjugation pathway predominates in its relevance to toxicity, 1,2-dibromo-3-chloropropane may be a more appropriate metabolic analogue.

Toxicity-Like Analogues—Oral

Table A-3 summarizes available oral toxicity values for 1-bromo-2-chloroethane and the compounds identified as potential structural analogues.

| | - | able Oral Toxicity l 7-04-0) and Candida | | |
|--------------------------------|----------------------------|---|--|---|
| Parameter | 1-Bromo- 2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo- 3-Chloropropane |
| Structure | CI Br | Br | CI CI | CI Br |
| CASRN | 107-04-0 | 106-93-4 | 107-06-2 | 96-12-8 |
| Repeated-dose toxic | ity—subchronic | | · | · |
| POD (mg/kg-d) | NA | NA | 58 | 0.7 |
| POD type | NA | NA | LOAEL | NOAEL |
| Subchronic UF _C | NA | NA | 3,000 (UF _A , UF _D , UF _H , UF _L) | $300 (UF_A, UF_D, UF_H)$ |
| Subchronic p-RfD (mg/kg-d) | NA | NA | 2×10^{-2} | 2×10^{-3} |
| Critical effects ^a | NA | NA | Increased absolute kidney weight in female rats | Male reproductive toxicity (reduced germ cell number) |
| Species | NA | NA | Rat | Rabbit |
| Duration | NA | NA | 13 wk | 10 wk |
| Route (method) | NA | NA | Drinking water | Drinking water |
| Source | NA | NA | <u>U.S. EPA (2010)</u> | <u>U.S. EPA (2006)</u> |
| Repeated-dose toxic | ity—chronic | | · | · |
| POD (mg/kg-d) | NA | 27 | 58 | 0.7 |
| POD type | NA | LOAEL (ADJ) | LOAEL | NOAEL |
| Chronic UF _C | NA | 3,000 (UF _A , UF _D , UF _H , UF _L) | 10,000 (UF _A , UF _D , UF _H , UF _L , UF _S) | 3,000 (UF _A , UF _D , UF _H , UF _S) |
| Chronic RfD/p-RfD (mg/kg-d) | NA | 9×10^{-3} | 6×10^{-3} (screening) | 2×10^{-4} |
| Critical effects ^a | NA | Liver peliosis, testicular atrophy, and adrenal cortical degeneration | Increased absolute kidney weight in female rats | Testicular toxicity (reduced germ cell number) |
| Species | NA | Rat (M and F) | Rat | Rabbit |
| Duration | NA | 49 wk (M); 61 wk (F) | 13 wk | 10 wk |
| Route (method) | NA | Gavage in corn oil (5 d/wk) | Drinking water | Drinking water |
| Source | NA | U.S. EPA (2004) | U.S. EPA (2010) | U.S. EPA (2006) |

| Table A-3. Comparison of Available Oral Toxicity Data for 1-Bromo-2-Chloroethane(CASRN 107-04-0) and Candidate Analogues | | | | | | | |
|--|--|--|-------------------------|-------------------------|--|--|--|
| 1-Bromo- 2-Chloroethane1,2-Dibromoethane1,2-Dibromo- 3-Chloropropane | | | | | | | |
| Acute oral lethality d | Acute oral lethality data | | | | | | |
| Rat oral LD50 (mg/kg) | 170 | | | | | | |
| Toxicity at rat LD ₅₀ NR | | NR | NR | NR | | | |
| Source | <u>U.S. EPA (2020b);</u> <u>U.S. EPA (1985)</u> | <u>U.S. EPA (2020b);</u> <u>U.S. EPA (2004)</u> | <u>U.S. EPA (2020a)</u> | <u>U.S. EPA (2020b)</u> | | | |

^aExposure-response arrays were prepared to illustrate the dose-response relationship for testicular, kidney, and liver effects across the candidate analogue compounds (Figures A-1, A-2, and A-3, respectively).

 $ADJ = duration adjusted; F = female(s); LD_{50} = median lethal dose; LOAEL = lowest-observed-adverse-effect level; M = male(s); NA = not applicable; NOAEL = no-observed-adverse-effect level; NR = not reported; POD = point of departure; p-RfD = provisional reference dose; RfD = oral reference dose; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_L = LOAEL-to-NOAEL uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.$

No repeated-dose oral toxicity data are available for 1-bromo-2-chloroethane. Acute toxicity data are limited to a reported median lethal dose (LD₅₀) of 64 mg/kg-day in rats; no further information was provided [Frear (1969) as cited in <u>NLM (2016)</u>].

Oral toxicity values are available for 1,2-dibromoethane, 1,2-dichloroethane, and 1,2-dibromo-3-chloropropane (see Table A-3), and the data supporting these values are extensive [as cited in <u>U.S. EPA (2010, 2006, 2004)</u>]. The target organs of toxicity for these potential analogues include the testis, kidney, and liver (see exposure-response arrays in Figures A-1, A-2, and A-3). After review of the available data for candidate analogues, testicular effects are identified as the most sensitive toxicity target after exposure to 1,2-dibromo-3-chloropropane and 1,2-dibromoethane.

Testicular effects occurred at the lowest oral doses for 1,2-dibromoethane and 1,2-dibromo-3-chloropropane and are identified as being among the critical effects for derivation of provisional reference dose (p-RfD) values for both compounds (see Table A-3, Figure A-1). In support of this endpoint, epidemiological studies of 1,2-dibromo-3-chloropropane-exposed production workers, farmers and pesticide applicators have demonstrated impaired testicular function (decreased spermatogenesis and sperm count and altered sperm morphology) in exposed humans. The testicular toxicity of 1,2-dibromo-3-chloropropane is known to be mediated exclusively by the GSH conjugation pathway because the inhibition of this pathway completely abrogated the observed testicular effects (Omichinski et al., 1988a; Søderlund et al., 1988). Reactive episulfonium metabolites are well known to bind to DNA and promote increased rates of DNA damage. Within the context of the 1,2-dibromo-3-chloropropane database, reactive episulfonium metabolites were observed to produce increased amounts of testicular DNA damage, promote testicular necrosis and atrophy at higher doses (Søderlund et al., 1988), and in turn, impair spermatogenesis and prolong oligospermia (Meistrich et al., 2003). This same sequence of events is likely to occur for 1,2-dibromoethane and 1-bromo-2-chloroethane, which are anticipated to undergo GSH conjugation at similar rates to 1,2-dibromo-3-chloropropane due to the presence of bromine leaving groups in all three compounds. This reaction leads to

formation of the reactive episulfonium metabolites that are implicated in testicular toxicity. The absence of testicular toxicity following exposure to 1,2-dichloroethane (see Figure A-1) may be attributable to reduced episulfonium formation secondary to the expected slower rate of GSH conjugation resulting from lower reactivity of the chlorine leaving group relative to bromine (see "Metabolic Analogues" section for more details).

In the absence of repeated-exposure oral toxicity data for 1-bromo-2-chloroethane, there is no information with which to clearly identify or rule out candidate analogues based on toxicity comparisons. However, based on the expected formation of the episulfonium ion following metabolism of 1-bromo-2-chloroethane, the mechanism of action for the critical testicular effects following exposure to 1,2-dibromo-3-chloropropane (and by analogy 1,2-dibromoethane) is plausible for 1-bromo-2-chloroethane as well.

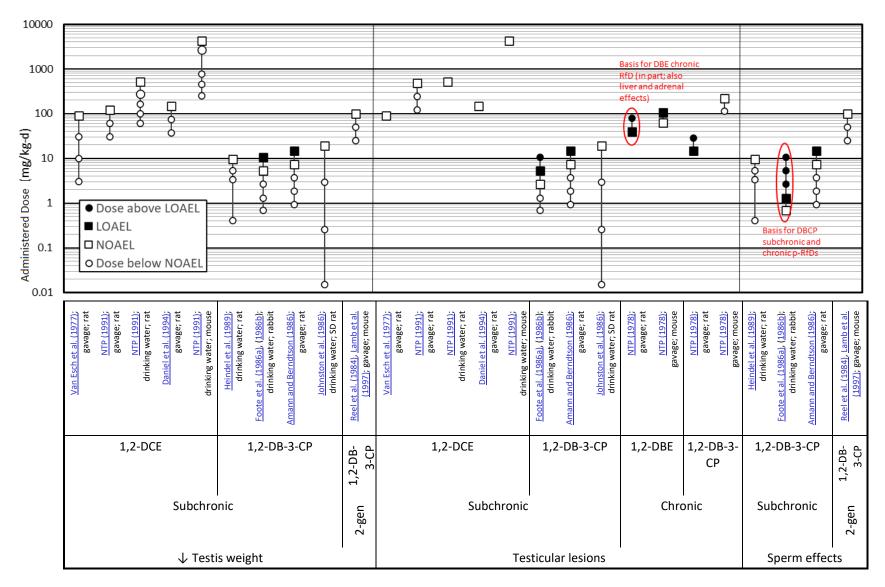


Figure A-1. Testicular Effects Following Oral Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004)]

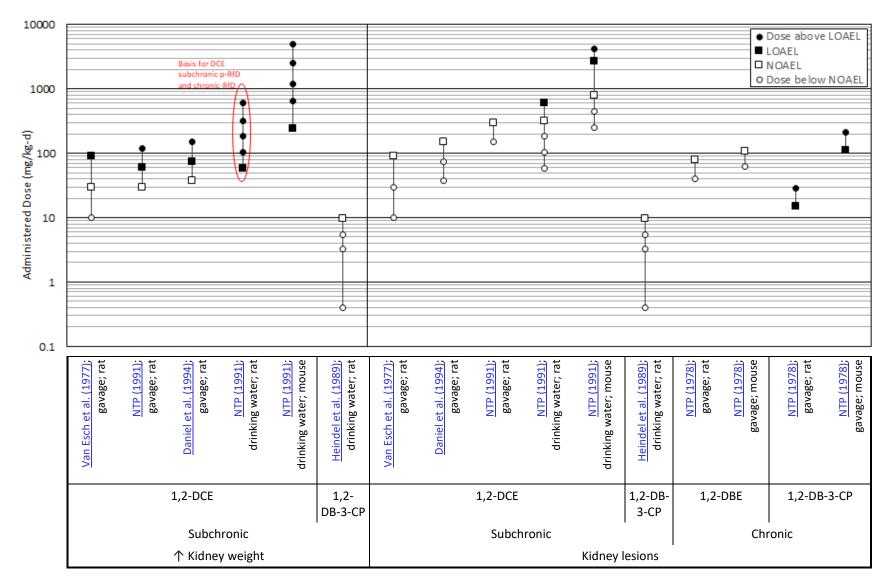


Figure A-2. Kidney Effects Following Oral Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004)]

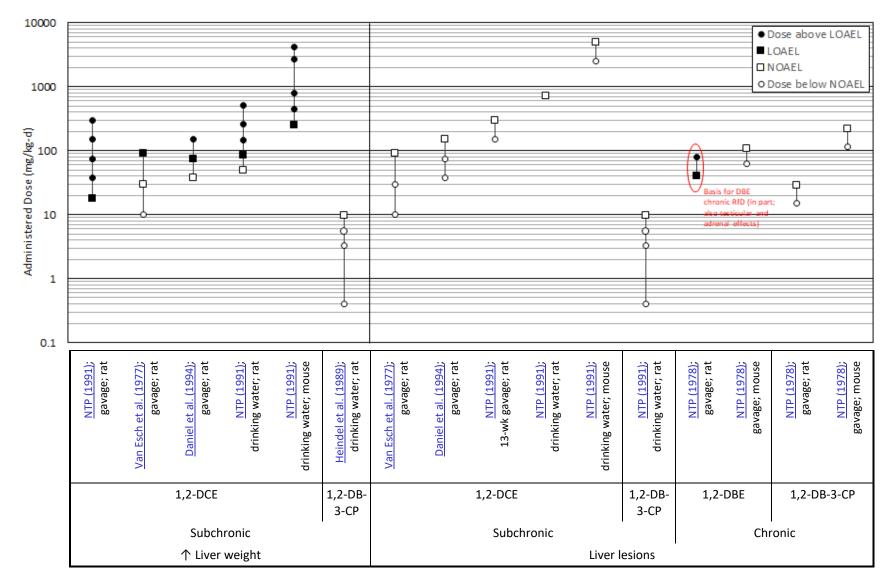


Figure A-3. Liver Effects Following Oral Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004)]

Toxicity-Like Analogues—Inhalation Table A-4 summarizes available inhalation toxicity values for the compounds identified as potential structural analogues for 1-bromo-2-chloroethane.

| Table A-4. Comparison of Available Inhalation Toxicity Data for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Candidate Analogues | | | | |
|--|----------------------------|---|--|---|
| Parameter | 1-Bromo- 2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo- 3-Chloropropane |
| Structure | ci~~ ^{Br} | Br | ci⁄⁄ ^{Ci} | Cl Br Br |
| CASRN | 107-04-0 | 106-93-4 | 107-06-2 | 96-12-8 |
| Repeated-dose toxic | ity—subchronic | | | |
| POD (mg/m ³) | NA | NA | 22 | 0.17 |
| POD type | NA | NA | LOAEL (ADJ) | NOAEL (HEC) |
| Subchronic UF _C | NA | NA | 300 (UF _D , UF _H , UF _L) | 100 (UF _A , UF _D , UF _H) |
| Subchronic p-RfC (mg/m ³) | NA | NA | 7×10^{-2} | 2×10^{-3} |
| | | Neurobehavioral impairment (impaired visual-motor reactions) | Testicular effects (decreased absolute testis-weight, atrophy, loss of spermatogenic elements in seminiferous tubules) | |
| Species | NA | NA | Human | Rabbit |
| Duration NA NA | | NA | No information on length of employment or duration of exposure was reported; treated as subchronic in the PPRTV assessment | 14 wk |
| Route (method) | NA | NA | Inhalation (occupational) | Inhalation (whole body; 6 h/d, 5 d/wk) |
| Source | NA | <u>U.S. EPA (2004)</u> | <u>U.S. EPA (2010)</u> | <u>U.S. EPA (2006);</u> <u>U.S. EPA (2003)</u> |
| Repeated-dose toxic | ity—chronic | | | |
| POD (mg/m ³) | NA | 2.8 | 22 | 0.17 |
| POD type | NA | BMCL ₁₀ (HEC) | LOAEL (ADJ) | NOAEL (HEC) |
| Chronic UF _C | NA | $300 (UF_A, UF_D, UF_H)$ | 3,000 (UF _D , UF _H , UF _L , UF _S) | 1,000 (UF _A , UF _D , UF _H , UF _S) |
| Chronic RfC/p-RfC (mg/m ³) | NA | 9 × 10 ⁻³ | 7×10^{-3} | 2×10^{-4} |

| Parameter | 1-Bromo- 2-Chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dibromo- 3-Chloropropane | |
|---------------------------------------|---|---|--|---|--|
| Critical effects ^a | NA | Suppurative inflammation of nasal cavity in female mice | Neurobehavioral impairment (impaired visual-motor reactions) | Testicular effects (decreased absolute testis weight, atrophy, loss of spermatogenic elements in seminiferous tubules) | |
| Species | NA | Mouse | Human | Rabbit | |
| Duration | NA | 78 wk (M) 90–106 wk (F) | No information on length of employment or duration of exposure was reported; treated as subchronic in the PPRTV assessment | 14 wk | |
| Route (method) NA | | Inhalation (whole body; 6 h/d, 5 d/wk) | Inhalation (occupational) | Inhalation (whole body; 6 h/d, 5 d/wk) | |
| Source | NA | <u>U.S. EPA (2004)</u> | <u>U.S. EPA (2010)</u> | U.S. EPA (2006); U.S. EPA (2003) | |
| Acute inhalation l | ethality data | | | | |
| LC ₅₀ (mg/m ³) | 15,000–25,000 (30-min; unspecified species) | ~1,500 (9 h; rat) | 7,758 (4 h; rat) | ND | |
| Source | U.S. EPA (1985) | U.S. EPA (2004) | U.S. EPA (2020a) | U.S. EPA (2020b) | |

Table A-4 Comparison of Available Inhalation Toxicity Data for

^aExposure-response arrays were prepared to illustrate the dose-response relationship for testicular, kidney, liver, and respiratory tract effects in experimental animal inhalation studies across the candidate analogue compounds (see Figures A-4, A-5, A-6, and A-7, respectively).

ADJ = duration adjusted; $BMCL_{10} = 10\%$ benchmark concentration lower confidence limit; F = female(s); HEC = human equivalent concentration; LC_{50} = median lethal concentration;

LOAEL = lowest-observed-adverse-effect level; M = male(s); NA = not applicable; ND = no data;

NOAEL = no-observed-adverse-effect level; POD = point of departure; PPRTV = provisional peer-reviewed toxicity value; p-RfC = provisional reference concentration; RfC = inhalation reference concentration;

 UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_{H} = intraspecies uncertainty factor; UF_{L} = LOAEL-to-NOAEL uncertainty factor; UF_{S} = subchronic-to-chronic uncertainty factor.

No repeated-exposure inhalation toxicity data are available for 1-bromo-2-chloroethane. Acute inhalation toxicity data are limited to a single study by Valade et al. (1957) as cited in U.S. <u>EPA (1985)</u>, which reports a 30-minute median lethal concentration (LC₅₀) value of 15,000–25,000 mg/m³ in unspecified laboratory animals. This study also indicates that acute inhalation of "several alkyl halides" caused ataxia in dogs, rats, and guinea pigs; however, it is unclear whether these halides included 1-bromo-2-chloroethane.

Inhalation toxicity values are available for 1,2-dichloroethane, 1,2-dibromoethane, and 1,2-dibromo-3-chloropropane (see Table A-4), and the data supporting these values are extensive [reviewed by <u>U.S. EPA (2010, 2006, 2004, 2003)</u>]. Target organs and systems of toxicity for potential analogues include the testis, kidney, liver, nervous system, and respiratory tract (see exposure-response arrays in Figures A-4, A-5, A-6, and A-7). Although neurobehavioral impairment was observed following occupational exposure to 1,2-dichloroethane (impaired visual-motor reactions in two of three tests), an exposure-response array was not prepared for this endpoint because no additional neurobehavioral studies were identified for any candidate analogue in humans or laboratory animals. No information is available regarding the potential mode of action (MOA) for the observed neurobehavioral effects in humans or animals. Upon review of the available inhalation toxicity data, testicular and respiratory effects are identified as the most sensitive toxicity endpoints among the candidate analogues.

Epidemiology studies of 1,2-dibromo-3-chloropropane-exposed production workers, farmers, and pesticide applicators have demonstrated impaired testicular function (decreased spermatogenesis and sperm count and altered sperm morphology) in exposed humans. Testicular effects were identified as critical effects for assessment of 1,2-dibromo-3-chloropropane inhalation toxicity in animal studies. A comparison of rat, mouse, and rabbit data indicates that the rabbit is the most sensitive of these species for testicular effects (by both inhalation and oral exposure). Oral data indicate that testicular toxicity of 1,2-dibromo-3-chloropropane is mediated exclusively by the GSH conjugation pathway (Omichinski et al., 1988a; Søderlund et al., 1988). This is likely the case also for 1,2-dibromoethane and 1-bromo-2-chloroethane, which are anticipated to undergo GSH conjugation at similar rates because of the expected similar reactivity of the bromine leaving groups. This reaction leads to the formation of the reactive episulfonium ion metabolites that are implicated in testicular toxicity (see "Toxicity-Like Analogues—Oral" section above for more details).

Inhalation exposure to the candidate analogues 1,2-dibromoethane and 1,2-dibromo-3-chloropropane in rats and mice produced respiratory tract lesions, with nasal effects occurring at lower concentrations than lesions observed in the bronchial, bronchiolar, and alveolar regions. Respiratory lesions may result, at least in part, from the production of HBr (from both CYP450 oxidation and GSH conjugation) in the respiratory tract. Nasal irritation has been observed in volunteers and laboratory animals exposed to HBr gas (NRC, 2014). By analogy, HBr formation and respiratory lesions may also be expected to occur for the 1-bromo-2-chloroethane target compound. In contrast, 1,2-dichloroethane does not produce nasal or other respiratory tract lesions following inhalation exposure. This may be partially explained by the slower release of chlorine relative to bromine during GSH conjugation, resulting in a reduced rate of generation of HCl in the respiratory tract of exposed animals (relative to HBr from the other compounds). The lack of effect by 1,2-dichloroethane supports the hypothesized role of HBr in the observed respiratory effects.

Other cytotoxic mechanisms may contribute to the nasal toxicity of 1,2-dibromoethane; these include lipid peroxidation and/or protein binding induced by anticipated or experimentally determined metabolites (e.g., 2-bromoacetaldehyde) (<u>U.S. EPA, 2004</u>). By analogy, formation of 2-bromoacetaldehyde and subsequent nasal and/or respiratory lesions may also be expected to occur for the 1-bromo-2-chloroethane target compound.

In the absence of repeated-exposure inhalation toxicity data for 1-bromo-2-chloroethane, there is no information with which to clearly identify or rule out candidate analogues based on toxicity comparisons. However, based on the expected formation of the episulfonium ion following metabolism of 1-bromo-2-chloroethane, the mechanism of action for testicular toxicity following exposure to 1,2-dibromo-3-chloropropane (and by analogy 1,2-dibromoethane) is plausible for 1-bromo-2-chloroethane. Based on the expected formation of HBr and 2-bromoacetaldehyde following metabolism of 1-bromo-2-chloroethane.

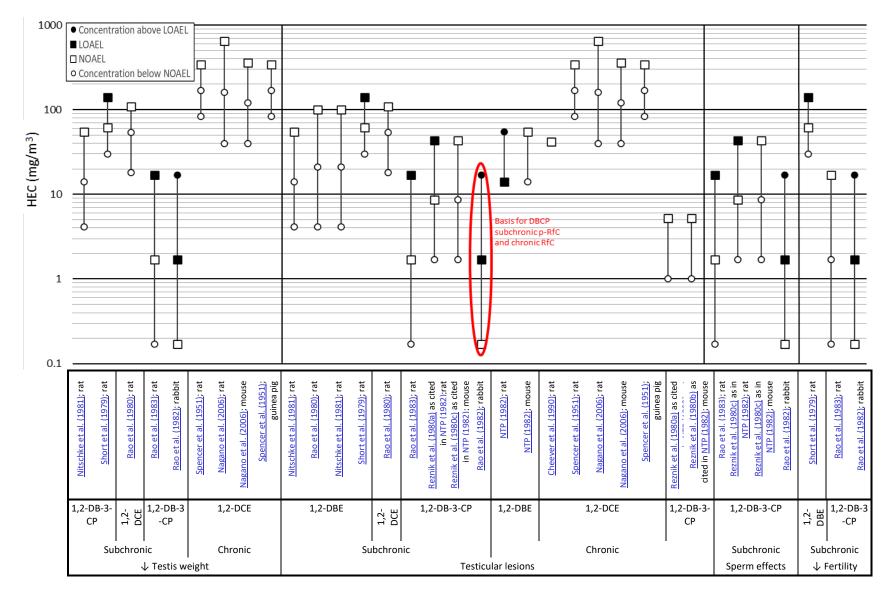


Figure A-4. Testicular Effects Following Inhalation Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004, 2003)]

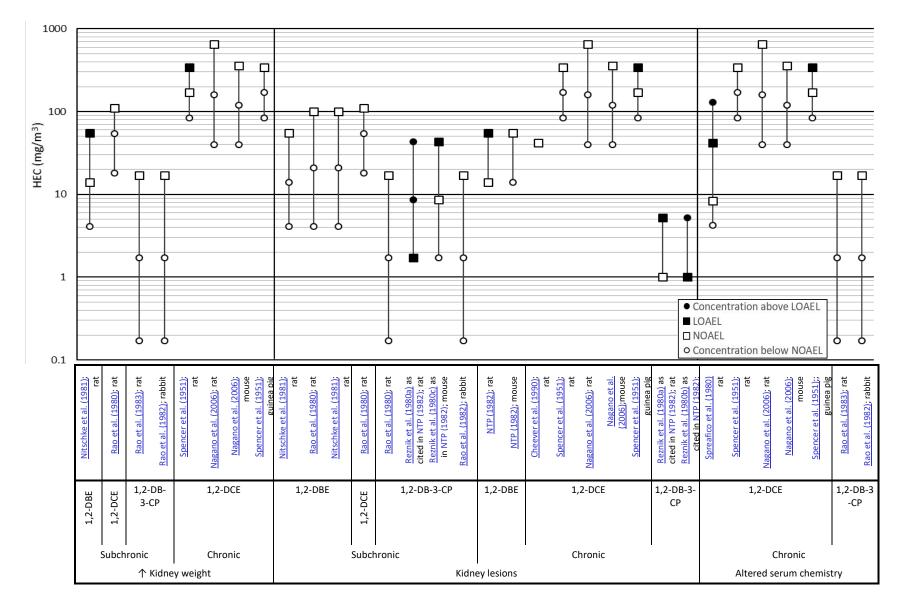


Figure A-5. Kidney Effects Following Inhalation Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004, 2003)]

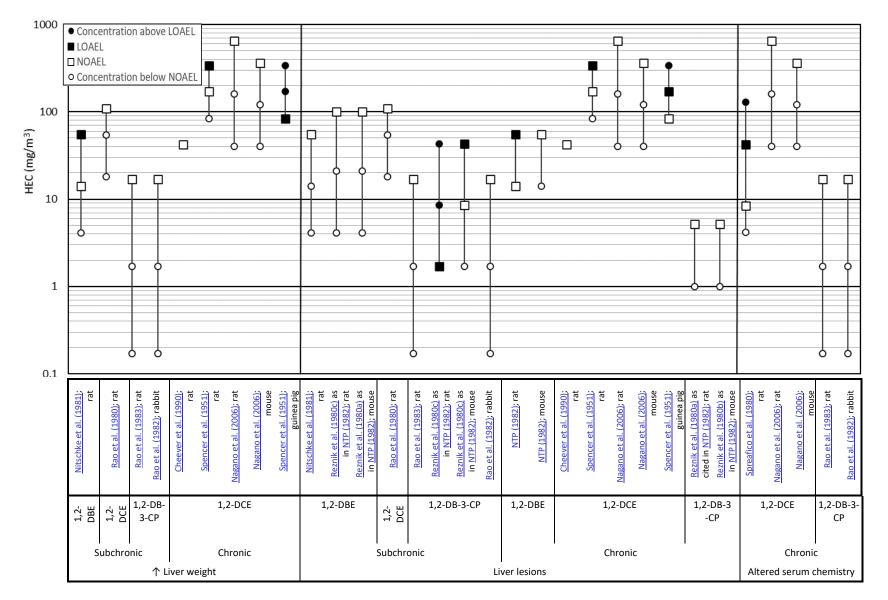


Figure A-6. Liver Effects Following Inhalation Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004, 2003)]

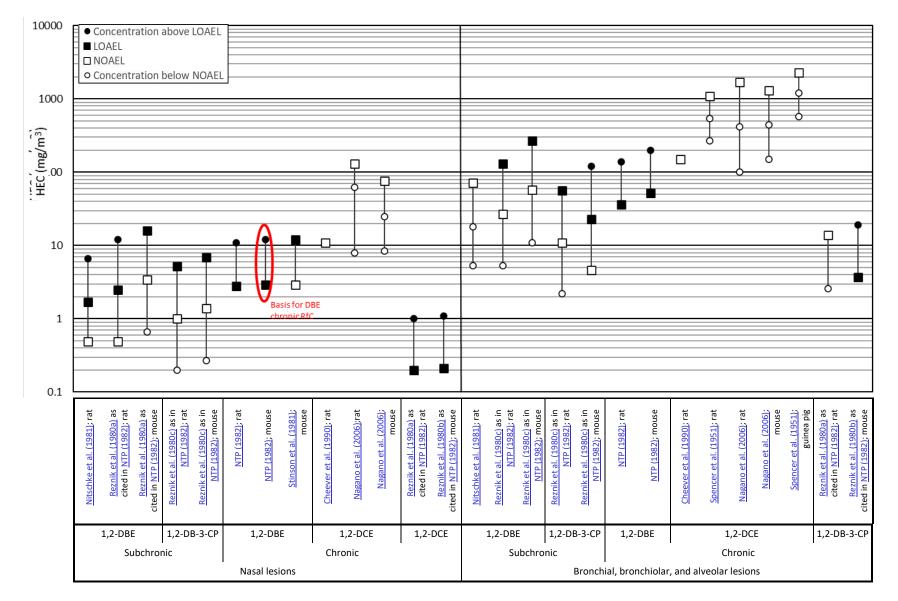


Figure A-7. Respiratory Tract Effects Following Inhalation Exposure to 1,2-Dichloroethane (1,2-DCE), 1,2-Dibromoethane (1,2-DBE), or 1,2-Dibromo-3-Chloropropane (1,2-DB-3-CP) [reviewed by U.S. EPA (2010, 2006, 2004, 2003)]

Weight-of-Evidence Approach

To select the best analogue chemical based on all of the information described within the three analogue similarity tiers prescribed in <u>Wang et al. (2012)</u>, the following considerations are used in a weight-of-evidence (WOE) approach: (1) lines of evidence from U.S. EPA assessments are preferred; (2) biological and toxicokinetic data are preferred over the structural similarity scores; (3) lines of evidence that indicate pertinence to humans are preferred; (4) chronic studies are preferred over subchronic studies when selecting an analogue for a chronic value; (5) chemicals with more sensitive toxicity values may be favored; and (6) if there are no clear indications as to the best analogue chemical based on the other considerations, then the candidate analogue with the highest structural similarity scores may be preferred.

Oral

The WOE approach used to select the analogue compound for 1-bromo-2-chloroethane is based on mechanistic and metabolism considerations related to the most sensitive critical effect (i.e., testicular toxicity). Unlike 1,2-dibromoethane and 1,2-dibromo-3-chloropropane, 1,2-dichloroethane is not a testicular toxicant. This may be explained by differences in the rate and extent of GSH adduct formation due in part to differences in the chemical reactivity of the brominated versus the chlorinated compounds. 1,2-Dibromoethane will react more rapidly with GSH than 1,2-dichloroethane because bromide is a better chemical leaving group in comparison to the chloride. That is, the C-Br bond is more readily cleaved than is the C-Cl bond. The results of this difference are demonstrated by the metabolic products and GSH adducts formed by 1-bromo-2-chloroethane and 1,2-dibromo-3-chloropropane. In both cases, bromine is preferentially displaced when both C-Br and C-Cl bonds are available. The only GSH-conjugated product detected from 1-bromo-2-chloroethane metabolism was CEG, and episulfonium products of 1,2-dibromo-3-chloropropane contain chlorine, indicating that bromine was preferentially released during metabolism (Dekant and Vamvakas, 1993; Jean and Reed, 1992; Humphreys et al., 1991; Marchand and Reed, 1989). These data suggest that from the available identified potential analogue chemicals, chemicals with a bromine leaving group are expected to be more relevant as an analogue for 1-bromo-2-chloroethane. The presence of this group is expected to significantly influence the rate of GSH conjugation (and ultimately the rate of toxic moiety production) over analogues containing chlorine leaving groups.

By analogy, 1-bromo-2-chloroethane can be expected to form the same episulfonium ion that is formed from 1,2-dibromoethane and 1,2-dichloroethane after GSH conjugation (see Table A-2). In the initial reaction, bromine will be preferentially displaced from 1-bromo-2-chloroethane. Although this compound and 1,2-dichloroethane are both expected to lead to the same initial adduct, 1-bromo-2-chloroethane is expected to react faster because of the presence of a bromide versus chloride leaving group, leading to a higher localized concentration of the proximal toxicant. Therefore, the rate of initial adduct formation is expected to be more similar to 1,2-dibromoethane and 1,2-dibromo-3-chloropropane than to 1,2-dichloroethane. Available data suggest that the cyclization of the GSH adducts for each respective analogue is rapid and immediate, and that this step (leading to the production of the toxic episulfonium ion moiety) is not likely to be the rate-limiting step in the metabolism of the parent chemicals.

Given the demonstrated significance of the presence of a bromide leaving group in metabolism and expected toxicity, 1-bromo-2-chloroethane, 1,2-dibromoethane, and

1,2-dibromo-3-chloropropane were determined to be more appropriate analogues compared with 1,2-dichloroethane. Ultimately, 1,2-dibromo-3-chloropropane is selected as the most appropriate analogue compound for both subchronic and chronic effects because the demonstrated testicular effect of this chemical on sperm is the most sensitive measure of toxicity among the favored brominated analogues.

Inhalation

The WOE approach used to select the analogue compound for 1-bromo-2-chloroethane is based on mechanistic considerations related to sensitive effects for the candidate brominated analogues following inhalation exposure (i.e., testicular toxicity and respiratory effects). The absence of testicular and respiratory tract effects for 1,2-dichloroethane may be explained, at least in part, by a decrease in the rate and extent of GSH conjugation (i.e., bromine is a better chemical leaving group than chlorine), leading to a lower localized concentration of proximal toxicants (i.e., episulfonium ion, HCl). 1-Bromo-2-chloroethane can be expected to form both the reactive episulfonium metabolite (responsible for testicular effects) and HBr (which may contribute to respiratory lesions) at a similar rate as 1,2-dibromoethane and 1,2-dibromo-3-chloropropane. Both 1,2-dibromoethane and 1,2-dibromo-3-chloropropane release HBr during GSH conjugation, and this is expected to occur for 1-bromo-2-chloroethane. As described above, the rate of GSH conjugation among brominated compounds, rapid cyclization of GSH conjugates to generate reactive episulfonium ions, and the release of HBr during GSH conjugation suggest that the brominated analogues are more appropriate surrogate chemicals for 1-bromo-2-chloroethane than 1,2-dichloroethane. However, these characteristics do not provide a means to sufficiently differentiate amongst the brominated analogues.

Ultimately, 1,2-dibromo-3-chloropropane is selected as the analogue compound for both subchronic and chronic inhalation exposure because the demonstrated effect of this chemical on sperm is among the most sensitive measures of toxicity among the favored brominated analogues, and the relevance of this endpoint to humans is supported by the observation of similar effects in exposed workers (U.S. EPA, 2006).

NONCANCER ORAL TOXICITY VALUES Derivation of a Screening Subchronic Provisional Reference Dose

Based on the overall analogue approach presented in this PPRTV assessment, 1,2-dibromo-3-chloropropane is selected as the analogue for 1-bromo-2-chloroethane for deriving a screening subchronic p-RfD. The study used for the U.S. EPA subchronic p-RfD for 1,2-dibromo-3-chloropropane was a 10-week reproductive study in rabbits [Foote et al. (1986a, b) as cited in U.S. EPA (2006)] described as follows:

Groups of 6 male Dutch rabbits were exposed to drinking water that provided reported DBCP intakes of 0, 0.94, 1.88, 3.75, 7.5 or 15.0 mg/kg on 5 days/week (0, 0.7, 1.3, 2.7, 5.4 or 10.7 mg/kg-day) for 10 weeks (Foote et al., 1986a, 1986b). General health, body weight, semen quality, and libido were evaluated throughout the study. Assessments of fertility (mated with untreated females) and serum levels of reproductive hormones (follicle stimulation hormone, luteinizing hormone and testosterone) were performed during the last week of the study. Endpoints evaluated following sacrifice at the end of the study

included organ weights (liver, kidneys, testes, epididymides, accessory sex glands), quantitative histology of testes and epididymides, and sperm morphology and forward motility and morphology. There were no statistically significant (p < 0.05) changes in any of the study endpoints at 0.7 mg/kg-day. Effects observed at higher doses included dose-related reductions in numbers of all germ cell types within Stage I seminiferous tubular cross sections (significantly reduced numbers of spermatogonia and preleptotene spermatocytes at >1.3 mg/kg-day, pachytene spermatocytes at >2.7 mg/kg-day, and round spermatids at >5.4 mg/kg-day) (Table 1 [in U.S. EPA 2006]). Other effects included doserelated significantly reduced numbers of leptotene primary spermatocytes per Sertoli cell at >2.7 mg/kg-day, and significantly reduced mean diameter of seminiferous tubules and increased percentage of sperm with abnormal tails at >5.4 mg/kg-day (Table 2 [in U.S. EPA 2006]). Testis weight and volume, and sperm production (number of seminiferous tubules with round or elongating spermatids), output (ejaculate volume times sperm concentration) and motility were reduced, and serum FSH level was increased, at 10.7 mg/kg-day (Table 3 [in U.S. EPA 2006].). Fertility was not affected at any dose level, as assessed by number of males producing young, number or percentage of live births, total number of young, average litter size, and gestation length. The results summarized above are based on comparisons of mean data from the treated and control groups. Regression analyses showed highly significant correlations between DBCP dosage and essentially all of the testicular responses. The findings of this study indicate that rabbits are more sensitive than rats to testicular effects of DBCP. This study identified a NOAEL of 0.7 mg/kg-day and LOAEL of 1.3 mg/kg-day for reproductive toxicity in male rabbits.

The critical effect in this study was testicular toxicity in male rabbits; the no-observed-adverse-effect level (NOAEL) of 0.7 mg/kg-day was used as the point of departure (POD) for 1,2-dibromo-3-chloropropane (U.S. EPA, 2006) and is adopted as the analogue POD for the current assessment of 1-bromo-2-chloroethane.

For the current assessment, the NOAEL of 0.7 mg/kg-day was converted to a human equivalent dose (HED) according to current (U.S. EPA, 2011c) guidance. In *Recommended Use of Body Weight^{3/4} as the Default Method in Derivation of the Oral Reference Dose* (U.S. EPA, 2011c), the Agency endorses body-weight scaling to the 3/4 power (i.e., BW^{3/4}) as a default to extrapolate toxicologically equivalent doses of orally administered agents from all laboratory animals to humans for the purpose of deriving an RfD from effects that are not portal-of-entry effects.

Following <u>U.S. EPA (2011c)</u> guidance, the POD for testicular effects in male rabbits is converted to an HED by applying a dosimetric adjustment factor (DAF) derived as follows:

 $DAF = (BW_a^{1/4} \div BW_h^{1/4})$

where

DAF = dosimetric adjustment factor BW_a = animal body weight BW_h = human body weight

Using a reference BW_a of 2.86 kg for rabbits and a reference BW_h of 70 kg for humans (U.S. EPA, 1988), the resulting DAF is 0.45. Applying this DAF to the NOAEL of 0.7 mg/kg-day yields a POD (HED) as follows:

POD (HED) = NOAEL (mg/kg-day) × DAF = $0.7 \text{ mg/kg-day} \times 0.45$ = 0.3 mg/kg-day

The <u>U.S. EPA (2006)</u> subchronic p-RfD for 1,2-dibromo-3-chloropropane was derived using a composite uncertainty factor (UF_C) of 300, reflecting 10-fold uncertainty factors for interspecies extrapolation and intraspecies variability and a 3-fold uncertainty factor for database uncertainties (UF_A, UF_H, and UF_D, respectively). <u>Wang et al. (2012)</u> indicated that the uncertainty factors typically applied in deriving a toxicity value for the selected analogue are the same as those applied to the chemical of concern unless additional information is available. For 1-bromo-2-chloroethane, a UF_A of 3 is applied because cross-species dosimetric adjustment was performed, and a UF_D of 10 was used to reflect the lack of repeated-dose toxicity data; the UF_H remained the same as for 1,2-dibromo-3-chloropropane. Thus, the screening subchronic p-RfD for 1-bromo-2-chloroethane is derived using a UF_C of 300.

| Screening Subchronic p-RfD | = Analogue POD (HED) \div UF _C |
|----------------------------|---|
| | $= 0.3 \text{ mg/kg-day} \div 300$ |
| | $= 1 \times 10^{-3}$ mg/kg-day |

Table A-5 summarizes the uncertainty factors for the screening subchronic p-RfD for 1-bromo-2-chloroethane.

Table A-5. Uncertainty Factors for the Screening Subchronic p-RfD for1-Bromo-2-Chloroethane (CASRN 107-04-0)

| | - | |
|-----------------|-------|---|
| UF | Value | Justification |
| UFA | 3 | A UF _A of 3 (10 ^{0.5}) is applied to account for uncertainty in characterizing the toxicokinetic or toxicodynamic differences between rabbits and humans following 1-bromo-2-chloroethane exposure. The toxicokinetic uncertainty has been accounted for by calculating an HED through application of a DAF as outlined in the U.S. EPA's <i>Recommended Use of Body Weight^{3/4} as the Default Method in Derivation of the Oral Reference Dose</i> (U.S. EPA, 2011c). Dosimetric adjustment calculations were performed on the POD for the selected analogue, 1,2-dibromo-3-chloropropane. |
| UF _D | 10 | A UF_D of 10 is applied to reflect database limitations for 1,2-dibromo-3-chloropropane and the absence of repeated-dose toxicity data for 1-bromo-2-chloroethane. |
| UF _H | 10 | A UF_H of 10 is applied for interindividual variability to account for human-to-human variability in susceptibility in the absence of quantitative information to assess the toxicokinetics and toxicodynamics of 1-bromo-2-chloroethane in humans. |
| UF _L | 1 | A UF _L of 1 is applied for LOAEL-to-NOAEL extrapolation because the POD is a NOAEL. |
| UFs | 1 | A UF _s of 1 is applied because a subchronic study was selected as the principal study for the subchronic assessment. |
| UF _C | 300 | Composite uncertainty factor = $UF_A \times UF_D \times UF_H \times UF_L \times UF_S$. |

DAF = dosimetric adjustment factor; HED = human equivalent dose; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfD = provisional reference dose; UF = uncertainty factor; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_L = LOAEL-to-NOAEL uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.

Derivation of a Screening Chronic Provisional Reference Dose

1,2-Dibromo-3-chloropropane is also selected as the analogue for 1-bromo-2-chloroethane for deriving a screening chronic p-RfD. The key study and calculation of the POD (HED) are described above for the subchronic p-RfD. The uncertainty factors used for the screening subchronic p-RfD (UF_A of 3, UF_H of 10, and UF_D of 10) are applied, and a UFs of 10 is applied to account for extrapolation from a subchronic to a chronic duration (consistent with the UF_S applied during the derivation of the chronic p-RfD for

1,2-dibromo-3-chloropropane). Thus, the screening chronic p-RfD for 1-bromo-2-chloroethane is derived using a UF_C of 3,000.

| Screening Chronic p-RfD | = | Analogue POD (HED) ÷ UF _C |
|-------------------------|---|--------------------------------------|
| | = | 0.3 mg/kg-day ÷ 3,000 |
| | = | 1 × 10 ⁻⁴ mg/kg-day |

Table A-6 summarizes the uncertainty factors for the screening chronic p-RfD for 1-bromo-2-chloroethane.

Table A-6. Uncertainty Factors for the Screening Chronic p-RfD for1-Bromo-2-Chloroethane (CASRN 107-04-0)

| UF | Value | Justification |
|-----------------|-------|---|
| UFA | 3 | A UF _A of 3 (10 ^{0.5}) is applied to account for uncertainty in characterizing the toxicokinetic or toxicodynamic differences between rabbits and humans following 1-bromo-2-chloroethane exposure. The toxicokinetic uncertainty has been accounted for by calculating an HED through application of a DAF as outlined in the U.S. EPA's <i>Recommended Use of Body Weight</i> ^{3/4} <i>as the Default Method in Derivation of the Oral Reference Dose</i> (U.S. EPA, 2011c). Dosimetric adjustment calculations were performed on the POD for the selected analogue, 1,2-dibromo-3-chloropropane. |
| UF _D | 10 | A UF _D of 10 is applied to reflect database limitations for 1,2-dibromo-3-chloropropane and the absence of repeated-dose toxicity data for 1-bromo-2-chloroethane. |
| UF _H | 10 | A UF_H of 10 is applied for interindividual variability to account for human-to-human variability in susceptibility in the absence of quantitative information to assess the toxicokinetics and toxicodynamics of 1-bromo-2-chloroethane in humans. |
| UF_L | 1 | A UF _L of 1 is applied for LOAEL-to-NOAEL extrapolation because the POD is a NOAEL. |
| UFs | 10 | A UF _s of 10 is applied because a subchronic study was selected as the principal study for the chronic assessment. |
| UF _C | 3,000 | Composite uncertainty factor = $UF_A \times UF_D \times UF_H \times UF_L \times UF_S$. |

DAF = dosimetric adjustment factor; HED = human equivalent dose; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfD = provisional reference dose; UF = uncertainty factor; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_L = LOAEL-to-NOAEL uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.

NONCANCER INHALATION TOXICITY VALUES

Derivation of a Screening Subchronic Provisional Reference Concentration

Based on the overall analogue approach presented in this PPRTV assessment, 1,2-dibromo-3-chloropropane is selected as the most appropriate analogue for 1-bromo-2-chloroethane for deriving a screening subchronic p-RfC. The study used for the subchronic p-RfC for 1,2-dibromo-3-chloropropane was a 14-week study in rabbits [Rao et al. (1982) as cited in U.S. EPA (2006)]. U.S. EPA (2006) described the study as follows:

In rabbits, reproductive toxicity was evaluated in groups of 10 New Zealand white males (age 6 months) that were exposed to 0, 0.1, 1 or 10 ppm (0, 0.94, 9.4 or 94 mg/m³) vapor for 6 hours/day, 5 days/week for 14 weeks, and observed for the following 32 weeks (0, 0.1 and 1 ppm groups) or 38 weeks (10 ppm group) (Rao et al., 1982). The 10 ppm rabbits were exposed for only 8 weeks due to high mortality (apparently from pneumonia). Body weight and hematological and clinical chemistry parameters were evaluated, but no exposure related changes were found. No gross lesions were found in the lungs or upper respiratory tract, but these tissues were not examined histologically. Semen was collected during the exposure and recovery periods to assess sperm motility, viability and counts. The average sperm count of the 10-ppm rabbits was significantly less than that of the controls after 7 weeks of exposure, and remained decreased for the duration of the exposure and observation periods. At 1 ppm (9.4 mg/m^3) , sperm counts were significantly reduced, compared with controls, from weeks 11 to 13 of exposure. At 0.1 ppm (0.94 mg/m^3) , sperm counts were sporadically lower than control values (significantly reduced at only one interim time point). The percentage of live sperm in the semen of the 10 ppm (94 mg/m^3) rabbits was also significantly reduced compared to controls during weeks 8-26. Rabbits exposed to 1 ppm (9.4 mg/m^3) , but not those exposed to 0.1 ppm (0.94 mg/m^3) , exhibited significant decreases in the percentage of live sperm during weeks 6, 12 and 13. From the 8th week of exposure onward, the 10-ppm (94 mg/m^3) rabbits had a marked decrease in the percentage of progressively motile sperm; no consistent statistically significant decreases in this endpoint were found at <1 ppm (9.4 mg/m^3) (Table 4 [in U.S. EPA 2006]). Abnormal spermatozoa within the seminiferous tubules were counted in 3-4 rabbits per group; the percentage of abnormal sperm at 14 weeks was 5% in controls, 10% at 0.1 ppm (0.94 mg/m^3) , and 18% at 1 ppm (9.4 mg/m^3) .

To assess the effects of DBCP on fertility in the rabbits, exposed males were mated to unexposed females at study weeks 14 and 41 (Rao et al., 1982) (Tables 5, 6 [in U.S. EPA 2006]). There were no effects on the libido of the exposed male rabbits during week 14, based on percentages of males (78–100%) that copulated with unexposed females. Five of the 10 males exposed to 10 ppm were infertile (none of the females that they were mated with became pregnant). The mean number of implantations/litter in the 1 ppm (9.4 mg/m³) group was significantly less than that of the control group. During week 41 (27 weeks post-exposure), all rabbits exposed to 0.1 or 1 ppm (0.94 or 9.4 mg/m³) DBCP produced normal litters, and 2 of the 5 infertile males exposed to 10 ppm (94 mg/m^3) recovered (sperm counts increased) and produced normal litters. Serum levels of follicle stimulating hormone (FSH) were significantly elevated at 14 weeks in the males exposed to 1 ppm (9.4 mg/m³) and at 46 weeks in the males exposed to 10 ppm (94 mg/m³), but serum levels of testosterone were unchanged (Table 7 [in U.S. EPA 2006]). The increases in serum FSH were consistent with the decreases in sperm count. Gross pathologic examinations showed small testes size in rabbits exposed to 1 or 10 ppm. Testes weight was significantly decreased to 50% of control values (week 14) in the group exposed to 1 ppm (9.4 mg/m³) and to 75% of control values (week 8) in the group exposed to 10 ppm. Histological examinations showed reproductive system effects that included atrophy of the testes, epididvmides, and accessory sex glands, including the prostate. The testicular atrophy was severe, as characterized by nearly complete or complete loss of spermatogenic elements in nearly all seminiferous tubules. Following the recovery period, tubular regeneration was observed in the testes of some 10 ppm (94 mg/m³) rabbits (3/5 had regeneration such that 25% of the seminiferous tubules appeared normal). At 1 ppm, testicular recovery was reported to be nearly complete in some rabbits (incidences not given). The testes of the 0.1 ppm rabbits appeared normal. The lack of exposure-related adverse testicular and fertility effects at 0.1 ppm indicates that this study identified a

NOAEL of 0.1 ppm (0.94 mg/m³) and LOAEL of 1 ppm (9.4 mg/m³) for reproductive effects in rabbits.

The critical effect in this study was testicular toxicity in male rabbits; the NOAEL of 0.94 mg/m^3 was used as the POD for 1,2-dibromo-3-chloropropane. The <u>U.S. EPA (2006)</u> calculated a POD (human equivalent concentration [HEC]) according to <u>U.S. EPA (1994)</u> guidance for Category 3 gases by adjusting intermittent exposure levels to a continuous exposure basis (<u>U.S. EPA, 2002</u>) and multiplying the result by a ratio of the animal blood-gas partition coefficient (concentration) for 1,2-dibromo-3-chloropropane to the human blood-gas partition coefficient (concentration \div concentration) for 1,2-dibromo-3-chloropropane. Because blood-air partition coefficients for 1,2-dibromo-3-chloropropane are unknown, a default value of 1 was assigned.

POD (HEC) = NOAEL (mg/m³) × (hours exposed/24 hours) × (days/week exposed/7 days) × (animal blood-gas partition coefficient ÷ human blood-gas partition coefficient) = 0.94 mg/m³ × (6/24) × (5/7) × 1 = 0.17 mg/m³

The POD (HEC) of 0.17 mg/m³ derived for 1,2-dibromo-3-chloropropane by <u>U.S. EPA</u> (2006) is adopted as the analogue POD (HEC) for the current assessment of 1-bromo-2-chloroethane.

The U.S. EPA (2006) subchronic p-RfC for 1,2-dibromo-3-chloropropane was derived using a UF_C of 100 to account for uncertainties due to interspecies extrapolation (UF_A = 3), intraspecies variability (UF_H = 10), and database uncertainties (UF_D = 3) (U.S. EPA, 2006). Wang et al. (2012) indicated that the uncertainty factors typically applied in deriving a toxicity value for the chemical of concern are the same as those applied to the analogue unless additional information is available. Because no repeated-exposure inhalation toxicity data are available for 1-bromo-2-chloroethane, a UF_D of 10 is used; other UF values remain the same as for 1,2-dibromo-3-chloropropane. Thus, the screening subchronic p-RfC for 1-bromo-2-chloroethane is derived using a UF_C of 300, reflecting a UF_A of 3, a UF_H of 10, and a UF_D of 10.

| Screening Subchronic p-RfC | = | Analogue POD (HEC) ÷ UF _C |
|----------------------------|---|--------------------------------------|
| | = | $0.17 \text{ mg/m}^3 \div 300$ |
| | = | $6 \times 10^{-4} \text{ mg/m}^3$ |

Table A-7 summarizes the uncertainty factors for the screening subchronic p-RfC for 1-bromo-2-chloroethane.

| | Table A-7. Uncertainty Factors for the Screening Subchronic p-RfC for1-Bromo-2-Chloroethane (CASRN 107-04-0) | | | | |
|-----------------|--|--|--|--|--|
| UF | Value | Justification | | | |
| UFA | 3 | A UF _A of 3 ($10^{0.5}$) is applied to account for uncertainty associated with extrapolating from animals to humans when cross-species dosimetric adjustment (HEC calculation) is performed. Dosimetric adjustment calculations were performed on the POD for the selected analogue, 1,2-dibromo-3-chloropropane. | | | |
| UF _D | 10 | A UF _D of 10 is applied to reflect database limitations for 1,2-dibromo-3-chloropropane and the absence of repeated-exposure toxicity data for 1-bromo-2-chloroethane. | | | |
| UF _H | 10 | A UF_H of 10 is applied for interindividual variability to account for human-to-human variability in susceptibility in the absence of quantitative information to assess the toxicokinetics and toxicodynamics of 1-bromo-2-chloroethane in humans. | | | |
| UFL | 1 | A UF _L of 1 is applied for LOAEL-to-NOAEL extrapolation because the POD is a NOAEL. | | | |
| UFs | 1 | A UF_s of 1 is applied because a subchronic study was selected as the principal study for the subchronic assessment. | | | |
| UF _C | 300 | Composite uncertainty factor = $UF_A \times UF_D \times UF_H \times UF_L \times UF_S$. | | | |

HEC = human equivalent concentration; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfC = provisional reference concentration; UF = uncertainty factor; UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_L = LOAEL-to-NOAEL uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.

Derivation of a Screening Chronic Provisional Reference Concentration

1,2-Dibromo-3-chloropropane is also selected as the analogue for 1-bromo-2-chloroethane for deriving a screening chronic p-RfC. The key study and calculation of the POD (HEC) are described above for the subchronic p-RfC. The uncertainty factor values used for the screening subchronic p-RfC (UF_A of 3, UF_H of 10, and UF_D of 10) are applied, and a UF_s of 10 is applied to account for extrapolation from a subchronic to a chronic duration (consistent with the UFs applied during the derivation of the chronic RfC for 1,2-dibromo-3-chloropropane) (U.S. EPA, 2003). Thus, the screening chronic p-RfC for 1-bromo-2-chloroethane is derived using a UF_C of 3,000.

| Screening Chronic p-RfC | = | Analogue POD (HEC) \div UF _C |
|-------------------------|---|---|
| | = | 0.17 mg/m3 ÷ 3,000 |
| | = | $6 \times 10^{-5} \text{ mg/m}^3$ |

Table A-8 summarizes the uncertainty factors for the screening chronic p-RfC for 1-bromo-2-chloroethane.

| Table A-8. Uncertainty Factors for the Screening Chronic p-RfC for | |
|--|--|
| 1-Bromo-2-Chloroethane (CASRN 107-04-0) | |

| UF | Value | Justification |
|-----------------|-------|--|
| UFA | 3 | A UF _A of 3 (10 ^{0.5}) is applied to account for uncertainty associated with extrapolating from animals to humans when cross-species dosimetric adjustment (HEC calculation) is performed. Dosimetric adjustment calculations were performed on the POD for the selected analogue, 1,2-dibromo-3-chloropropane. |
| UF _D | 10 | A UF _D of 10 is applied to reflect database limitations for 1,2-dibromo-3-chloropropane and the absence of repeated-exposure toxicity data for 1-bromo-2-chloroethane. |
| UF _H | 10 | A UF_H of 10 is applied for interindividual variability to account for human-to-human variability in susceptibility in the absence of quantitative information to assess the toxicokinetics and toxicodynamics of 1-bromo-2-chloroethane in humans. |
| UF_{L} | 1 | A UF _L of 1 is applied for LOAEL-to-NOAEL extrapolation because the POD is a NOAEL. |
| UFs | 10 | A UF_s of 10 is applied because a subchronic study was selected as the principal study for the chronic assessment. |
| UF_{C} | 3,000 | Composite uncertainty factor = $UF_A \times UF_D \times UF_H \times UF_L \times UF_S$. |

HEC = human equivalent concentration; LOAEL = lowest-observed-adverse-effect level; NOAEL = no-observed-adverse-effect level; POD = point of departure; p-RfC = provisional reference concentration; UF = uncertainty factor(s); UF_A = interspecies uncertainty factor; UF_C = composite uncertainty factor; UF_D = database uncertainty factor; UF_H = intraspecies uncertainty factor; UF_L = LOAEL-to-NOAEL uncertainty factor; UF_S = subchronic-to-chronic uncertainty factor.

APPENDIX B. BACKGROUND AND METHODOLOGY FOR THE SCREENING EVALUATION OF POTENTIAL CARCINOGENICITY

For reasons noted in the main Provisional Peer-Reviewed Toxicity Value (PPRTV) document, there is inadequate information to assess the carcinogenic potential of 1-bromo-2-chloroethane. However, information is available for this chemical which, although insufficient to support a weight-of-evidence (WOE) descriptor and derivation of provisional cancer risk estimates under current guidelines, may be of limited use to risk assessors. In such cases, the Center for Public Health and Environmental Assessment (CPHEA) summarizes available information in an appendix and develops a "screening evaluation of potential carcinogenicity." Appendices receive the same level of internal and external scientific peer review as the provisional cancer assessments in PPRTVs to ensure their appropriateness within the limitations detailed in the document. Users of the information regarding potential carcinogenicity in this appendix should understand that there could be more uncertainty associated with this evaluation than for the cancer WOE descriptors presented in the body of the assessment. Questions or concerns about the appropriate use of the screening evaluation of potential carcinogenicity should be directed to the CPHEA.

The screening evaluation of potential carcinogenicity includes the general steps shown in Figure B-1. The methods for Steps 1 through 8 apply to any target chemical and are described in this appendix. Chemical-specific data for all steps in this process are summarized in Appendix C.

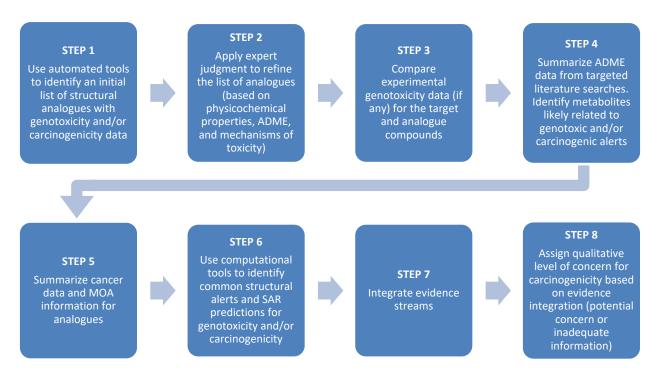


Figure B-1. Steps Used in the Screening Evaluation of Potential Carcinogenicity

STEP 1. USE OF AUTOMATED TOOLS TO IDENTIFY STRUCTURAL ANALOGUES WITH CARCINOGENICITY AND/OR GENOTOXICITY DATA ChemACE Clustering

The U.S. EPA's Chemical Assessment Clustering Engine [ChemACE; U.S. EPA (2011a)] is an automated tool that groups (or clusters) a user-defined list of chemicals based on chemical structure fragments. The methodology used to develop ChemACE was derived from U.S. EPA's Analog Identification Methodology (AIM) tool, which identifies structural analogues for a chemical based on common structural fragments. ChemACE uses the AIM structural fragment recognition approach for analogue identification and applies advanced queries and user-defined rules to create the chemical clusters. The ChemACE cluster outputs are available in several formats and layouts (i.e., Microsoft Excel, Adobe PDF) to allow rapid evaluation of structures, properties, mechanisms, and other parameters, which are customizable based on an individual user's needs. ChemACE clustering has been successfully used with chemical inventories for identifying trends within a series of structurally similar chemicals, demonstrating structural diversity in a chemical inventory, and detecting structural analogues to fill data gaps and/or perform read-across.

For this project, ChemACE is used to identify potential structural analogues of the target compound that have available carcinogenicity assessments and/or carcinogenicity data. An overview of the ChemACE process is shown in Figure B-2.

Create and curate an inventory of chemicals with carcinogenicity assessments and/or cancer data

Figure B-2. Overview of ChemACE Process

The chemical inventory was populated with chemicals from the following databases and lists:

- Carcinogenic Potency Database [CPDB; <u>CPDB (2011)</u>]
- Agents classified by the International Agency for Research on Cancer (IARC) monographs (IARC, 2018)
- National Toxicology Program (NTP) Report on Carcinogens [ROC; NTP (2016a)]
- NTP technical reports (<u>NTP, 2017</u>)
- Integrated Risk Information (IRIS) carcinogens (U.S. EPA, 2017)
- California EPA Prop 65 list (CalEPA, 2020)
- European Chemicals Agency (ECHA) carcinogenicity data available in the Organisation for Economic Co-operation and Development (OECD) Quantitative Structure-Activity Relationship (QSAR) Toolbox (<u>OECD, 2018</u>)
- PPRTVs for Superfund (U.S. EPA, 2020d)

In total, 2,123 distinct substances were identified from the sources above. For the purpose of ChemACE clustering, each individual substance needed to meet the following criteria:

- 1) Substance is not a polymer, metal, inorganic, or complex salt because ChemACE is not designed to accommodate these substances;
- 2) Substance has a CASRN or unambiguous chemical identification; and
- 3) Substance has a unique Simplified Molecular Input Line Entry System (SMILES) notation (encoded molecular structure format used in ChemACE) that can be identified from one of these sources:
 - a. Syracuse Research Corporation (SRC) and Distributed Structure-Searchable Toxicity (DSSTox) database lists of known SMILES associated with unique CASRNs (the combined lists contained >200,000 SMILES) or
 - b. ChemIDplus, U.S. EPA Chemicals Dashboard, or internet searches.

Of the initial list of 2,123 substances, 201 were removed because they did not meet one of the first two criteria, and 155 were removed because they did not meet the third. The final inventory of substances contained 1,767 unique compounds.

Two separate ChemACE approaches were compared for clustering of the chemical inventory. The restrictive clustering approach, in which all compounds in a cluster contain all of the same fragments and no different fragments, resulted in 208 clusters. The less restrictive approach employed the following rules for remapping the chemical inventory:

- treat adjacent halogens as equivalent, allowing fluorine (F) to be substituted for chlorine (Cl), Cl for bromine (Br), Br for iodine (I);
- allow methyl, methylene, and methane to be equivalent;
- allow primary, secondary, and tertiary amines to be equivalent; and
- exclude aromatic thiols (removes thiols from consideration).

Clustering using the less restrictive approach (Pass 2) resulted in 284 clusters. ChemACE results for clustering of the target chemical within the clusters of the chemical inventory are described in Appendix C.

Analogue Searches in the OECD QSAR Toolbox (Dice Method)

The OECD QSAR Toolbox (Version 4.1) is used to search for additional structural analogues of the target compound. There are several structural similarity score equations available in the toolbox (Dice, Tanimoto, Kulczynski-2, Ochiai/Cosine, and Yule). Dice is considered the default equation. The specific options that are selected for the performance of this search include a comparison of molecular features (atom-centered fragments) and atom characteristics (atom type, count hydrogens attached, and hybridization). Chemicals identified in these similarity searches are selected if their similarity scores exceeded 50%.

The OECD QSAR Toolbox Profiler is used to identify those structural analogues from the Dice search that have carcinogenicity and/or genotoxicity data. Nine databases in the OECD QSAR Toolbox (Version 4.1) provide data for carcinogenicity or genotoxicity (see Table B-1).

Analogue search results for the target chemical are described in Appendix C.

| Table B-1. Databases Providing Carcinogenicity and Genotoxicity Data in the OECD |
|--|
| QSAR Toolbox (Version 4.1) |

| Database Name | Toolbox Database Description ^a | |
|---|---|--|
| CPDB | The CPDB provides access to bioassay literature with qualitative and quantitative analysis of published experiments from the general literature (through 2001) and from the NCI/NTP (through 2004). Reported results include bioassays in rats, mice, hamsters, dogs, and nonhuman primates. A calculated carcinogenic potency (TD ₅₀) is provided to standardize quantitative measures for comparison across chemicals. The CPDB contains 1,531 chemicals and 3,501 data points. | |
| ISSCAN | The ISSCAN database provides information on carcinogenicity bioassays in rats and mice reported in sources including NTP, CPDB, CCRIS, and IARC. This database reports a carcinogenicity TD ₅₀ . There are 1,149 chemicals and 4,518 data points included in the ISSCAN database. | |
| ECHA CHEM | The ECHA CHEM database provides information on chemicals manufactured or imported in Europe from registration dossiers submitted by companies to ECHA to comply with the REACH Regulation framework. The ECHA database includes 9,229 chemicals with almost 430,000 data points for a variety of endpoints including carcinogenicity and genotoxicity. ECHA does not verify the information provided by the submitters. | |
| ECVAM Genotoxicity and Carcinogenicity | The ECVAM Genotoxicity and Carcinogenicity database provides genotoxicity and carcinogenicity data for Ames positive chemicals in a harmonized format. ECVAM contains in vitro and in vivo bacteria mutagenicity, carcinogenicity, CA, CA/aneuploidy, DNA damage, DNA damage and repair, mammalian culture cell mutagenicity, and rodent gene mutation data for 744 chemicals and 9,186 data points. | |
| ISSCTA | ISSCTA provides results of four types of in vitro cell transformation assays including Syrian hamster embryo cells, mouse BALB/c 3T3, mouse C3H/10T1/2, and mouse Bhas 42 assays that inform nongenotoxic carcinogenicity. ISSCTA consists of 352 chemicals and 760 data points. | |
| Bacterial mutagenicity ISSSTY | The ISSSTY database provides data on in vitro <i>Salmonella typhimurium</i> Ames test mutagenicity (positive and negative) taken from the CCRIS database in TOXNET. The ISSSTY database provides data for 7,367 chemicals and 41,634 data points. | |
| Genotoxicity OASIS | The Genotoxicity OASIS database provides experimental results for mutagenicity results from "Ames tests (with and without metabolic activation), in vitro chromosomal aberrations and MN and MLA evaluated in vivo and in vitro, respectively." The Genotoxicity OASIS database consists of 7,920 chemicals with 29,940 data points from 7 sources. | |
| Micronucleus OASIS | The Micronucleus OASIS database provides experimental results for in vivo bone marrow and peripheral blood MNT CA studies in blood erythrocytes, bone marrow cells, and polychromatic erythrocytes of humans, mice, rabbits, and rats for 557 chemicals. | |

| Table B-1. Databases Providing Carcinogenicity and Genotoxicity Data in the OECDQSAR Toolbox (Version 4.1) | | | |
|--|---|--|--|
| | | | |
| Database Name | Toolbox Database Description ^a | | |

| Database I valle | Toobox Database Description | |
|------------------|---|--|
| ISSMIC | The ISSMIC database provides data on the results of in vivo MN mutagenicity assa | |
| | to detect CAs in bone marrow cells, peripheral blood cells, and splenocytes in mice and | |
| | rats. Sources include TOXNET, NTP, and the Leadscope FDA CRADA Toxicity | |
| | Database. The ISSMIC database includes data for 563 chemicals and 1,022 data points. | |

^aDescriptions were obtained from the OECD QSAR Toolbox documentation [Version 4.1; <u>OECD (2018)</u>].

CA = chromosomal aberration; CCRIS = Chemical Carcinogenesis Research Information System; CPBD = Carcinogenic Potency Database; CRADA = Cooperative Research and Development Agreement; DNA = deoxyribonucleic acid; ECHA = European Chemicals Agency; ECVAM = European Centre for the Validation of Alternative Methods; FDA = Food and Drug Administration; IARC = International Agency for Research on Cancer; ISSCAN = Istituto Superiore di Sanità Chemical Carcinogen; ISSCTA = Istituto Superiore di Sanità Cell Transformation Assay; ISSMIC = Istituto Superiore di Sanità Micronucleus; ISSSTY = Istituto Superiore di Sanità *Salmonella typhimurium*; MLA = mouse lymphoma gene mutation assay; MN = micronuclei; MNT = micronucleus test; NCI = National Cancer Institute; NTP = National Toxicology Program; OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; REACH = Registration, Evaluation, Authorisation and Restriction of Chemicals; TD₅₀ = median toxic dose.

STEPS 2–5. ANALOGUE REFINEMENT AND SUMMARY OF EXPERIMENTAL DATA FOR GENOTOXICITY, TOXICOKINETICS, CARCINOGENICITY, AND MODE OF ACTION

The outcome of the Step 1 analogue identification process using ChemACE and the OECD QSAR Toolbox is an initial list of structural analogues with genotoxicity and/or carcinogenicity data. Expert judgment is applied in Step 2 to refine the list of analogues based on physicochemical properties; absorption, distribution, metabolism, and excretion (ADME); and mechanisms of toxicity. The analogue refinement process is chemical specific and is described in Appendix C. Steps 3, 4, and 5 (summary of experimental data for genotoxicity, toxicokinetics, carcinogenicity, and mode of action [MOA]) are also chemical specific (see Appendix C for further details).

STEP 6. STRUCTURAL ALERTS AND STRUCTURE-ACTIVITY RELATIONSHIP PREDICTIONS FOR 1-BROMO-2-CHLOROETHANE AND ANALOGUES

Structural alerts (SAs) and predictions for genotoxicity and carcinogenicity are identified using six freely available structure-based tools (described in Table B-2).

| Genotoxicity | | | |
|---------------------------------------|--|--|--|
| Name | Description ^a | | |
| OECD QSAR Toolbox (Version 4.1) | Seven OECD QSAR Toolbox profiling methods were used, including: Carcinogenicity (genotox and nongenotox) alerts by ISS (Version 2.3); updated version of the module originally implemented in Toxtree. It is a decision tree for estimating carcinogenicity, based on 55 SAs (35 from the Toxtree module and 20 newly derived). DNA alerts for Ames by OASIS (Version 1.4); based on the Ames mutagenicity TIMES model, uses 85 SAs responsible for interaction of chemicals with DNA. DNA alerts for CA and MNT by OASIS (Version 1.1); based on the DNA reactivity of the CAs TIMES model, uses 85 SAs for interaction of chemicals with DNA. In vitro mutagenicity (Ames test) alerts by ISS (Version 2.3); based on the Mutagenicity, based on a list of 43 SAs relevant for the investigation of chemical genotoxicity via DNA adduct formation. In vivo mutagenicity (MN) alerts by ISS (Version 2.3); based on the ToxMic rulebase in Toxtree. The rule base has 35 SAs for in vivo MN assay in rodents. OncoLogic Primary Classification (Version 4.0); "developed by LMC and OECD to mimic the structural criteria of chemical classes of potential carcinogens covered by the U.S. EPA's OncoLogic Cancer Expert System for Predicting the Carcinogenicity. It is applicable to organic chemicals with at least one of the 48 alerts specified. Protein binding alerts for CAs by OASIS (Version 1.3); based on 33 SAs for interactions with specific proteins including topoisomerases, cellular protein adducts, etc. | | |
| OncoLogic (Version 7) | OncoLogic is a tool for predicting the potential carcinogenicity of chemicals based on the application of rules for SAR analysis, developed by experts. Results may range from "low" to "high" concern level. | | |
| ToxAlerts | | | |

| Name | Description ^a |
|-----------------------------|--|
| ToxRead (Version 0.9) | ToxRead is a tool designed to assist in making read-across evaluations reproducible. SAs for mutagenicity are extracted from similar molecules with available experimental data in its database. Five similar compounds were selected for this project. The rule sets included: Benigni/Bossa as available in Toxtree (Version 1) SARpy rules extracted by Politecnico di Milano, with the automatic tool SARpy IRFMN rules extracted by human experts CRS4 rules extracted by CRS4 with automatic tools |
| Toxtree (Version 2.6.13) | Toxtree estimates toxic hazard by applying a decision tree approach. Chemicals were queried in Toxtree using the Benigni/Bossa rulebase for mutagenicity and carcinogenicity. If a potential carcinogenic alert based on any QSAR model or if any SA for genotoxic and nongenotoxic carcinogenicity was reported, then the prediction was recorded as a positive carcinogenicity prediction for the test chemical. The output definitions from the tool manual are listed below: SA for genotoxic carcinogenicity (recognizes the presence of one of more SAs and specifies a genotoxic mechanism) SA for nongenotoxic carcinogenicity (recognizes the presence of one or more SAs, and specifies a nongenotoxic mechanism) Potential <i>Salmonella typhimurium</i> TA100 mutagen based on QSAR Unlikely to be a <i>S. typhimurium</i> TA100 mutagen based on QSAR Potential carcinogen based on QSAR (assigned according to the output of QSAR8 aromati amines) Unlikely to be a carcinogen based on QSAR (assigned according to the output of QSAR8 aromatic amines) Negative for genotoxic carcinogenicity (no alert for genotoxic carcinogenicity) Negative for nongenotoxic carcinogenicity (no alert for nongenotoxic carcinogenicity) |

| Name | Description ^a | | |
|------|---|--|--|
| VEGA | VEGA applies several QSARs to a given chemical, as described below: Mutagenicity (Ames test) CONSENSUS model: a consensus assessment is performed based on predictions of the VEGA mutagenicity models (CAESAR, SARpy, ISS, and <i>k</i>-NN) Mutagenicity (Ames test) model (CAESAR): integrates two models, one is a trained SVM classifier, and the other is for FN removal based on SAs matching Mutagenicity (Ames test) model (SARpy/IRFMN): rule-based approach with 112 rules for mutagenicity and 93 for nonmutagenicity, extracted with SARpy software from the origina training set from the Caesar model; includes rules for both mutagenicity and nonmutagenicity Mutagenicity (Ames test) model (ISS): rule-based approach based on the work of Benigni and Bossa (ISS) as implemented in the software Toxtree Version 2.6 Mutagenicity model (CAESAR): Counter Propagation Artificial neural network developed using data for carcinogenicity in rats extracted from the CPDB database Carcinogenicity model (ISS): built implementing the same alerts Benigni and Bossa (ISS) implemented (ISS): built implementing the same alerts Benigni and Bossa (ISS) as implemented in the software Toxtree Version 2.6 Carcinogenicity model (CAESAR): Counter Propagation Artificial neural network developed using data for carcinogenicity in rats extracted from the CPDB database Carcinogenicity model (IRFMN/Alternative Non-Testing Methods Assessed for REACH Substances [ANTARES]): a set of rules (127 SAs), extracted with the SARpy software from a data set of 1,543 chemicals obtained from the carcinogenicity database of Europear Union-funded project ANTARES Carcinogenicity model (IRFMN/ISSCAN-CGX): based on a set of rules (43 SAs) extracte with the SARpy software from a data set of set of 986 compounds; the data set of carcinogenicity of different species was provided by Kirkland et al. (2005) | | |

^aThere is some overlap between the tools. For example, OncoLogic classification is provided by the QSAR Toolbox, but the prediction is available only through OncoLogic, and alerts or decision trees were used in or adapted from several models (e.g., Benigni and Bossa alerts and Toxtree decision tree).

ANTARES = Alternative Non-Testing Methods Assessed for REACH Substances; CA = chromosomal aberration; CAESAR = Computer-Assisted Evaluation of industrial chemical Substances According to Regulations; CONSENSUS = consensus assessment based on multiple models (CAESAR, SARpy, ISS, and *k*-NN); CPDB = Carcinogenic Potency Database; CRS4 = Center for Advanced Studies, Research and Development in Sardinia; DNA = deoxyribonucleic acid; FN = false negative; IRFMN = Istituto di Ricerche Farmacologiche Mario Negri; ISS = Istituto Superiore di Sanità; ISSCAN-CGX = Istituto Superiore di Sanità Chemical Carcinogen; *k*-NN = *k*-nearest neighbor; LMC = Laboratory for Mathematical Chemistry; MN = micronucleus; MNT = micronucleus test; OCHEM = Online Chemical Monitoring Environment; OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; REACH = Registration, Evaluation, Authorisation and Restriction of Chemicals; SA = structural alert; SAR = structure-activity relationship; SVM = support vector machine; TIMES = The Integrated MARKEL-EFOM System; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

The tool results for the target and analogue compounds are provided in Appendix C.

STEP 7. EVIDENCE INTEGRATION FOR SCREENING EVALUATION OF 1-BROMO-2-CHLOROETHANE CARCINOGENICITY

Available data across multiple lines of evidence from Steps 1–6 (outlined above) are integrated to determine the qualitative level of *concern for potential carcinogenicity* of the target compound (Step 8). In the absence of information supporting carcinogenic portal-of-entry effects, the qualitative level of concern for the target chemical should be considered applicable to all routes of exposure.

Evidence integration for the target compound is provided in Appendix C.

APPENDIX C. RESULTS OF THE SCREENING EVALUATION OF POTENTIAL CARCINOGENICITY

STEP 1. USE OF AUTOMATED TOOLS TO IDENTIFY STRUCTURAL ANALOGUES WITH CARCINOGENICITY AND/OR GENOTOXICITY DATA

U.S. EPA's Chemical Assessment Clustering Engine (ChemACE) clustering was performed as described in Appendix B. Using the most restrictive clustering rules (where ChemACE assigns a unique definition for each fragment, ensuring that each chemical submitted for clustering appears in only one ChemACE cluster), 1-bromo-2-chloroethane appeared in Cluster 67, which did not contain any other compounds. Using the less restrictive clustering option, ChemACE treats adjacent halogens as equivalent fragments, resulting in the inclusion of 1-bromo-2-chloroethane in multiple clusters. Using the more permissive fragment definition, halogens (i.e., bromine [Br] and chlorine [Cl]) in the structures could be swapped in the clustering scheme and appear in more than one cluster. Five clusters (15, 31, 34, 49, and 152) contain 1-bromo-2-chloroethane and a total of 38 other halogenated chemicals (see Table C-1 below). All of these clusters contain halogenated alkanes with chain lengths of 1–3 carbon atoms and 1–6 halogen substituents.

| Table C-1. Clusters Containing 1-Bromo-2-Chloroethane (CASRN 107-04-0) and theAssociated Fragments | | |
|--|---|--|
| Cluster | Fragments | |
| 15 | Cl-R or F-R or Br-R | |
| 31 | Cl-R or F-R or Br-R (but not Br-R only structures) | |
| 34 | Cl-R or F-R or Br-R (but not F-R only structures) | |
| 49 | Cl-R or F-R or Br-R or I-R | |
| 152 | Cl-R or F-R or Br-R (but not Br-R or F-R only structures) | |

Br = bromine; Cl = chlorine; F = fluorine; I = iodine; R = functional group (must be an aliphatic carbon attachment).

The Organisation for Economic Co-operation and Development (OECD) Quantitative Structure-Activity Relationship (QSAR) Toolbox Profiler was used to identify structural analogues from the Dice analogue search with carcinogenicity and/or genotoxicity data (see Step 1 methods in Appendix B). Only one analogue was identified in the Dice analogue search (1-bromo-3-chloropropane); this compound was also identified by ChemACE. Refinement of selection of final analogues is described below.

STEP 2. ANALOGUE REFINEMENT USING EXPERT JUDGMENT

Expert judgment was applied to refine the initial list of 38 potential analogues based on physicochemical properties; absorption, distribution, metabolism, and excretion (ADME); and mechanisms of toxicity.

1-Bromo-2-chloroethane contains two halogen substituents with the following attributes: (1) they are labile (easily displaced) leaving groups that can undergo nucleophilic substitution; (2) they are attached to adjacent carbon atoms (1,2-substitution pattern); and (3) they are both attached at primary carbon atoms (R-CH₂-X). Each of these attributes influences the reactivity of the molecule, its available metabolic pathways, and its potential bioactivity. Therefore, compounds were considered potential analogues if they contain (1) two or more labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom.

Of the 38 chemicals identified as potential analogues by ChemACE clustering and the OECD Toolbox analogue selection tool (Dice), 32 were not selected for further review, including halogenated methanes, monohalogenated compounds, and compounds with more than one halogen per carbon atom. Each of these attributes introduce significant differences in bioavailability, reactivity, and applicable metabolic pathways relative to 1-bromo-2-chloroethane.

The remaining six possible analogues for 1-bromo-2-chloroethane are listed in Table C-2. The existence of a cancer risk estimate and/or a weight-of-evidence (WOE) determination for cancer is indicated for each analogue. All of the potential analogues were included in each of the five clusters, except for 1,2-dibromoethane, which was excluded from Clusters 31 and 152.

| Table C-2. Summary of Cancer Assessment Information for Analogues of1-Bromo-2-Chloroethane (CASRN 107-04-0) ^a | | | |
|--|--|--|--|
| Analogue Name (CASRN) | Cancer Risk Estimates (if available) | WOE Determinations | |
| 1,2-Dibromoethane (106-93-4) ^b | <u>U.S. EPA (2004)</u> —OSF, IUR <u>CalEPA (2011)</u> —OSF, IUR | U.S. EPA (2004)—likely IARC (1999)—probably (Group 2A) NTP (2016c)—reasonably anticipated CalEPA (2018)—known | |
| 1,2-Dichloroethane (107-06-2) ^b | <u>U.S. EPA (1993)</u> —OSF, IUR <u>CalEPA (1999b)</u> —OSF | U.S. EPA (1993)—probable IARC (1999)—possibly (Group 2B) NTP (2016d)—reasonably anticipated CalEPA (2018)—known | |
| 1-Bromo-3-chloropropane (109-70-6) ^{b, c} | None ^d | None | |
| 1,2-Dichloropropane (78-87-5) ^b | <u>U.S. EPA (2020c)</u> —p-OSF, p-IUR <u>CalEPA (2004); CalEPA</u> (1999b)—OSF | U.S. EPA (2020c)—likely IARC (2017)—carcinogenic (Group 1) CalEPA (2018)—known | |
| 1,2-Dibromo-3-chloropropane (96-12-8) ^b | <u>U.S. EPA (2006)</u> —p-OSF, p-IUR <u>CalEPA (2020)</u> —OSF, IUR | U.S. EPA (2006)—likely IARC (1999)—possibly (Group 2B) NTP (2016b)—reasonably anticipated CalEPA (2018)—known | |
| 1,2,3-Trichloropropane (96-18-4) ^b | <u>U.S. EPA (2009)</u> —OSF <u>CalEPA (2009)</u> —OSF | U.S. EPA (2009)—likely IARC (1995)—probably (Group 2A) NTP (2016b)—reasonably anticipated CalEPA (2018)—known | |

^aGray shading indicates that there was not a cancer risk estimate and/or a WOE determination for cancer for that analogue.

^bFound by ChemACE.

^cFound by Dice.

^dNo cancer toxicity values are available; however, a 2-year inhalation bioassay in rats and mice with clear evidence of carcinogenicity is available from the Japanese Industrial Safety and Health Association (Japan Industrial Safety and Health Association, 2005a, b).

ChemACE = Chemical Assessment Clustering Engine; IUR = inhalation unit risk; OSF = oral slope factor; p-IUR = provisional inhalation unit risk; p-OSF = provisional oral slope factor; WOE = weight of evidence.

1-Bromo-3-chloropropane, which lacks a cancer risk estimate or a WOE determination for cancer (highlighted in gray in Table C-2), was not further considered as a potential analogue for the screening evaluation of potential carcinogenicity of 1-bromo-2-chloroethane. Compounds selected for further consideration were 1,2-dibromoethane (1,2-DBE), 1,2-dichloroethane (1,2-DCE), 1,2-dichloropropane (1,2-DCP), 1,2-dibromo-3-chloropropane (1,2-DB-3-CP), and 1,2,3-trichloropropane (TCP).

STEP 3. COMPARISON OF THE EXPERIMENTAL GENOTOXICITY DATA FOR 1-BROMO-2-CHLOROETHANE AND ANALOGUES

The genotoxicity data available for 1-bromo-2-chloroethane are described in the "Other Data" section in the main body of this report. Available data indicate that

1-bromo-2-chloroethane and/or its metabolites display genotoxic, mutagenic, clastogenic, and deoxyribonucleic acid (DNA)-damaging activity. Genotoxicity data for the analogue compounds have been extensively reviewed. A summary of the genotoxicity data for analogue halogenated alkanes is provided in Table C-3. Overall data indicate that these analogues are mutagenic and clastogenic, and capable of binding DNA and causing DNA damage.

| Table C- | 3. Comparison of Av | vailable Genotoxici | ty Data for 1-Bron Analogues | no-2-Chloroethane | e (CASRN 107-04-0 |)) and Candidate |
|--------------|--|---|--|---|--|---|
| Parameter | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloropropane CASRN 96-18-4 |
| Mutagenicity | Mutagenic in Salmonella typhimurium Mutagenic in mammalian cells in vitro | Mutagenic in S. typhimurium and Streptomyces coelicolor; mixed results in Escherichia coli Mutagenic in Aspergillus nidulans Induced somatic and sex-linked recessive mutations in Drosophila melanogaster Mutagenic in mammalian cells in vitro Did not induce dominant lethal mutations in mice | mixed results in E. coli; not mutagenic in S. coelicolor or Bacillus subtilis Not mutagenic in A. nidulans Induced somatic and sex-linked recessive mutations Mutagenic in mammalian cells in vitro | Mixed results in <i>S. typhimurium</i>; not mutagenic in <i>E. coli</i> or <i>S. coelicolor</i> Mutagenic in <i>A. nidulans</i> Induced somatic mutations in <i>D. melanogaster</i>, but not sex-linked recessive mutations Mutagenic in mammalian cells in vitro Did not induce dominant lethal mutations in rats Did not induce <i>Pig-a</i> or <i>Gpt</i> mutations in mice | Mutagenic in <i>S. typhimurium</i> with and without metabolic activation Induced sex-linked recessive mutations in <i>D. melanogaster</i> Mutagenic in mammalian cells with and without metabolic activation Induced somatic mutations in mice (spot test) Induced dominant lethal mutations in rats, but not mice | Mutagenic in <i>S. typhimurium</i> with metabolic activation; not mutagenic in <i>E. coli</i> Not mutagenic in <i>A. nidulans</i> Induced somatic mutations in <i>D. melanogaster</i> Mutagenic in mammalian cells with metabolic activation Did not induce dominant lethal mutations in rats |

| Parameter | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloropropane CASRN 96-18-4 |
|---------------------------|--|--|--|--|--|---|
| Clastogenicity | Induced CAs in mammalian cells in vitro Induced mitotic malsegregation/ aneuploidy in <i>A. nidulans</i> Induced mitotic recombination in <i>D. melanogaster</i> | Induced SCEs in mammalian cells in vitro and in vivo Induced CAs and MNs in mammalian cells in vitro, but not in vivo | vivo • Induced MNs in mammalian cells in | mammalian cells in vitro | Induced CAs in mammalian cells in vitro and in vivo Induced MNs in mammalian cells in vivo Induced SCEs in mammalian cells in vitro Induced heritable translocation and mitotic recombination in <i>D. melanogaster</i> | Induced polyploidy in mammalian cells in vivo Induced MNs in mammalian cells in vitro, but not in vivo Induced CAs and SCEs in mammalian cells in vitro Induced MN without metabolic activation Induced mitotic gene conversion in <i>S. cerevisiae</i> Did not induce chromosomal abnormalities in <i>A. nidulans</i> |
| DNA damage and adducts | Induced DNA damage in mammalian cells in vivo Induced DNA damage in <i>S. typhimurium</i> | damage/repair in mammalian cells in | Induced DNA damage/repair in mammalian cells in vitro and in vivo Equivocal DNA damage detected via comet assay in rodent liver cells. Forms DNA adducts | • Induced DNA damage/repair in mammalian cells in vitro and in vivo | Induced DNA damage/repair in mammalian cells in vitro and in vivo Forms DNA adducts | Induced DNA damage/repair in mammalian cells in vitro and in vivo Forms DNA adducts |

| Table C-3 | 3. Comparison of Av | ailable Genotoxici | ty Data for 1-Bron Analogues | no-2-Chloroethane | (CASRN 107-04-0 |)) and Candidate |
|------------------------|--|--|---|--|--|--|
| Parameter | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloropropane CASRN 96-18-4 |
| Cell transformation | ND | • Induced neoplastic transformation in mammalian cells | • Mixed results for cell transformation | ND | • Induced neoplastic transformation in mammalian cells | • Induced neoplastic transformation in mammalian cells |
| References | See "Genotoxicity" section in main document for references | <u>U.S. EPA (2004);</u> <u>IARC (1999)</u> | Gwinn et al. (2011); ATSDR (2001); CalEPA (1999a); IARC (1999); U.S. EPA (1993); Lebaron et al. (2021) | <u>IARC (2017); U.S.</u> <u>EPA (2020c);</u> <u>CalEPA (1999b)</u> | <u>U.S. EPA (2006);</u> <u>Clark and Snedeker</u> (2005); <u>IARC (1999)</u> | <u>U.S. EPA (2009); IARC (1995)</u> |

Г

CA = chromosomal aberration; DNA = deoxyribonucleic acid; MN = micronuclei; ND = no data; SCE = sister chromatid exchange.

STEP 4. TOXICOKINETICS OF 1-BROMO-2-CHLOROETHANE AND ANALOGUES

The toxicokinetics of 1-bromo-2-chloroethane and potential analogues are briefly described in Table C-4. Each of the candidate analogue compounds is metabolized by a CYP450 oxidation pathway and a direct glutathione (GSH) conjugation pathway (<u>U.S. EPA, 2020c;</u> <u>IARC, 2017; CalEPA, 2009; U.S. EPA, 2009, 2006, 2004; ATSDR, 2001; IARC, 1999, 1995</u>). Contributions from both pathways are presumed by analogy for 1-bromo-2-chloroethane as well, although experimental data for the target compound are available for the GSH conjugation pathway only (Jean and Reed, 1992; Marchand and Reed, 1989).

Experimental data show that GSH conjugation leads to the formation of episulfonium ions via cyclization of GSH adducts for 1,2-DBE, 1,2-DCE, 1,2-DB-3-CP, and TCP (<u>CalEPA</u>, 2009; U.S. EPA, 2009, 2004; ATSDR, 2001; IARC, 1995; <u>Guengerich</u>, 1994; <u>van Beerendonk et al.</u>, 1994; <u>Dekant and Vamvakas</u>, 1993; <u>Pearson et al.</u>, 1990; <u>Dohn et al.</u>, 1988; <u>Omichinski et al.</u>, 1988b; <u>Omichinski et al.</u>, 1988a; <u>Guengerich et al.</u>, 1987). Based on analogy and identified glutathione-conjugated metabolites, episulfonium ion generation is also expected following GSH conjugation of 1-bromo-2-chloroethane (based on formation of *S*-[2-chloroethyl]glutathione [CEG]) (Jean and Reed, 1992; <u>Marchand and Reed, 1989</u>) and 1,2-DCP (based on the formation of *S*-[2-hydroxypropyl]glutathione) (IARC, 2017).

No data are available for absorption, distribution, or excretion of 1-bromo-2-chloroethane. Experimental data indicate that all candidate analogues show rapid and extensive absorption, wide distribution, and primary excretion in urine (U.S. EPA, 2020c; IARC, 2017; CalEPA, 2009; U.S. EPA, 2009, 2006, 2004; ATSDR, 2001; IARC, 1999, 1995).

| | Table C-4. Summary of Toxicokinetic Data for 1-Bromo-2-Chloroethane and Analogues | | | | | | |
|---|---|--|---|--|--|--|--|
| Compound (CASRN)Absorption, Distribution, Excretion1-Bromo- 2-chloroethane (107-04-0)ND | | Metabolism | References | | | | |
| | | Primary metabolic pathways are oxidation by CYP450 (by analogy to other dihaloalkanes) and GSH conjugation (demonstrated) 1. Primary metabolites formed via oxidation expected to be: 2-bromoacetaldehyde 2-chloroacetaldehyde HBr HCl 2. Primary metabolites formed via GSH conjugation: CEG, with HBr release (minor HCl release) Inferred from other dihaloalkanes: CEG will cyclize to form episulfonium ion | Guengerich (1994); Dekant and Vamvakas (1993); Jean and Reed (1992); Gargas et al. (1989); Marchand and Reed (1989); Guengerich et al. (1987) | | | | |
| 1,2-Dibromoethane (106-93-4) | Rapid and extensive absorption (oral and inhalation) Wide distribution Primary excretion in urine, small amounts in feces and exhaled air | Primary metabolic pathways are oxidation by CYP450 and GSH conjugation 1. Primary metabolites formed via oxidation: 2-bromoacetaldehyde HBr Primary metabolites formed via GSH conjugation: BEG, with HBr release Episulfonium ion (from cyclization of GSH adduct in BEG) | <u>U.S. EPA (2004); Gargas</u> et al. (1989); <u>Guengerich et</u> al. (1987) | | | | |
| 1,2-Dichloroethane (107-06-2) | Rapid and extensive absorption (oral and inhalation) Wide distribution Primary excretion in urine, small amounts in feces and exhaled air | Primary metabolic pathways are oxidation by CYP450 and GSH conjugation 1. Primary metabolites formed via oxidation: 2-chloroacetaldehyde HCl Primary metabolites formed via GSH conjugation: CEG, with HCl release Episulfonium ion (from cyclization of GSH adduct in CEG) | <u>ATSDR (2001); Gargas et</u> <u>al. (1989); Guengerich et</u> <u>al. (1987)</u> | | | | |

| Compound (CASRN) | Absorption, Distribution, Excretion | oxicokinetic Data for 1-Bromo-2-Chloroethane and Analog Metabolism | References |
|--|--|---|---|
| 1,2-Dichloropropane (78-87-5) | Readily absorbed (oral, inhalation, dermal) Wide distribution, with preferential disposition in body fat at high exposures Rapid elimination, primarily via urine with small amounts via exhaled breath (proportion increases with increased exposure) | Primary metabolic pathway is oxidation by CYP2E1 followed by GSH conjugation (or vice versa) to form: S-(2-oxopropyl)glutathione (identified) S-(1-carboxyethyl)glutathione (identified) S-(2-hydroxypropyl)glutathione (presumed based on cysteine-conjugated urinary metabolite) Presumed secondary pathways (based on identified metabolites and known metabolism of similar haloalkanes) Formation of episulfonium ions from cyclization of GSH adduct in <i>S</i>-(2-hydroxypropyl)glutathione Oxidative dechlorination, leading to formation of lactate and release of carbon dioxide | <u>IARC (2017); U.S. EPA</u> (2020c) |
| 1,2-Dibromo- 3-chloropropane (96-12-8) | Rapid and extensive absorption (oral; no inhalation data) Wide distribution Most excretion in urine, some in exhaled air, small amounts in feces | Primary metabolic pathways are oxidation by CYP450 and GSH conjugation Primary metabolites formed via oxidation: 2-chloro-3-(bromomethyl)oxirane 1-bromo-3-chloroacetone HBr HCl Primary metabolites formed via GSH conjugation: CBPG, with HBr release (minor HCl release) Episulfonium ion (from cyclization of GSH adduct in CBPG) | Guengerich (1994); van Beerendonk et al. (1994); Dekant and Vamvakas (1993); Humphreys et al. (1991); Pearson et al. (1990); Dohn et al. (1988); Omichinski et al. (1988a); Omichinski et al. (1988b); Gingell et al. (1987); Guengerich et al. (1987) |

| | Table C-4. Summary of T | Coxicokinetic Data for 1-Bromo-2-Chloroethane and Analog | jues |
|-------------------------------------|---|--|--|
| Compound (CASRN) | Absorption, Distribution, Excretion | Metabolism | References |
| 1,2,3-Trichloropropane (96-18-4) | Rapid and extensive absorption (oral; no inhalation data) Rapid distribution, initially to adipose tissue, liver, and kidney Primary excretion in urine, smaller amounts in feces and exhaled air | Primary metabolic pathways are oxidation by CYP450 and GSH conjugation Primary metabolites formed via oxidation: 1,3-dichloroacetone 2,3-dichloropropanal Chloroacrolein HCl Primary metabolites formed via GSH conjugation: B-chlorothioether with HCl release Episulfonium ion (from cyclization of GSH adduct in thioether) | <u>CalEPA (2009); U.S. EPA</u> (2009); <u>IARC (1995)</u> |

BEG = S-(2-bromoethyl)glutathione; CBPG = S-(3-chloro-2-bromopropyl) glutathione; CEG = S-(2-chloroethyl)glutathione; CYP450 = cytochrome P450; GSH = glutathione; HBr = hydrogen bromide; HCl = hydrogen chloride; ND = no data.

STEP 5. CARCINOGENICITY OF 1-BROMO-2-CHLOROETHANE ANALOGUES AND MODE-OF-ACTION DISCUSSION

U.S. EPA cancer WOE descriptors for 1-bromo-2-chloroethane and analogue compounds are shown in Tables C-5 (oral) and C-6 (inhalation). As noted in the PPRTV document, there is *"Inadequate Information to Assess the Carcinogenic Potential"* of 1-bromo-2-chloroethane. All analogues are characterized as having evidence of carcinogenic potential. 1,2-DBE, 1,2-DCP, 1,2-DB-3-CP, and TCP were all classified as *"Likely to be Carcinogenic to Humans"* (U.S. EPA, 2020c, 2009, 2006, 2004) under the 2005 or 1999 *Guidelines for Carcinogen Risk Assessment* (U.S. EPA, 2005). 1,2-DCE is considered a *"Probable Human Carcinogen"* (U.S. EPA, 1993) under the 1986 *Guidelines for Carcinogen Risk Assessment* (U.S. EPA, 1986). OSF values varied by several orders of magnitude, with the highest potency value calculated for TCP and the lowest (provisional) potency value for 1,2-DCP. IUR values also varied by several orders of magnitude, with the highest potency value for 1,2-DCP (provisional) and the lowest potency value for 1,2-DCP (provisional).

Several epidemiology studies in print shop workers in Japan reported a potential correlation between exposure to 1,2-DCP (and other solvents) and cholangiocarcinoma, a rare cancer of the bile duct (U.S. EPA, 2020c; IARC, 2017). However, the population size is small in most of these studies, and the workers were exposed to numerous other solvents, including dichloromethane and 1,1,1-trichloroethane, as well as kerosene and printing ink, confounding interpretation of the results. Some human epidemiology studies for 1,2-DCE, 1,2-DBE, and 1,2-DB-3-CP have found limited evidence of increased risk of certain cancers (e.g., lymphatic, hematopoietic, pancreatic, and stomach cancer); however, due to multiple exposures in the available studies, these data are inadequate to evaluate a relationship between human cancer and exposure to a specific compound (NTP, 2016b, c, d; U.S. EPA, 2006, 2004; IARC, 1999). No human cancer data are available for TCP (NTP, 2016e; U.S. EPA, 2009).

All analogues are associated with tumor induction in animal bioassays (U.S. EPA, 2020c; <u>IARC, 2017; NTP, 2016b, c, d, e; U.S. EPA, 2006, 2004; IARC, 1999, 1995; U.S. EPA, 1993</u>). These are briefly summarized below and in Table C-5 (oral) and Table C-6 (inhalation).

Common target organs and systems in rats and/or mice following exposure to analogues include the liver, respiratory system, and mammary gland. All analogues induced hepatocellular neoplasms (adenomas and/or carcinomas) following chronic oral exposure; liver tumors were also increased following chronic inhalation exposure to 1,2-DCE. All analogues tested via the inhalation route also induced respiratory tract tumors, including nasal cavity tumors (1,2-DBE, 1,2-DCP, 1,2-DB-3-CP) and lung tumors (1,2-DBE, 1,2-DCE, 1,2-DCP, 1,2-DB-3-CP). Lung tumors were also observed following oral exposure to 1,2-DBE and 1,2-DCE. Mammary gland tumors were increased following exposure to analogues via the oral route (1,2-DCE, 1,2-DCP, 1,2-DCP

The gastrointestinal, endocrine, and circulatory systems were also identified as cancer targets for the majority of analogues. Most analogues induced tumors of the gastrointestinal system following chronic oral exposure, including forestomach tumors (1,2-DBE, 1,2-DCE, 1,2-DB-3-CP, TCP); tongue, pharynx, and stomach tumors (1,2-DB-3-CP); oral cavity and alimentary system tumors (TCP); and esophageal papillomas (1,2-DBE). Endocrine tumors observed included thyroid tumors (1,2-DBE, 1,2-DCP) and pancreatic tumors (TCP) following

oral exposure and adrenal tumors (1,2-DBE and 1,2-DB-3-CP) following inhalation exposure. Hemangiosarcoma, most notably in the spleen, was observed following oral or inhalation exposure to 1,2-DBE, oral exposure to 1,2-DCE, and inhalation exposure to 1,2-DCP.

Other tumor types were identified in only one or two of the analogues and therefore do not represent common targets of the identified analogue chemicals (see Tables C-5 and C-6 for more details). Observed tumors included female reproductive tumors (1,2-DCE, TCP), renal tumors (TCP, 1,2-DB-3-CP), malignant lymphoma (1,2-DCE), Harderian gland tumors (TCP, 1,2-DCP), Zymbal gland tumors (TCP), preputial gland tumors (TCP), mesothelioma of the peritoneum (1,2-DCE), mesothelioma of the tunica vaginalis (1,2-DBA, 1,2-DB-3-CP), and subcutaneous tumors (1,2-DCE).

U.S. EPA (2009) and U.S. EPA (2006) concluded that 1,2-DB-3-CP and TCP are carcinogenic through a mutagenic MOA, supported by experimental evidence of mutagenicity. The U.S. EPA (2004) report did not make a formal determination regarding a carcinogenic MOA for 1,2-DBE; however, available evidence suggested potential mutagenicity and/or DNA damage as potential MOAs. Available data for 1,2-DCP were considered inadequate for a formal MOA analysis by U.S. EPA (2020c); however, DNA damage has been proposed as a potential mechanism. The U.S. EPA (1993) report did not evaluate potential carcinogenic MOAs for 1,2-DCE, but experimental data indicate that the chemical is both DNA damaging and mutagenic (see Step 3). Taken together, all analogues except 1,2-DCP have sufficient evidence of mutagenicity (see Step 3) to indicate a potential common mutagenic MOA for halogenated alkanes. Specifically, experimental data show that 1,2-DBE, 1,2-DCE, 1,2-DB-3-CP, and TCP form DNA-reactive episulfonium ions via cyclization of GSH adducts during metabolism. By analogy, both 1,2-DCP and the target compound, 1-bromo-2-chloroethane, are expected to form episulfonium ions as well (see Table C-4). Data for 1,2-DCE indicate that although oxidative metabolites appear to influence CAs, the formation of the episulfonium ion is likely the primary mutagen (Gwinn et al., 2011). However, 1,2-DCP is not considered a potent mutagen, and experimental data do not support the potential for GSH-mediated mutagenicity (U.S. EPA, 2020c; Akiba et al., 2017; IARC, 2017).

| Table C- | 5. Comparison of | f Available Oral Ca | arcinogenicity Data Candidate Ana | | lloroethane (CASR) | N 107-04-0) and |
|--|--|--|--------------------------------------|--|--|---|
| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 |
| Structure | CI | Br | CICI | CI CI | Cl Br Br | CI CI CI |
| U.S. EPA WOE characterization | "Inadequate Information to Assess Carcinogenic Potential" (see Table 6) | "Likely to Be Carcinogenic to Humans" | "Probable Human Carcinogen" | "Likely to Be Carcinogenic to Humans" | "Likely to Be Carcinogenic to Humans" | "Likely to Be Carcinogenic to Humans" |
| OSF (mg/kg-d) ⁻¹ | NDr | 2 (upper bound) ^a | 9.1 × 10 ⁻² | 3.7×10^{-2} (provisional) | 8×10^{-1} (provisional) | 3×10^1 (upper bound) ^b |
| Data set(s) used for slope factor derivation | NA | Forestomach tumors, hemangiosarcomas, and thyroid follicular cell adenomas or carcinomas in male rats | Hemangiosarcoma in male rats | Hepatocellular adenoma or carcinoma in male mice | Renal adenoma or carcinoma in male rat | Alimentary system, liver, Harderian gland, and uterine tumors in female mice |

| Table C-5. Comparison of Available Oral Carcinogenicity Data for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Candidate Analogues | | | | | | | | | |
|---|--|---|--|---|--|---|--|--|--|
| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 | | | |
| Other tumors observed in animal oral bioassays | ND | Rat: forestomach, thyroid, or liver tumors and hemangiosarcomas (females) Mouse: forestomach and lung tumors, esophageal papilloma (female) | Rat: forestomach and mammary gland tumors Mouse: lung, liver, and uterus tumors; malignant lymphoma | Rat: mammary gland tumors (marginal increase in females) Mouse: liver tumors (female), thyroid tumors (female) | Rat: mammary gland and kidney tumors (female), liver tumors (male), forestomach and stomach tumors (both) Mouse: forestomach and stomach tumors | Rat: alimentary system and Zymbal gland tumors (both); pancreatic, preputial gland, and kidney tumors (males); clitoral and mammary gland tumors (females) Mouse: alimentary system, Harderian gland, and liver tumors (males) | | | |
| Study doses (mg/kg-d) | NA | TWA: 0, 38, 41 HED ^b : 0, 11, 22 | TWA: 0, 47, 95 HED: 0, 4.46, 8.23 | TWA: 0, 89.3, 179 HED: 0, 12.5, 25.1 | TWA: 0, 0.24, 0.80, 2.39 HED: not reported per dose (HED conversion done after OSF calculation) | 0, 6, 20, 60 HED: not reported per dose (HED conversion done after BMD modeling) | | | |
| Route (method) | NA | Gavage in corn oil (5 d/wk) | Gavage in corn oil (5 d/wk) | Gavage in corn oil (5 d/wk) | Diet | Gavage | | | |
| Duration | NA | 49 wk | 78 wk, followed by untreated observation period up to 32 wk | 103 wk | 104 wk | 2 yr; all high-dose females sacrificed after 73 wk | | | |
| POD type | NA | BMDL ₁₀ (HED) | BMDL (linearized multistage with time- to-death analysis, extra risk) | BMDL ₁₀ (HED) | BMDL ₁₀ (HED) | BMDL ₁₀ (HED) | | | |

| Table C- | Table C-5. Comparison of Available Oral Carcinogenicity Data for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Candidate Analogues | | | | | | | |
|--------------|---|-------------------------------------|--|--------------------------------------|--|--|--|--|
| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 | | |
| Source | NA | NTP (2016c); U.S. EPA (2004) | NTP (2016d); U.S. EPA (2020c); U.S. EPA (2010) | <u>U.S. EPA (2020c)</u> | <u>U.S. EPA (2006)</u> | <u>U.S. EPA (2009)</u> | | |

^aAn upper bound on cancer risk was estimated by adding the central tendency risk estimates for the three tumor types and calculating an upper confidence limit on the sum, using an estimate of the variance pooled across the three slope factors. The resulting (upper bound) slope factor was adjusted for daily exposure by multiplying by 5 days/7 days and for lifetime exposure by dividing by (49 weeks/104 weeks).

^bHED values were reported in the modeling section of <u>U.S. EPA (2004)</u> and appear to correlate with initial doses of 40 and 80 mg/kg-day, not TWA doses that reflect lowered dose in the high-dose group.

BMD = benchmark dose; BMDL = benchmark dose lower confidence limit; $BMDL_{10} = 10\%$ benchmark dose lower confidence limit; HED = human equivalent dose; NA = not applicable; ND = no data; NDr = not determined; OSF = oral slope factor; POD = point of departure; TWA = time-weighted average; WOE = weight of evidence.

| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloro- propane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 |
|--|---|--|---|---|--|--|
| Structure | CI | Br | ci~~ ^{Ci} | CI CI | Cl Br | CI CI |
| WOE characterization | "Inadequate Information to Assess Carcinogenic Potential" (see Table 6) | "Likely to Be Carcinogenic to Humans" | "Probable Human Carcinogen" | "Likely to Be Carcinogenic to Humans" | "Likely to Be Carcinogenic to Humans" | "Likely to Be Carcinogenic to Humans" |
| IUR $(\mu g/m^3)^{-1}$ | NDr | 6×10^{-4} (upper bound) ^a | $2.6 \times 10^{-5 \text{ b}}$ $1.0 \times 10^{-6} (95\%$ upper bound) ^c | 3.7×10^{-6} (provisional) | 6 × 10 ⁰ (provisional) | NDr |
| Data set(s) used for slope factor derivation | NA | Nasal tumors (male and female rats, female mice), hemangiosarcomas (male and female rats, female mice), mesothelioma (male rats), lung tumors (female rats and mice), mammary tumors (female rats and mice), fibrosarcoma (female mice) | Hemangiosarcoma (oral study) ^a | Nasal cavity tumors in male rats | Nasal cavity tumors in male rats | NA |

г

| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | -04-0) and Candid 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloro- propane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 |
|--|--|---|--|--|---|--|
| Human carcinogenicity data | ND | Available occupational data inadequate to evaluate relationship between 1,2-DBE and cancer | Available occupational data inadequate to evaluate relationship between 1,2-DCE and cancer | Elevated risk of cholangiocarcinoma in print shop workers from several cohorts exposed to 1,2-DCP (and other chlorinated solvents) | Available occupational data inadequate to evaluate relationship between 1,2-DB-3-CP and cancer; no associations between gastric cancer or leukemia and 1,2-DB-3-CP in drinking water (case-control) | ND |
| Other tumors observed in animal inhalation bioassays | ND | Rat: splenic hemangiosarcoma, adrenal tumors, subcutaneous mesenchymal tumor | Inhalation study published since derivation of IUR: Rat: mammary gland tumors, peritoneal mesothelioma, subcutaneous fibroma Mouse: lung, liver, and mammary gland tumors (females) | Rats: nasal cavity tumors (female) Mice: Harderian gland tumors (male), lung tumors (female), hemangiocarcinoma (male) | Rats: nasal cavity, pharynx, adrenal gland, and mammary gland tumors (females), tongue tumors (both), mesothelioma (males) Mice: nasal and lung tumor (both) | ND |

| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloro- propane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 |
|----------------------|--|---|--|--|---|--|
| Study concentrations | NA | Reported (ppm): 0, 10, 40 | TWA (mg/kg-d): 0, 47, 95 HED (mg/kg-d): 0, | Analytical (ppm): 0, 80.2, 200.5, 500.2 | Reported (ppm): 0, 0.6, 3 | NA |
| | | HEC _{ER} (ppm): 0, 1.8, 7.1 (rat, mouse) | 4.46, 8.23 | HEC _{ET} (mg/m ³): 0, 16.2, 40.54, 101.1 | HEC _{ET} (mg/m ³): 0, 0.23, 1.13 | |
| | | HEC _{ET} (ppm): 0, 0.36, 1.42 (male rat, female mouse); 0, 0.25, 0.99 (female rat) | | | | |
| | | HEC _{PU} (ppm) 0, 3.1, 12.3 (female rat); 0, 5.8, 22.7 (female mouse) | | | | |
| | | HEC _{TB} (ppm) 0, 4.86, 19.2 (female mouse) | | | | |
| Route (method) | NA | Inhalation (6 h/d; 5 d/wk) | Gavage in corn oil (5 d/wk) ^b | Inhalation (6 h/d; 5 d/wk) | Inhalation (6 h/d; 5 d/wk) | NA |
| Duration | NA | 79–104 wk | 78 wk, plus untreated observation period up to 32 wk | 104 wk | Control: 105–107 wk; 0.6 ppm: 103 wk plus 1-wk observation; 3 ppm: 84 wk plus 0–1-wk observation | NA |

| Table C-6. Comparison of Available Inhalation Carcinogenicity Toxicity Data for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Candidate Analogues | | | | | | | |
|--|--|---|--|--|---|---|--|
| Type of Data | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloro- propane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 | |
| POD type | NA | BMCL ₁₀ (HEC) | BMDL (linearized multistage with time- to-death analysis, extra risk) | BMCL ₁₀ (HEC) | BMCL ₁₀ (HEC) | NA | |
| Source | NA | <u>NTP (2016c); U.S.</u> <u>EPA (2004)</u> | NTP (2016d); U.S. EPA (2010); U.S. EPA (1993) | <u>IARC (2017); U.S.</u> <u>EPA (2020c)</u> | <u>NTP (2016b); U.S.</u> <u>EPA (2006); IARC</u> (1995) | <u>NTP (2016e); U.S.</u> <u>EPA (2009)</u> | |

^aAn upper bound on cancer risk was estimated by using the multistage model with Poly-3 adjusted incidence data central tendency estimate for tumors at six sites in two species.

^bAvailable inhalation data in 1987 was inadequate to derive an IUR (no induction of tumors); therefore, data from oral studies used to derive IUR in 1987 (the same major urinary metabolites and concentration from both oral and inhalation exposure) (<u>Reitz et al., 1982</u>). The IUR was calculated from oral data, assuming 100% absorption and metabolism at the low dose. See findings from oral study in Table C-5. ^cInferred unit risk from negative inhalation study (Maltoni et al., 1980).

"Interred unit risk from negative innalation study (<u>Maltoni et al., 1980</u>).

1,2-DB-3-CP = 1,2-dibromo-3-chloropropane; 1,2-DBE = 1,2-dibromoethane; 1,2-DCE = 1,2-dichloropethane; 1,2-DCP = 1,2-dichloropropane; BMCL₁₀ = 10% benchmark concentration lower confidence limit; BMDL = benchmark dose lower confidence limit; ER = extrarespiratory; ET = extrathoracic; HEC = human equivalent concentration; HED = human equivalent dose; IUR = inhalation unit risk; NA = not applicable; ND = no data; NDr = not determined; POD = point of departure; PU = pulmonary; TB = tracheobronchial; TWA = time-weighted average; WOE = weight of evidence.

STEP 6. STRUCTURAL ALERTS AND STRUCTURE-ACTIVITY RELATIONSHIP PREDICTIONS FOR 1-BROMO-2-CHLOROETHANE AND ANALOGUES

Structural alerts (SAs) and predictions for genotoxicity and carcinogenicity were identified using computational tools as described in Appendix B. The model results for 1-bromo-2-chloroethane and its analogue compounds are shown in Table C-7. Concerns for carcinogenicity and/or mutagenicity of 1-bromo-2-chloroethane and its analogues were indicated by several models within each predictive tool (see Table C-7). Table C-8 provides a list of the specific SAs that underlie the findings of a concern for carcinogenicity or mutagenicity in Table C-7.

OECD QSAR Toolbox models showed a concern for in vitro and in vivo mutagenicity for 1-bromo-2-chloroethane and all analogues based on structural alerts, a concern for CAs for 1-bromo-2-chloroethane, and concerns for all analogues based on protein binding alerts (see Table C-7). OECD QSAR Toolbox models also showed a concern for mutagenicity, CAs, and MN induction for 1-bromo-2-chloroethane, 1,2-DBE, 1,2-DCE, and 1,2-DCP; no results were reported for 1,2-DB-3-CP or TCP. The ToxRead and VEGA models also indicated a concern for mutagenicity for 1-bromo-2-chloroethane and all analogues.

OECD QSAR Toolbox models showed a concern for carcinogenicity for 1-bromo-2-chloroethane and all analogues based on structural alerts (see Table C-8). The level of carcinogenicity concern in OncoLogic was "moderate" for 1-bromo-2-chloroethane and 1,2-DCE based on SAR analysis only (haloalkanes and haloalkenes structural alert) and 1,2-dichloropropane based on experimental data and SAR analysis (haloalkanes and haloalkenes structural alert). The level of carcinogenicity concern in OncoLogic was "high-moderate" for 1,2-DBE, 1,2-DB-3-CP, and TCP based on experimental data and SAR analysis (haloalkanes and haloalkenes structural alert). Two carcinogenicity models in VEGA showed concern for carcinogenicity for 1-bromo-2-chloroethane (ISS and IFRMN/ANTARES). The CAESAR and IRFMN/ISSCAN-CGX models did not have reliable data for 1-bromo-2-chloroethane. For analogues, all four carcinogenicity models in VEGA showed concern for carcinogenicity for 1,2-DBE, 1,2-DCE, 1,2-DB-3-CP, and TCP. For 1,2-DCP, the ISS and IRFMN/ISSCAN-CGX models in VEGA showed concern for carcinogenicity, but the CAESAR and IRFMN/ANTARES models showed no concern for carcinogenicity. The ToxAlerts tool showed concern for nongenotoxic carcinogenicity for 1-bromo-2-chloroethane and all analogues based on structural alerts (aliphatic halogens). In contrast, the Toxtree tool indicated that there was no concern for nongenotoxic carcinogenicity for 1-bromo-2-chloroethane or any of its analogues.

The ToxAlerts and Toxtree tools showed a concern for genotoxic carcinogenicity for 1-bromo-2-chloroethane and all analogues based on various structural alerts (see Table C-8).

| Table (| C-7. Heat Map Illustrating the Structural Alert and SAR Pre 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Anal | | | Res | sult | s fo | r |
|-------------------------|---|------------------------|-------------------|--------------------|---------------------|-----------------------------|------------------------|
| Tool | Modelª | 1-Bromo-2-chloroethane | 1,2-Dibromoethane | 1,2-Dichloroethane | 1,2-Dichloropropane | 1,2-Dibromo-3-chloropropane | 1,2,3-Trichloropropane |
| Mutagenici | ty/genotoxicity alerts | | | | | | |
| | DNA alerts for Ames by OASIS | | | | | | |
| OECD | DNA alerts for CA and MNT by OASIS | | | | | | |
| QSAR | In vitro mutagenicity (Ames test) alerts by ISS | | | | | | |
| Toolbox | In vivo mutagenicity (micronucleus) alerts by ISS | | | | | | |
| | Protein binding alerts for CA by OASIS | | | | | | |
| ToxRead | ToxRead (mutagenicity) | | | | | | |
| | Mutagenicity (Ames test) CONSENSUS model-assessment | | | | | | |
| | Mutagenicity (Ames test) model (CAESAR)-assessment | | | | | | |
| VEGA | Mutagenicity (Ames test) model (SARpy/IRFMN)—assessment | | | | | | |
| | Mutagenicity (Ames test) model (ISS)-assessment | | | | | | |
| | Mutagenicity (Ames test) model (k-NN/read-across)—assessment | | | | | | |
| Carcinoger | icity alerts | | | | | | |
| OECD QSAR Toolbox | Carcinogenicity (genotoxicity and nongenotoxicity) alerts by ISS | | | | | | |
| OncoLogic | OncoLogic (prediction of the carcinogenic potential of the chemical) | | | | | | |
| | Carcinogenicity model (CAESAR)—assessment | | | | | | |
| VEGA | Carcinogenicity model (ISS)—assessment | | | | | | |
| | Carcinogenicity model (IRFMN/ANTARES)—assessment | | | | | | |
| | Carcinogenicity model (IRFMN/ISSCAN-CGX)—assessment | | | | | | |
| ToxAlerts | Aliphatic halogens (for nongenotoxic carcinogenicity) | | | | | | |
| Toxtree | Nongenotoxic carcinogenicity | | | | | | |

| Table | C-7. Heat Map Illustrating the Structural Alert and SAR Pro 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Ana | | | Res | sult | s foi | r | |
|-----------|--|------------------------|------------------|-------------------|--------------------|----------------------------|-----------------------|--|
| | | l-Bromo-2-chloroethane | ,2-Dibromoethane | ,2-Dichloroethane | ,2-Dichloropropane | ,2-Dibromo-3-chloropropane | ,2,3-Trichloropropane | |
| Tool | Model ^a | 4 | 1, | 1, | 1, | 1,; | 1, | |
| Combined | | | | | | | | |
| | Aliphatic halide (general) (for genotoxic carcinogenicity, mutagenicity) | | | | | | | |
| | Aliphatic halide (specific) (for genotoxic carcinogenicity, mutagenicity) | | | | | | | |
| ToxAlerts | Aliphatic halogens (for genotoxic carcinogenicity, mutagenicity) | | | | | | | |
| | Aromatic and aliphatic substituted primary alkyl halides (for genotoxic carcinogenicity, mutagenicity) | | | | | | | |
| | Structural alert for genotoxic carcinogenicity | | | | | | | |
| Toxtree | Model results outside the applicability domain for carcinogenicity/mutagenicity. | | | | | | | |
| Model | results or alerts indicating no concern for carcinogenicity/mutagenicity. | | | | | | | |
| Model | results outside the applicability domain for carcinogenicity/mutagenicity. | | | | | | | |
| Model | results or alerts indicating concern for carcinogenicity/mutagenicity. | | | | | | | |

^aAll tools and models described in Appendix B were used. Models with results or alerts are presented in the heat map (models without results were omitted).

ANTARES = Alternative Non-Testing Methods Assessed for REACH Substances; CA = chromosomal aberration; CAESAR = Computer-Assisted Evaluation of industrial chemical Substances According to Regulations; CONSENSUS = consensus assessment based on multiple models (CAESAR, SARpy, ISS, and *k*-NN); DNA = deoxyribonucleic acid; IRFMN = Istituto di Ricerche Farmacologiche Mario Negri; ISS = Istituto Superiore di Sanità; ISSCAN-CGX = Istituto Superiore di Sanità Chemical Carcinogen; *k*-NN = *k*-nearest neighbor; MNT = micronucleus test; OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; REACH = Registration, Evaluation, Authorisation and Restriction of Chemicals; SAR = structure-activity relationship; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

| Table C-8. Structural Alerts for 1-Bromo-2-Chloroethane (CASRN 107-04-0) andAnalogues | | | | | | | | |
|---|-------------------|---|--|--|--|--|--|--|
| SA | Tool | Compounds | | | | | | |
| Haloalkanes and haloalkenes | OncoLogic | 1-Bromo-2-chloroethane ^a ,1,2-DBE ^b , 1,2-DCE, ^a 1,2-DCP ^c , 1,2-DB-3-CP ^b , TCP ^b | | | | | | |
| Aliphatic halogens | ToxAlerts | 1-Bromo-2-chloroethane, 1,2-DBE, 1,2-DCE, | | | | | | |
| | Toxtree | 1,2-DCP, 1,2-DB-3-CP, TCP | | | | | | |
| | OECD QSAR Toolbox | | | | | | | |
| Aliphatic halide | ToxAlerts | 1-Bromo-2-chloroethane, 1,2-DBE, 1,2-DCE, | | | | | | |
| Aromatic and aliphatic substituted primary alkyl halides | | 1,2-DCP, 1,2-DB-3-CP, TCP | | | | | | |
| Vicinal dihaloalkanes | OECD QSAR Toolbox | 1-Bromo-2-chloroethane, 1,2-DBE, 1,2-DCE, 1,2-DCP | | | | | | |
| Halogenated vicinal hydrocarbons | | 1-Bromo-2-chloroethane, 1,2-DBE, 1,2-DCE, 1,2-DCP, 1,2-DB-3-CP, TCP | | | | | | |

^aIdentified as moderate alert based on SAR analysis only.

Г

^bIdentified as high-moderate alert based on experimental data and SAR analysis.

^cIdentified as moderate alert based on experimental data and SAR analysis.

1,2-DB-3-CP = 1,2-dibromo-3-chloropropane; 1,2-DBE = 1,2-dibromoethane; 1,2-DCE = 1,2-dichloropthane; 1,2-DCP = 1,2-dichloropropane; OECD = Organisation for Economic Co-operation and Development; QSAR = quantitative structure-activity relationship; SA = structural alert; SAR = structure-activity relationship; TCP = 1,2,3-trichloropropane.

STEP 7. EVIDENCE INTEGRATION FOR SCREENING EVALUATION OF 1-BROMO-2-CHLOROETHANE CARCINOGENICITY

Table C-9 presents the data for multiple lines of evidence pertinent to the screening evaluation of the carcinogenic potential of 1-bromo-2-chloroethane.

| Table C-9. Integration of Evidence for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Analogues | | | | | | | |
|--|--|--|--|--|--|--|--|
| Evidence Streams | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 | |
| Structure | CI | Br | CICI | CI CI CH ₃ | Cl Br Br | CICICI | |
| Analogue selection and evaluation (see Steps 1 and 2) | Target compound Contains: (1) two labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | Contains: (1) two labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | Contains: (1) two labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | Contains: (1) two labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | Contains: (1) three labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | Contains: (1) three labile halogen substituents; (2) halogens attached to adjacent, primary carbon atoms; and (3) no more than one halogen per carbon atom | |
| Experimental genotoxicity data (see Step 3) | Mutagenic Clastogenic DNA damaging Forms DNA adducts | Mutagenic Clastogenic DNA damaging Forms DNA adducts Induces cell transformation | Mutagenic Clastogenic DNA damaging Forms DNA adducts | Not a potent mutagen Clastogenic DNA damaging | Mutagenic Clastogenic DNA damaging Forms DNA adducts Induces cell transformation | Mutagenic Clastogenic DNA damaging Forms DNA adducts Induces cell transformation | |
| ADME evaluation (see Step 4) | Oxidation (based on analogy to other dihaloalkanes) and GSH conjugation (demonstrated) Expected to form episulfonium ion | Common metabolic pathways with other haloalkanes (oxidation and GSH conjugation) Forms episulfonium | Common metabolic pathways with other haloalkanes (oxidation and GSH conjugation) Forms episulfonium | Common metabolic pathways with other haloalkanes (oxidation and GSH conjugation) Expected to form episulfonium ion | Common metabolic pathways with other haloalkanes (oxidation and GSH conjugation) Forms episulfonium ion | Common metabolic pathways with other haloalkanes (oxidation and GSH conjugation) Forms episulfonium ion | |

| | Table C-9. Integration of Evidence for 1-Bromo-2-Chloroethane (CASRN 107-04-0) and Analogues | | | | | | | |
|---|--|--|---|--|---|--|--|--|
| Evidence Streams | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 | | |
| Cancer data and MOA ^a (see Step 5) | No data | Multisite carcinogen in rats and mice following oral or inhalation exposure Proposed MOA: Mutagenicity and/or DNA damage | Multisite carcinogen in rats and mice following oral or inhalation exposure Potential MOA: Mutagenicity and/or DNA damage | Liver tumors (mice) and marginal increase in mammary gland tumors (rats) following oral exposure; nasal tumors (rats) and lung and Harderian gland tumors (mice) following inhalation exposure | Multisite carcinogen in rats and mice following oral exposure; primarily respiratory tract cancer (nasal, pharynx, lung) following inhalation exposure MOA: mutagenicity | e | | |
| | | | | Proposed MOA: DNA damage | | | | |

| Evidence Streams | 1-Bromo- 2-Chloroethane CASRN 107-04-0 | 1,2-Dibromoethane CASRN 106-93-4 | 1,2-Dichloroethane CASRN 107-06-2 | 1,2-Dichloropropane CASRN 78-87-5 | 1,2-Dibromo- 3-Chloropropane CASRN 96-12-8 | 1,2,3-Trichloro- propane CASRN 96-18-4 |
|--|---|--|--------------------------------------|---|--|---|
| Common structural alerts and SAR predictions (see Step 6) | ALERTS • Haloalkanes and haloalkenes • Aliphatic halogens • Aliphatic halide • Vicinal dihaloalkanes • Halogenated vicinal hydrocarbons SAR PREDICTIONS: Concerns for mutagenicity and carcinogenicity in most models; no concern for nongenotoxic carcinogenicity in Toxtree | ALERTS Haloalkanes and haloalkenes Aliphatic halogens Aliphatic halide Vicinal dihaloalkanes Halogenated vicinal hydrocarbons SAR PREDICTIONS: Concerns for mutagenicity and carcinogenicity in most models; no concern for nongenotoxic carcinogenicity in Toxtree | hydrocarbons | Halogenated vicinal | ALERTS Haloalkanes and haloalkenes Aliphatic halogens Aliphatic halide Halogenated vicinal hydrocarbons SAR PREDICTIONS: Concerns for mutagenicity and carcinogenicity in most models; no concern for nongenotoxic carcinogenicity in Toxtree | ALERTS Haloalkanes and haloalkenes Aliphatic halogens Aliphatic halide Halogenated vicinal hydrocarbons SAR PREDICTIONS: Concerns for mutagenicity and carcinogenicity in mos models; no concern for nongenotoxic carcinogenicity in Toxtree |

^aFor MOA descriptions, "MOA" was used if the U.S. EPA did a formal MOA evaluation and made a conclusion regarding cancer MOA(s); "proposed MOA" was used if the U.S. EPA did not do formal MOA evaluation, but proposed possible MOAs; and "potential MOA" was used if the U.S. EPA did not discuss potential MOAs but available experimental data suggest an MOA.

ADME = absorption, distribution, metabolism, excretion; DNA = deoxyribonucleic acid; GSH = glutathione; MOA = mode of action; SAR = structure-activity relationship; VEGA = Virtual models for property Evaluation of chemicals within a Global Architecture.

STEP 8. QUALITATIVE LEVEL OF CONCERN FOR 1-BROMO-2-CHLOROETHANE POTENTIAL CARCINOGENICITY

Table C-10 identifies the qualitative level of concern for potential carcinogenicity of 1-bromo-2-chloroethane based on the multiple lines of evidence described above. Due to lack of information supporting carcinogenic portal-of-entry effects, the qualitative level of concern for this chemical is considered to be applicable to all routes of exposure.

Г

| Table C-10. Qualitative Level of Concern for Carcinogenicity of1-Bromo-2-Chloroethane (CASRN 107-04-0) | | | | | | |
|--|--------------|--|--|--|--|--|
| Level of Concern | Designation | Comments | | | | |
| Concern for potential carcinogenicity | Selected | All analogues of 1-bromo-2-chloroethane have carcinogenic potential based on tumors observed in rodent studies. The U.S. EPA identified a mutagenic MOA for 1,2-DB-3-CP and TCP, and all analogues except 1,2-DCP are mutagenic (evidence is limited for 1,2-DCP). DNA damage is also a potential MOA suggested for 1,2-DBE and 1,2-DCP, and experimental evidence indicates that all analogues are DNA damaging and all except 1,2-DCP form DNA adducts (no data for 1,2-DCP). Common metabolic pathways exist between 1-bromo-2-chloroethane and the analogue compounds. 1-Bromo-2-chloroethane and its analogues have similar structural alerts (e.g., haloalkanes and haloalkenes, aliphatic halogens, aliphatic halide, halogenated vicinal hydrocarbons) and similar SAR predictions showing concern for carcinogenicity and/ or genotoxicity. | | | | |
| Inadequate information for assigning qualitative level of concern | Not selected | NA | | | | |

1,2-DB-3-CP = 1,2-dibromo-3-chloropropane; 1,2-DBE = 1,2-dibromoethane; 1,2-DCP = 1,2-dichloropropane; DNA = deoxyribonucleic acid; MOA = mode of action; NA = not applicable; SAR = structure-activity relationship; TCP = 1,2,3-trichloropropane.

APPENDIX D. REFERENCES

- <u>ACGIH</u> (American Conference of Governmental Industrial Hygienists). (2020). 2020 TLVs and BEIs: Based on the documentation of the threshold limit values for chemical substances and physical agents & biological exposure indices. Cincinnati, OH.
- <u>Akiba, N; Shiizaki, K; Matsushima, Y; Endo, O; Inaba, K; Totsuka, Y.</u> (2017). Influence of GSH S-transferase on the mutagenicity induced by dichloromethane and 1,2-dichloropropane. Mutagenesis 32: 455-462. <u>http://dx.doi.org/10.1093/mutage/gex014</u>
- <u>Amann, RP; Berndtson, WE.</u> (1986). Assessment of procedures for screening agents for effects on male reproduction: Effects of dibromochloropropane (DBCP) on the rat. Fundam Appl Toxicol 7: 244-255. <u>http://dx.doi.org/10.1016/0272-0590(86)90154-5</u>
- ATSDR (Agency for Toxic Substances and Disease Registry). (2001). Toxicological profile for 1,2-dichloroethane [ATSDR Tox Profile]. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service.

https://www.atsdr.cdc.gov/ToxProfiles/tp.asp?id=592&tid=110

- ATSDR (Agency for Toxic Substances and Disease Registry). (2018). Minimal risk levels (MRLs). June 2018. Atlanta, GA: Agency for Toxic Substances and Disease Registry (ATSDR).
- Barber, ED; Donish, WH; Mueller, KR. (1981). A procedure for the quantitative measurement of the mutagenicity of volatile liquids in the Ames Salmonella/microsome assay. Mutat Res Genet Toxicol 90: 31-48. <u>http://dx.doi.org/10.1016/0165-1218(81)90048-3</u>
- Brem, H; Stein, AB; Rosenkranz, HS. (1974). The mutagenicity and DNA-modifying effect of haloalkanes. Cancer Res 34: 2576-2579.
- <u>CalEPA</u> (California Environmental Protection Agency). (1999a). Public health goal for 1,2dichloroethane in drinking water. Sacramento, CA: Office of Environmental Health Hazard Assessment. <u>https://oehha.ca.gov/media/downloads/pesticides/report/12dcaf.pdf</u>
- <u>CalEPA</u> (California Environmental Protection Agency). (1999b). Public health goal for 1,2dichloropropane in drinking water. Sacramento, CA: Office of Environmental Health Hazard Assessment, Pesticide and Environmental Toxicology Section. <u>https://oehha.ca.gov/media/downloads/water/public-health-goal/12dcpf.pdf</u>
- <u>CalEPA</u> (California Environmental Protection Agency). (2004). No significant risk level (NSRL) for the proposition 65 carcinogen 1,2-dichloropropane. Sacramento, CA: Office of Environmental Health Hazard Assessment. https://oehha.ca.gov/media/downloads/crnr/oehha2004a.pdf
- CalEPA (California Environmental Protection Agency). (2009). Public health goals for chemicals in drinking water: 1,2,3-trichloropropane. Sacramento, CA: Office of Environmental Health Hazard Assessment.

https://oehha.ca.gov/media/downloads/water/chemicals/phg/082009tcpphg.pdf

- <u>CalEPA</u> (California Environmental Protection Agency). (2011). Air Toxics Hot Spots program technical support document for cancer potencies. Appendix B. Chemical-specific summaries of the information used to derive unit risk and cancer potency values. Updated 2011. Sacramento, CA: Office of Environmental Health Hazard Assessment. https://oehha.ca.gov/air/crnr/technical-support-document-cancer-potency-factors-2009
- <u>CalEPA</u> (California Environmental Protection Agency). (2018). Chemicals known to the state to cause cancer or reproductive toxicity May 25, 2018. (Proposition 65 list). Sacramento, CA: Office of Enironmental Health Hazard Assessment. <u>http://oehha.ca.gov/proposition-65/proposition-65-list</u>

- <u>CalEPA</u> (California Environmental Protection Agency). (2019). Consolidated table of OEHHA/ARB approved risk assessment health values (September 19, 2019 ed.). Sacramento, CA: California Air Resources Board. https://www.arb.ca.gov/toxics/healthval/contable.pdf
- <u>CalEPA</u> (California Environmental Protection Agency). (2020). Hot spots unit risk and cancer potency values. Appendix A. Updated October 2020. Sacramento, CA: Office of Environmental Health Hazard Assessment.

https://oehha.ca.gov/media/downloads/crnr/appendixa.pdf

- <u>Cheever, KL; Cholakis, JM; El-Hawari, AM; Kovatch, RM; Weisburger, EK.</u> (1990). Ethylene dichloride: The influence of disulfiram or ethanol on oncogenicity, metabolism, and DNA covalent binding in rats. Toxicol Sci 14: 243-261. <u>http://dx.doi.org/10.1016/0272-0590(90)90205-X</u>
- <u>ChemIDplus.</u> (2021). ChemIDplus advanced. Available online at <u>https://chem.nlm.nih.gov/chemidplus/</u>
- <u>Chroust, K; Pavlová, M; Prokop, Z; Mendel, J; Božková, K; Kubát, Z; Zaji, CV; Damborský, J.</u>
 (2006). Quantitative structure-activity relationships for toxicity and genotoxicity of halogenated aliphatic compounds: Wing spot test of Drosophila melanogaster.
 Chemosphere 67: 152-159. <u>http://dx.doi.org/10.1016/j.chemosphere.2006.09.020</u>
- Clark, HA; Snedeker, SM. (2005). Critical evaluation of the cancer risk of dibromochloropropane (DBCP) [Review]. J Environ Sci Health C Environ Carcinog Ecotoxicol Rev 23: 215-260. <u>http://dx.doi.org/10.1080/10590500500234996</u>
- <u>CPDB</u> (Carcinogenic Potency Database). (2011). The carcinogenic potency project: The carcinogenic potency database [Database]: Department of Energy; National Cancer Institute; Environmental Protection Agency; National Institute of Environmental Health Sciences; National Toxicology Program; University of California, Berkeley. Retrieved from <u>https://www.nlm.nih.gov/databases/download/cpdb.html</u>
- Crebelli, R; Andreoli, C; Carere, A; Conti, L; Crochi, B; Cotta-Ramusino, M; Benigni, R. (1995). Toxicology of halogenated aliphatic hydrocarbons: Structural and molecular determinants for the disturbance of chromosome segregation and the induction of lipid peroxidation. Chem Biol Interact 98: 113-129. <u>http://dx.doi.org/10.1016/0009-2797(95)03639-3</u>
- Daniel, FB; Robinson, M; Olson, GR; York, RG; Condie, LW. (1994). Ten and ninety-day toxicity studies of 1,2-dichloroethane in Sprague-Dawley rats. Drug Chem Toxicol 17: 463-477. http://dx.doi.org/10.3109/01480549409014312
- Dekant, W; Vamvakas, S. (1993). Glutathione-dependent bioactivation of xenobiotics [Review]. Xenobiotica 23: 873-887. <u>http://dx.doi.org/10.3109/00498259309059415</u>
- Dohn, DR; Graziano, MJ; Casida, JE. (1988). Metabolites of [3-13C]1,2-dibromo-3chloropropane in male rats studied by 13C and 1H-13C correlated two-dimensional NMR spectroscopy. Biochem Pharmacol 37: 3485-3495. <u>http://dx.doi.org/10.1016/0006-2952(88)90701-0</u>
- ECHA (European Chemicals Agency). (2015). Registered substances: Perylene-3,4:9,10tetracarboxydiimide [Database]. Helsinki, Finland. Retrieved from <u>https://echa.europa.eu/registration-dossier/-/registered-dossier/10330</u>
- Foote, RH; Berndtson, WE; Rounsaville, TR. (1986a). Use of quantitative testicular histology to assess the effect of dibromochloropropane (DBCP) on reproduction in rabbits. Toxicol Sci 6: 638-647. http://dx.doi.org/10.1016/0272-0590(86)90176-4

- Foote, RH; Schermerhorn, EC; Simkin, ME. (1986b). Measurement of semen quality, fertility, and reproductive hormones to assess dibromochloropropane (DBCP) effects in live rabbits. Fundam Appl Toxicol 6: 628-637. <u>http://dx.doi.org/10.1016/0272-</u>0590(86)90175-2
- <u>Gargas, ML; Burgess, RJ; Voisard, DE; Cason, GH; Andersen, ME.</u> (1989). Partition coefficient of low-molecular-weight volatile chemicals in various tissues and liquids. Toxicol Appl Pharmacol 98: 87-99. <u>http://dx.doi.org/10.1016/0041-008x(89)90137-3</u>
- <u>Gingell, R; Beatty, PW; Mitschke, HR; Page, AC; Sawin, VL; Putcha, L; Kramer, WG.</u> (1987). Toxicokinetics of 1,2-dibromo-3-chloropropane (DBCP) in the rat. Toxicol Appl Pharmacol 91: 386-394. http://dx.doi.org/10.1016/0041-008x(87)90060-3
- <u>Guengerich, FP.</u> (1994). Metabolism and genotoxicity of dihaloalkanes [Review]. Adv Pharmacol 27: 211-236. http://dx.doi.org/10.1016/S1054-3589(08)61034-0
- <u>Guengerich, FP; Peterson, LA; Cmarik, JL; Koga, N; Inskeep, PB.</u> (1987). Activation of dihaloalkanes by glutathione conjugation and formation of DNA adducts. Environ Health Perspect 76: 15-18. <u>http://dx.doi.org/10.1289/ehp.877615</u>
- <u>Gwinn, MR; Johns, DO; Bateson, TF; Guyton, KZ.</u> (2011). A review of the genotoxicity of 1,2dichloroethane (EDC) [Review]. Mutat Res 727: 42-53. http://dx.doi.org/10.1016/j.mrrev.2011.01.001
- Heindel, JJ; Berkowitz, AS; Kyle, G; Luthra, R; Bruckner, JV. (1989). Assessment in rats of the gonadotoxic and hepatorenal toxic potential of dibromochloropropane (DBCP) in drinking water. Fundam Appl Toxicol 13: 804-815. http://dx.doi.org/10.1093/toxsci/13.4.804
- <u>Hughes, TJ; Simmons, DS; Monteith, LG; Myers, LE; Claxton, LD.</u> (1987). Mutagenicity of 31 organic compounds in a modified preincubation ames assay with Salmonella typhimurium strains TA100 and TA102 [Abstract]. Environ Mutagen 9: 49. <u>http://dx.doi.org/10.1002/em.2860090602</u>
- Humphreys, WG; Kim, DH; Guengerich, FP. (1991). Isolation and characterization of N7-guanyl adducts derived from 1,2-dibromo-3-chloropropane. Chem Res Toxicol 4: 445-453. http://dx.doi.org/10.1021/tx00022a008
- <u>IARC</u> (International Agency for Research on Cancer). (1995). Dry cleaning, some chlorinated solvents and other industrial chemicals: Summary of data reported and evaluation. Lyon, France.
- <u>IARC</u> (International Agency for Research on Cancer). (1999). Re-evaluation of some organic chemicals, hydrazine, and hydrogen peroxide [IARC Monograph]. In IARC monographs on the evaluation of carcinogenic risks to humans. Lyon, France. <u>http://monographs.iarc.fr/ENG/Monographs/vol71/mono71.pdf</u>
- <u>IARC</u> (International Agency for Research on Cancer). (2017). 1,2-Dichloropropane. Some chemicals used as solvents and in polymer manufacture [IARC Monograph]. In IARC monographs on the evaluation of carcinogenic risks to humans (pp. 141-175). Lyon, France.
- <u>IARC</u> (International Agency for Research on Cancer). (2018). IARC monographs on the evaluation of carcinogenic risk to humans. http://monographs.iarc.fr/ENG/Monographs/PDFs/index.php
- IPCS (International Programme on Chemical Safety). (2020). INCHEM: Chemical safety information from intergovernmental organizations [Database]. Geneva, Switzerland: World Health Organization, Canadian Centre for Occupational Health and Safety. Inter-Organization Programme for the Sound Management of Chemicals. Retrieved from <u>http://www.inchem.org/</u>

Japan Industrial Safety and Health Association. (2005a). Summary of inhalation carcinogenicity study of 1-bromo-3-chloropropane in BDF1 mice. (Study No.0418). Ministry of Health, Labour and Welfare. Japan Bioassay Research Center.

http://anzeninfo.mhlw.go.jp/user/anzen/kag/pdf/gan/1-Bromo-3-Chloropropane_Mice.pdf

Japan Industrial Safety and Health Association. (2005b). Summary of inhalation carcinogenicity study of 1-bromo-3-chloropropane in F344 rats. (Study No.0417). Ministry of Health, Labour and Welfare. Japan Bioassay Research Center.

http://anzeninfo.mhlw.go.jp/user/anzen/kag/pdf/gan/1-Bromo-3-Chloropropane_Rats.pdf

- Jean, PA; Reed, DJ. (1992). Utilization of glutathione during 1,2-dihaloethane metabolism in rat hepatocytes. Chem Res Toxicol 5: 386-391. <u>http://dx.doi.org/10.1021/tx00027a011</u>
- Johnston, RV; Mensik, DC; Taylor, HW; Jersey, GC; Dietz, FK. (1986). Single-generation drinking water reproduction study of 1,2-dibromo-3-chloropropane in Sprague-Dawley rats. Bull Environ Contam Toxicol 37: 531–537. <u>http://dx.doi.org/10.1007/BF01607800</u>
- Kirkland, D; Aardema, M; Henderson, L; Müller, L. (2005). Evaluation of the ability of a battery of three in vitro genotoxicity tests to discriminate rodent carcinogens and noncarcinogens: I. Sensitivity, specificity and relative predictivity. Mutat Res 584: 1-256. <u>http://dx.doi.org/10.1016/j.mrgentox.2005.02.004</u>
- Lamb, JC; Tyl, R; Lawton, AD. (1997). Reproductive toxicology. Dibromochloropropane [Abstract]. Environ Health Perspect 105 Suppl 1: 299-300.
- Lebaron, MJ; Hotchkiss, JA; Zhang, F; Koehler, MW; Boverhof, DR. (2021). Investigation of potential early key events and mode of action for 1,2-dichloroethane-induced mammary tumors in female rats. J Appl Toxicol 41: 362-374. <u>http://dx.doi.org/10.1002/jat.4048</u>
- Lewis, RJ, Sr; Hawley, GG. (2007). Sym-bromochloroethane. In RJ Lewis, Sr.; GG Hawley (Eds.), Hawley's condensed chemical dictionary (15th ed., pp. 183). Hoboken, NJ: John Wiley & Sons. <u>http://dx.doi.org/10.1002/9780470114735.hawley02335</u>
- Maltoni, C; Valgimigli, L; Scarnato, C. (1980). Long-term carcinogenic bioassays on ethylene dichloride administered by inhalation to rats and mice. In B Ames; P Infante; R Reitz (Eds.), Ethylene dichloride: A potential health risk? (pp. 3-29). Cold Spring Harbor, NY: Cold Spring Harbor Laboratory.
- Marchand, DH; Reed, DJ. (1989). Identification of the reactive glutathione conjugate S-(2chloroethyl)glutathione in the bile of 1-bromo-2-chloroethane-treated rats by highpressure liquid chromatography and precolumn derivatization with o-phthalaldehyde. Chem Res Toxicol 2: 449-454. http://dx.doi.org/10.1021/tx00012a015
- Meistrich, ML; Wilson, G; Shuttlesworth, GA; Porter, KL. (2003). Dibromochloropropane inhibits spermatogonial development in rats. Reprod Toxicol 17: 263-271. http://dx.doi.org/10.1016/S0890-6238(03)00007-8
- Moody, DE; James, JL; Smuckler, EA. (1980). Cytochrome P-450 lowering effect of alkyl halides, correlation with decrease in arachidonic acid. Biochem Biophys Res Commun 97: 673-679. <u>http://dx.doi.org/10.1016/0006-291X(80)90317-4</u>
- Nagano, K; Umeda, Y; Senoh, H; Gotoh, K; Arito, H; Yamamoto, S; Matsushima, T. (2006). Carcinogenicity and chronic toxicity in rats and mice exposed by inhalation to 1,2dichloroethane for two years. J Occup Health 48: 424-436. http://dx.doi.org/10.1539/joh.48.424
- NIOSH (National Institute for Occupational Safety and Health). (2018). NIOSH pocket guide to chemical hazards. Index of chemical abstracts service registry numbers (CAS No.). Atlanta, GA. <u>http://www.cdc.gov/niosh/npg/npgdcas.html</u>

- Nitschke, KD; Kociba, RJ; Keyes, DG; McKenna, MJ. (1981). A thirteen week repeated inhalation study of ethylene dibromide in rats. Fundam Appl Toxicol 1: 437-442. http://dx.doi.org/10.1016/s0272-0590(81)80024-3
- NLM (National Library of Medicine). (2005). Chemical Carcinogenesis Research Information System (CCRIS). 1-Bromo-2-chloroethane. Retrieved from https://www.nlm.nih.gov/databases/download/ccris.html
- NLM (National Library of Medicine). (2016). ChemIDplus advanced. 1-Bromo-2-chloroethane, CASRN: 107-04-0. Available online at https://chem.nlm.nih.gov/chemidplus/rn/startswith/107-04-0
- NOAA (National Oceanic and Atmospheric Administration). (2015). CAMEO Chemicals. Version 2.4.2. 1-Chloro-2-bromoethane, CAS number 107-04-0 (Version 2.5 rev 1). Retrieved from <u>http://m.cameochemicals.noaa.gov</u>
- <u>NRC</u> (National Research Council). (2014). Acute exposure guideline levels for selected airborne chemicals: Volume 17. Washington, DC: National Academies Press. <u>http://dx.doi.org/10.17226/18796</u>
- NTP (National Toxicology Program). (1978). Bioassay of dibromochloropropane for possible carcinogenicity. Natl Cancer Inst Carcinog Tech Rep Ser 28: 1-66.
- NTP (National Toxicology Program). (1982). Carcinogenesis bioassay of 1,2-dibromoethane (CAS no 106-93-4) in F344 rats and B6C3F1 mice (inhalation study) (pp. 1-163). https://search.proquest.com/docview/1859405604?accountid=171501
- NTP (National Toxicology Program). (1990). Genetic toxicity evaluation of 1-chloro-2bromoethane in Salmonella/E. coli mutagenicity test or Ames test study 406801. Chemical Effects in Biological Systems (CEBS). Available online at http://tools.niehs.nih.gov/cebs3/ntpViews/?studyNumber=002-01115-0001-0000-1
- NTP (National Toxicology Program). (1991). NTP technical report on the toxicity studies of 1,2-dichloroethane (ethylene dichloride) in F344/N rats, Sprague Dawley rats, Osborne-Mendel rats, and B6C3F1 mice (drinking water and gavage Studies) (CAS No. 107-06-2). (NTP TOX 4. NIH Publication No. 91-3123). https://ntp.niehs.nih.gov/ntp/htdocs/st_rpts/tox004.pdf
- <u>NTP</u> (National Toxicology Program). (2016a). 14th Report on carcinogens. Research Triangle Park, NC. <u>https://ntp.niehs.nih.gov/pubhealth/roc/index-1.html</u>
- NTP (National Toxicology Program). (2016b). Report on carcinogens, 14th edition: 1,2-Dibromo-3-chloropropane. In Report on Carcinogens. Research Triangle Park, NC: U.S. Department of Health and Human Services, National Toxicology Program. <u>https://ntp.niehs.nih.gov/ntp/roc/content/profiles/dibromochloropropane.pdf</u>
- NTP (National Toxicology Program). (2016c). Report on carcinogens, 14th edition: 1,2-Dibromoethane. In Report on Carcinogens. Research Triangle Park, NC: U.S. Department of Health and Human Services, National Toxicology Program. <u>https://ntp.niehs.nih.gov/ntp/roc/content/profiles/dibromoethane.pdf</u>
- NTP (National Toxicology Program). (2016d). Report on carcinogens, 14th edition: 1,2-Dichloroethane. In Report on Carcinogens. Research Triangle Park, NC: U.S. Department of Health and Human Services, National Toxicology Program. <u>https://ntp.niehs.nih.gov/ntp/roc/content/profiles/dichloroethane.pdf</u>
- NTP (National Toxicology Program). (2016e). Report on carcinogens, 14th edition: 1,2,3-Trichloropropane. In Report on Carcinogens. Research Triangle Park, NC: U.S. Department of Health and Human Services, National Toxicology Program. <u>https://ntp.niehs.nih.gov/ntp/roc/content/profiles/trichloropropane.pdf</u>

- <u>NTP</u> (National Toxicology Program). (2017). NTP technical reports index. Available online at <u>https://ntp.niehs.nih.gov/results/summaries/chronicstudies/index.html</u>
- <u>NTP</u> (National Toxicology Program). (2021). Testing status of 1,2-dichloroethane 10962-L. Available online at <u>https://ntp.niehs.nih.gov/go/ts-10962-1</u>
- Oda, Y; Yamazaki, H; Thier, R; Ketterer, B; Guengerich, FP; Shimada, T. (1996). A new Salmonella typhimurium NM5004 strain expressing rat glutathione S-transferase 5-5: Use in detection of genotoxicity of dihaloalkanes using an SOS/umu test system. Carcinogenesis 17: 297-302. http://dx.doi.org/10.1093/carcin/17.2.297
- <u>OECD</u> (Organisation for Economic Co-operation and Development). (2018). The OECD QSAR toolbox for grouping chemicals into categories. Retrieved from <u>https://www.qsartoolbox.org/</u>
- <u>Omichinski, JG; Brunborg, G; Holme, JA; Søderlund, EJ; Nelson, SD; Dybing, E.</u> (1988a). The role of oxidative and conjugative pathways in the activation of 1,2-dibromo-3-chloropropane to DNA-damaging products in rat testicular cells. Mol Pharmacol 34: 74-79.
- <u>Omichinski, JG; Søderlund, EJ; Dybing, E; Pearson, PG; Nelson, SD.</u> (1988b). Detection and mechanism of formation of the potent direct-acting mutagen 2-bromoacrolein from 1,2-dibromo-3-chloropropane. Toxicol Appl Pharmacol 92: 286-294.
- OSHA (Occupational Safety & Health Administration). (2020a). Air contaminants: Occupational safety and health standards for shipyard employment, subpart Z, toxic and hazardous substances. (OSHA Standard 1915.1000). Washington, DC. https://www.osha.gov/pls/oshaweb/owadisp.show_document?p_table=STANDARDS&p_id=10286
- OSHA (Occupational Safety & Health Administration). (2020b). Table Z-1: Limits for air contaminants. Occupational safety and health standards, subpart Z, toxic and hazardous substances. Available online at

http://www.osha.gov/pls/oshaweb/owadisp.show_document?p_table=STANDARDS&p_id=9992

- Pearson, PG; Omichinski, JG; Myers, TG; Søderlund, EJ; Dybing, E; Nelson, SD. (1990). Metabolic activation of 1,2-dibromo-3-chloropropane to mutagenic metabolites: detection and mechanism of formation of (Z)- and (E)-2-chloro-3-(bromomethyl)oxirane. Chem Res Toxicol 3: 458-466.
- Rao, KS; Burek, JD; Murray, FJ; John, JA; Schwetz, BA; Bell, TJ; Potts, WJ; Parker, CM. (1983). Toxicologic and reproductive effects of inhaled 1,2-dibromo-3-chloropropane in rats. Toxicol Sci 3: 104-110110. http://dx.doi.org/10.1016/S0272-0590(83)80064-5
- Rao, KS; Burek, JD; Murray, FJ; John, JA; Schwetz, BA; Beyer, JE; Parker, CM. (1982). Toxicologic and reproductive effects of inhaled 1,2-dibromo-3-chloropropane in male rabbits. Toxicol Sci 2: 241-251.
- <u>Rao, KS; Murray, JS; Deacon, MM; John, JA; Calhoun, LL; Young, JT.</u> (1980). Teratogenicity and reproduction studies in animals inhaling ethylene dichloride. In B Ames; P Infante; R Reitz (Eds.), Banbury Report, Vol 5 Ethylene Dichloride: A Potential Health Risk (pp. P149-P166). Cold Spring Harbor, NY: Cold Spring Harbor Laboratory.
- Reel, JR; Wolkowski-Tyl, R; Lawton, AD. (1984). Dibromochloropropane: reproduction and fertility assessment in CD-1 mice when administered by gavage. Reel, JR; Wolkowski-Tyl, R; Lawton, AD.

- Reitz, RH; Fox, TR; Ramsey, JC; Quast, JF; Langvardt, PW; Watanabe, PG. (1982). Pharmacokinetics and macromolecular interactions of ethylene dichloride in rats after inhalation or gavage. Toxicol Appl Pharmacol 62: 190-204. http://dx.doi.org/10.1016/0041-008X(82)90117-X
- Reznik, G; Reznik-Schuller, H; Ward, JM; Stinson, SF. (1980a). Morphology of nasal-cavity tumours in rats after chronic inhalation of 1,2-dibromo-3-chloropropane. Br J Cancer 42: 772-781. http://dx.doi.org/10.1038/bjc.1980.311
- Reznik, G; Stinson, SF; Ward, JM. (1980b). Lung tumors induced by chronic inhalation of 1,2dibromo-3-chloropropane in B6C3F1 mice. Cancer Lett 10: 339-342. http://dx.doi.org/10.1016/0304-3835(80)90051-8
- <u>Reznik, G; Stinson, SF; Ward, JM.</u> (1980c). Respiratory pathology in rats and mice after inhalation of 1,2-dibromo-3-chloropropane or 1,2 dibromoethane for 13 weeks. Arch Toxicol 46: 233-240. <u>http://dx.doi.org/10.1007/BF00310439</u>
- Shimada, T; Yamazaki, H; Oda, Y; Hiratsuka, A; Watabe, T; Guengerich, FP. (1996). Activation and inactivation of carcinogenic dihaloalkanes and other compounds by glutathione Stransferase 5-5 in Salmonella typhimurium tester strain NM5004. Chem Res Toxicol 9: 333-340. http://dx.doi.org/10.1021/tx950125v
- Short, RD; Winston, JM; Hong, CB; Minor, JL; Lee, CC; Seifter, J. (1979). Effects of ethylene dibromide on reproduction in male and female rats. Toxicol Appl Pharmacol 49: 97-105. http://dx.doi.org/10.1016/0041-008X(79)90281-3
- <u>Søderlund, EJ; Brunborg, G; Omichinski, JG; Holme, JA; Dahl, JE; Nelson, SD; Dybing, E.</u> (1988). Testicular necrosis and DNA damage caused by deuterated and methylated analogs of 1,2-dibromo-3-chloropropane in the rat. Toxicol Appl Pharmacol 94: 437-447. <u>http://dx.doi.org/10.1016/0041-008X(88)90284-0</u>
- Spencer, HC; Rowe, VK; Adams, EM; McCollister, DD; Irish, DD. (1951). Vapor toxicity of ethylene dichloride determined by experiments on laboratory animals. Arch Ind Hyg Occup Med 4: 482-493.
- Spreafico, F; Zuccato, E; Marcucci, F; Sironi, M; Paglialunga, S; Madonna, M; Mussini, E. (1980). Pharmacokinetics of ethylene dichloride in rats treated by different routes and its long-term inhalatory toxicity. In B Ames; P Infante; R Reitz (Eds.), Ethylene dichloride: A potential health risk? (Banbury Report 5 ed., pp. 107-133). Cold Spring Harbor, NY: Cold Spring Harbor Laboratory.
- Stinson, SF; Reznik, G; Ward, JM. (1981). Characteristics of proliferative lesions in the nasal cavities of mice following chronic inhalation of 1,2-dibromoethane. Cancer Lett 12: 121-129. <u>http://dx.doi.org/10.1016/0304-3835(81)90047-1</u>
- Storer, RD; Conolly, RB. (1983). Comparative in vivo genotoxicity and acute hepatotoxicity of three 1,2-dihaloethanes. Carcinogenesis 4: 1491-1494. http://dx.doi.org/10.1093/carcin/4.11.1491
- Tan, EL; Hsie, AW. (1981). Mutagenicity and cytotoxicity of haloethanes as studied in the CHO/HGPRT system. Mutat Res 90: 183-191. <u>http://dx.doi.org/10.1016/0165-1218(81)90081-1</u>
- U.S. EPA (U.S. Environmental Protection Agency). (1985). Health and environmental effects profile for bromochloroethanes [EPA Report]. (EPA/600/X-85/391). Cincinnati, OH: U.S. Environmental Protection Agency, Office of Research and Development.

- U.S. EPA (U.S. Environmental Protection Agency). (1986). Guidelines for carcinogen risk assessment [EPA Report] (pp. 33993-34003). (EPA/630/R-00/004). Washington, DC: U.S. Environmental Protection Agency, Risk Assessment Forum. <u>https://www.epa.gov/iris/basic-information-about-integrated-risk-informationsystem#risk</u>
- <u>U.S. EPA</u> (U.S. Environmental Protection Agency). (1988). Recommendations for and documentation of biological values for use in risk assessment [EPA Report]. (EPA/600/6-87/008). Cincinnati, OH. <u>http://cfpub.epa.gov/ncea/cfm/recordisplay.cfm?deid=34855</u>
- U.S. EPA (U.S. Environmental Protection Agency). (1993). Integrated Risk Information System (IRIS) chemical assessment summary for 1,2-dichloroethane (CASRN 107-06-2). https://cfpub.epa.gov/ncea/iris/iris_documents/documents/subst/0149_summary.pdf
- U.S. EPA (U.S. Environmental Protection Agency). (1994). Methods for derivation of inhalation reference concentrations and application of inhalation dosimetry [EPA Report]. (EPA/600/8-90/066F). Research Triangle Park, NC. https://cfpub.epa.gov/ncea/risk/recordisplay.cfm?deid=71993&CFID=51174829&CFTO KEN=25006317
- U.S. EPA (U.S. Environmental Protection Agency). (2002). A review of the reference dose and reference concentration processes. (EPA/630/P-02/002F). Washington, DC. https://www.epa.gov/sites/production/files/2014-12/documents/rfd-final.pdf
- U.S. EPA (U.S. Environmental Protection Agency). (2003). Integrated Risk Information System (IRIS) chemical assessment summary for 1,2-dibromo-3-chloropropane (DBCP); CASRN 96-12-8.

https://cfpub.epa.gov/ncea/iris/iris_documents/documents/subst/0414_summary.pdf

- U.S. EPA (U.S. Environmental Protection Agency). (2004). Toxicological review of 1,2dibromoethane (CASRN 106-93-4) in support of summary information on the Integrated Risk Information System (IRIS) [EPA Report]. (EPA/635/R-04/067). Washington, DC. https://iris.epa.gov/static/pdfs/0361tr.pdf
- U.S. EPA (U.S. Environmental Protection Agency). (2005). Guidelines for carcinogen risk assessment [EPA Report]. (EPA/630/P-03/001F). Washington, DC. <u>https://www.epa.gov/sites/production/files/2013-</u>09/documents/cancer_guidelines_final_3-25-05.pdf
- U.S. EPA (U.S. Environmental Protection Agency). (2006). Provisional peer-reviewed toxicity values for 1,2-dibromo-3-chloropropane (CASRN 96-12-8) [EPA Report]. Cincinnati, OH. <u>http://hhpprtv.ornl.gov/issue_papers/Dibromo3Chloropropane12.pdf</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2009). Toxicological review of 1,2,3trichloropropane (CASRN 96-18-4): In support of summary information on the Integrated Risk Information System (IRIS) [EPA Report] (pp. 247). (EPA/635/R-08/010F). Washington, DC. https://iris.epa.gov/static/pdfs/0200tr.pdf
- U.S. EPA (U.S. Environmental Protection Agency). (2010). Provisional peer-reviewed toxicity values for 1,2-dichloroethane (CASRN 107-06-2) [EPA Report]. Cincinnati, OH. http://hhpprtv.ornl.gov/issue_papers/Dichloroethane12.pdf
- <u>U.S. EPA</u> (U.S. Environmental Protection Agency). (2011a). Chemical Assessment Clustering Engine (ChemACE) [Database]. Retrieved from <u>https://www.epa.gov/tsca-screening-tools/chemical-assessment-clustering-engine-chemace</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2011b). Health effects assessment summary tables (HEAST) for superfund. Available online at <u>https://epa-heast.ornl.gov/heast.php</u>

- U.S. EPA (U.S. Environmental Protection Agency). (2011c). Recommended use of body weight 3/4 as the default method in derivation of the oral reference dose. (EPA/100/R-11/0001). Washington, DC. <u>https://www.epa.gov/sites/production/files/2013-09/documents/recommended-use-of-bw34.pdf</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2012). Estimation Programs Interface Suite[™] for Microsoft® Windows, v 4.11: 1 Bromo-2-chloroethane (CASRN 107-04-0) [Fact Sheet]. Washington, DC. <u>https://www.epa.gov/tsca-screening-tools/epi-suitetm-estimation-program-interface</u>
- <u>U.S. EPA</u> (U.S. Environmental Protection Agency). (2015a). About the TSCA chemical substance inventory. Download the non-confidential TSCA inventory [Database]. Retrieved from <u>http://www2.epa.gov/tsca-inventory/how-access-tsca-inventory</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2015b). Bromochloroethane. ChemView [Computer Program]. Retrieved from <u>https://java.epa.gov/chemview#dashboard</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2015c). Significant New Use Rules (SNUR). Reviewing new chemicals under the Toxic Substances Control Act (TSCA). <u>http://www.epa.gov/reviewing-new-chemicals-under-toxic-substances-control-act-tsca/epa-actions-reduce-risk-new#SNURs</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2017). IRIS carcinogens: §69502.2(a)(1)(E). Chemicals that are identified as "carcinogenic to humans," "likely to be carcinogenic to humans," or Group A, B1, or B2 carcinogens in the United States Environmental Protection Agency's Integrated Risk Information System. Available online
- U.S. EPA (U.S. Environmental Protection Agency). (2018). Integrated Risk Information System. IRIS Assessments. Washington, DC. Retrieved from <u>http://www.epa.gov/iris/</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2020a). CompTox Chemicals Dashboard: 1,2-Dibromoethane [Database]. Washington, DC. Retrieved from <u>https://comptox.epa.gov/dashboard/dsstoxdb/results?search=DTXSID3020415#toxicity-values</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2020b). CompTox Chemicals Dashboard: 1,2-Dichloroethane [Database]. Retrieved from <u>https://comptox.epa.gov/dashboard/dsstoxdb/results?search=1,2-dichloroethane#toxicity-values</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2020c). Integrated risk information system. IRIS assessments [Database]. Washington, DC. Retrieved from <u>http://www.epa.gov/iris/</u>
- U.S. EPA (U.S. Environmental Protection Agency). (2020d). Provisional peer-reviewed toxicity values (PPRTVs) for superfund: Derivation support documents [Database]. Washington, DC. Retrieved from <u>https://www.epa.gov/pprtv</u>
- van Beerendonk, GJ; van Gog, FB; Vrieling, H; Pearson, PG; Nelson, SD; Meerman, JH. (1994). Blocking of in vitro DNA replication by deoxycytidine adducts of the mutagen and clastogen 2-bromoacrolein. Cancer Res 54: 679-684.
- van Bladeren, PJ; Breimer, DD; Rotteveel-Smijs, GM; de Knijff, P; Mohn, GR; van Meeteren-Wälchli, B; Buijs, W; van der Gen, A. (1981). The relation between the structure of vicinal dihalogen compounds and their mutagenic activation via conjugation to glutathione. Carcinogenesis 2: 499-505.
- Van Esch, G; Kroes, R; Van Logten, M; Den Tonkelaar, E. (1977). Ninety-day toxicity study with 1,2-dichloroethane (DCE) in rats. (Report 195/77 Alg.Tox). Bilthoven, the Netherlands: National Institute of Public Health and Environmental Protection.

Wang, NC; Zhao, QJ; Wesselkamper, SC; Lambert, JC; Petersen, D; Hess-Wilson, JK. (2012).

Application of computational toxicological approaches in human health risk assessment. I. A tiered surrogate approach. Regul Toxicol Pharmacol 63: 10-19. http://dx.doi.org/10.1016/j.yrtph.2012.02.006

<u>Wheeler, JB; Stourman, NV; Armstrong, RN; Guengerich, FP.</u> (2001). Conjugation of haloalkanes by bacterial and mammalian glutathione transferases: mono- and vicinal dihalothanes. Chem Res Toxicol 14: 1107-1117. <u>http://dx.doi.org/10.1021/tx0100183</u>